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3 **Title:** Multi-scale integration of satellite remote sensing improves characterization of dry-season
4 green-up in an Amazon tropical evergreen forest

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26

27 **Highlights**

- 28 • A robust method cross-calibrated PlanetScope using BRDF-adjusted MODIS.
- 29 • Calibrated data accurately assessed ecosystem-scale and crown-scale reflectance.
- 30 • A dry-season decrease in non-photosynthetic vegetation (NPV) fraction was detected.
- 31 • Large seasonal trend variability in crown-scale NPV fraction was quantified.

32 **Abstract**

33 In tropical forests, leaf phenology—particularly the pronounced dry-season green-up—
34 strongly regulates biogeochemical cycles of carbon and water fluxes. However, uncertainties
35 remain in the understanding of tropical forest leaf phenology at different spatial scales. Phenocams
36 accurately characterize leaf phenology at the crown and ecosystem scales but are limited to a few
37 sites and time spans of a few years. Time-series satellite observations might fill this gap, but the
38 commonly used satellites (e.g. MODIS, Landsat and Sentinel-2) have resolutions too coarse to
39 characterize single crowns. To resolve this observational challenge, we used the PlanetScope
40 constellation with a 3m resolution and near daily nadir-view coverage. We first developed a
41 rigorous method to cross-calibrate PlanetScope surface reflectance using daily BRDF-adjusted
42 MODIS as the reference. We then used linear spectral unmixing of calibrated PlanetScope to
43 obtain dry-season change in the fractional cover of green vegetation (GV) and non-photosynthetic
44 vegetation (NPV) at the PlanetScope pixel level. We used the Central Amazon Tapajos National
45 Forest k67 site, as all necessary data (from field to phenocam and satellite observations) was
46 available. For this proof of concept, we chose a set of 22 dates of PlanetScope measurements in
47 2018 and 16 in 2019, all from the six drier months of the year to provide the highest possible cloud-
48 free temporal resolution. Our results show that MODIS-calibrated dry-season PlanetScope data (1)
49 accurately assessed seasonal changes in ecosystem-scale and crown-scale spectral reflectance; (2)
50 detected an increase in ecosystem-scale GV fraction (and a decrease in NPV fraction) from June
51 to November of both years, consistent with local phenocam observations with R^2 around 0.8; and
52 (3) monitored large seasonal trend variability in crown-scale NPV fraction. Our results highlight
53 the potential of integrating multi-scale satellite observations to extend fine-scale leaf phenology
54 monitoring beyond the spatial limits of phenocams.

55

56

57 **Keywords:** Multi-scale satellite observations, PlanetScope, MODIS, BRDF correction,
58 reflectance cross-calibration, leaf phenology, non-photosynthetic vegetation, individual tree
59 crowns.

60

61 **1. Introduction**

62 Leaf phenology dominates seasonal and spatial variability in carbon and water fluxes (Jung
63 et al., 2019; Restrepo-Coupe et al., 2013), with important vegetation mediated feedbacks to
64 regional and global climates (Bonan, 2008; Wright et al., 2017). At the ecosystem scale, leaf
65 phenology emerges from all individuals and species living in a plant community, and the
66 phenology of these individuals shows large differential sensitivity response to climate change,
67 even within a temperate deciduous forest community (Richardson et al., 2018). Therefore, it is
68 increasingly important for the field to move towards the study of leaf phenology at the individual
69 tree-crown level.

70 Compared with the highly predictable phenological cycles in the temperate biomes, leaf
71 phenology in tropical evergreen forests is even more complex and less understood (Albert et al.,
72 2019; Reich, 1995). Much recent evidence from ground observations (Detto et al., 2018; Xu et al.,
73 2017) and phenocams (de Moura et al., 2017; Lopes et al., 2016) shows unusual leaf phenology
74 patterns in tropical “evergreen” forests. That is, the forest ecosystem appears evergreen all year
75 round, but strong seasonal leaf phenology dynamics occur at the tree-crown level with two typical
76 patterns. First, about 60-70% of all individuals rapidly exchange old leaves for new leaves during
77 the high-sunlight dry season (Gonçalves et al., 2020; Wu et al., 2016). Second, also in the dry
78 season, some upper canopy crowns drop part or all of their leaves and remain the leafless status
79 for few weeks prior to massive new leaf flush. These unique phenology patterns further cause
80 strong seasonal variation in ecosystem-scale leaf quality (i.e. photosynthetic capacity and optical
81 properties) as a function of leaf age mix, which helps explain the large dry season increase in
82 tropical forest photosynthesis (Albert et al., 2018; Wu et al., 2016) and satellite-detected greenness
83 (Wu et al., 2018). Despite the increasing importance of crown-scale phenology study in multiple
84 ecology-related fields, there is yet lacking high resolution monitoring that can help interpret fine-
85 scale phenological dynamics and explain large spatial heterogeneity across forest landscapes.
86 Therefore, accurate characterization and understanding of tropical leaf phenology (i.e. particularly
87 the pronounced dry-season phenological variations and green-up) at different spatial scales remain
88 an essential problem in tropical ecology studies.

89 However, several challenges remain. Phenocams may be the most accurate way to quantify
90 tropical leaf phenology from individual tree-crowns up to landscapes (Alberton et al., 2017; Lopes
91 et al., 2016; Moore et al., 2016), but are very limited in their footprints and time spans. For example,

92 a phenocam mounted on a 60m tower typically covers only dozens of upper canopy tree-crowns
93 within an area of several hectares (Wu et al., 2016). In addition, existing phenocams have been
94 deployed at only a few forest sites and span a few years (e.g. e-phenocam network in Brazil,
95 <http://www.recod.ic.unicamp.br/ephenology/client/index.html#/phenocamNetwork>). Satellite
96 remote sensing with large area coverage and frequent revisits can be a powerful alternative solution
97 (Huete et al., 2002, 2006). Unfortunately, as shown in a recent ground-based tree survey study in
98 an Amazon evergreen forest of French Guiana (Blanchard et al., 2016), the crown diameter for
99 tropical canopy crowns is normally small, ranging from a few meters to tens of meters at most. As
100 such, most commonly used satellite observations, such as Moderate Resolution Imaging
101 Spectroradiometer (MODIS) of 500m per pixel, Landsat of 30m, and Sentinel-2 of 10m, remain
102 too coarse to monitor leaf phenology dynamics at the individual tree-crown scale.

103 The increasing availability of high spatial and temporal resolution satellite data offers an
104 unprecedented opportunity to help resolve both the spatial coverage limitation of phenocams and
105 the lack of tree-crown scale observations from coarse-resolution satellite remote sensing.
106 Particularly, the PlanetScope constellation of more than 120 sensors (Planet Labs Inc., San
107 Francisco, CA) has several advantages, including daily-to-weekly global coverage at a 3m spatial
108 resolution and near-nadir view (Planet Team, 2018), but has not yet been fully explored. As with
109 other optical orbital sensors (e.g. Galvao et al., 2011; Samanta et al., 2010), the PlanetScope
110 reflectance products are also subject to cloud/aerosol contamination and the Bidirectional
111 Reflectance Distribution Function (BRDF) effect that is associated with image acquisition under
112 variable illumination and sensor viewing geometries. Individual PlanetScope sensors also have
113 inconsistencies in their DN scaling (Houborg and McCabe, 2018a, 2018b). However, a rigorous
114 method to utilize PlanetScope data to aid assessments of land surface reflectance seasonality is
115 neither developed yet nor rigorously evaluated. Additionally, multiple biophysical processes, such
116 as seasonal variations in canopy leaf area index (LAI), in leaf age mix and in canopy structure, can
117 affect canopy reflectance seasonality simultaneously (Wu et al., 2018), making it difficult to
118 directly connect observed canopy reflectance seasonality with leaf phenology (e.g. leafy versus
119 leafless phenostages) at the tree-crown scale.

120 Recent advances in satellite data fusion techniques and improved biophysical
121 understanding of satellite reflectance products appear to be promising for application of fine-scale
122 phenology monitoring in tropical evergreen forests. For example, by first calibrating Landsat using

123 BRDF-adjusted MODIS data and then explicitly accounting for both spatial (from 30m spatial
124 resolution Landsat) and temporal (from daily MODIS) variations, Luo et al. (2018) demonstrated
125 the feasibility to combine MODIS and Landsat satellites to enable land surface monitoring at daily,
126 30m resolution. This suggests the technical feasibility to integrate MODIS and PlanetScope data
127 to enable high-resolution phenology monitoring at a 3m resolution in tropical evergreen forests.
128 Additionally, several recent studies demonstrate the feasibility to differentiate leafless tree-crowns
129 from leafy tree-crowns using high-resolution satellite images, such as QuickBird of a 2.62m
130 resolution (Lopes et al., 2016) and WorldView-2 of a 1.84m resolution (Wu et al., 2018). The
131 underlying biophysical basis is that the reflectance spectra of leafless tree-crowns are significantly
132 different from those of leafy tree-crowns—a phenomenon that has been commonly observed
133 across multiple tropical forest sites over large tropical areas (Clark and Roberts, 2012; Lopes et
134 al., 2016; Viennois et al., 2013; Wu et al., 2018). As such, we believe high-resolution satellite data
135 like PlanetScope could offer a novel means to quantitatively differentiate the green vegetation (GV)
136 fraction from that of non-photosynthetic vegetation (NPV) in the upper canopy of tropical
137 evergreen forests.

138 The goal of this study is to investigate the technical feasibility and mechanistic soundness
139 of integrating MODIS and PlanetScope data for cross-scale (from fine-scale of 3m to landscapes
140 of a few kilometers) phenology monitoring, with a particular focus on the dry-season phenological
141 trend. Specifically, we first developed a method to rigorously cross-calibrate PlanetScope
142 reflectance data using BRDF-adjusted MODIS. We then evaluated the fine-scale robustness of the
143 developed method by assessing the seasonal reflectance pattern of permanent objects and assessed
144 the large-scale robustness by comparing ecosystem-scale seasonal reflectance pattern of the
145 calibrated data with the corresponding pattern from MODIS. Further, we estimated fractions of
146 GV and NPV at the pixel and ecosystem levels, using a linear spectral unmixing model. By this
147 means, we hope to use a metric of NPV fraction (or GV fraction) with clear biophysical meaning
148 for tropical phenology monitoring. For this proof-of-concept, we focus on the Central Amazon dry
149 season, when crown-scale and ecosystem-scale leaf phenology changes are more pronounced and
150 when more frequent cloud-free images are available.

151

152 **2. Study site and materials**

153 **2.1 Study site**

154 A Central Amazon tropical evergreen forest at the k67 eddy covariance tower site
155 (54°58'W, 2°51'S) was used in this study (Fig. 1a). It is the Tapajos National Forest, near Santarém,
156 Pará, Brazil. We selected this site for three reasons. First, there were rich related field observations
157 previously made at this forest site, including both field and tower-phenocam measurements of leaf
158 phenology (Brando et al., 2010; Wu et al., 2016). Second, the forest is on an extensive well-drained
159 clay-soil plateau (Rice et al., 2004), which minimizes the effects of topography on satellite-
160 detected canopy reflectance (Matsushita et al., 2007) and thus makes it easy to interpret satellite
161 data. Third, it is a typical tropical evergreen forest in the Central Amazon, and ground-observed
162 phenology pattern in this site is very comparable with that of other Central Amazon tropical forests
163 near Manaus, Brazil (Lopes et al., 2016; Wu et al 2016). Additionally, similar to other tropical
164 evergreen forests (Eamus, 1999), it has rich plant diversity and includes vast variability in crown-
165 level leaf phenology ranging from evergreen to semi deciduous and fully (but briefly) deciduous.
166 The forest has a mean annual air temperature of 26 °C (Hutyra et al., 2007), and a mean annual
167 precipitation of 2022 mm yr⁻¹ with a 5-month-long dry season from July to November (Wu et al.,
168 2016). For details about forest composition and structure of the k67 site see Rice et al. (2004).
169 About 37 km north of this forest site is the town of Alter do Chão, which is predominated by the
170 urban land cover type with some mixture of forests as well. The urban area, particularly the
171 building materials with constant reflectance spectra after BRDF correction, was included to
172 evaluate the robustness of our method (see Section 3.2.4).

173

174 **2.2 Materials**

175 Four different kinds of data were available at k67 to characterize leaf phenology patterns.
176 These include field measurements of LAI, tower-phenocam measurements of tree-crown
177 phenostages (i.e. leafy versus leafless), and two types of optical satellite remote sensing
178 (PlanetScope and MODIS). Since different datasets were sampled in different time periods with
179 various durations while optical satellites were subject to heavy cloud contamination in the wet
180 season, to minimize all these effects, we limited our study to trends in the long dry season.
181 Specifically, we aimed to use field- and tower-based phenology measurements to help evaluate
182 satellite-derived phenology metrics.

183 Field measurements of LAI (projected leaf area per unit ground area, m²m⁻²) were
184 previously made monthly from January 2000 to December 2005 at 100 grid points systematically

185 distributed in a 1-ha plot, ~5km from the k67 tower site, using two LiCor-2000 Plant Canopy
186 Analyzers (LiCOR Inc., Lincoln, NE). For details regarding the data and data collection procedure
187 see Brando et al. (2010). We here used the mean annual cycle of monthly field-observed LAI to
188 indicate the average ecosystem-scale phenology pattern of this forest site.

189 A 3-band (NIR, red, and green) Tetracam Agricultural Digital Camera (Tetracam Inc.,
190 Chatsworth, CA) was mounted on the k67 eddy covariance tower for leaf phenology monitoring
191 (Fig. 1b). It had a field of view of about $200 \times 300 \text{ m}^2$, covering around 65 upper canopy tree crowns.
192 The phenocam was programmed for automatic image acquisition at a 30-minute interval from
193 January 2010 to December 2011. The image acquisition stopped afterwards. For more details on
194 the phenocam data as well as the method used for phenology analysis, see Wu et al. (2016). Below
195 we briefly summarized the visual assessment approach (or ‘crown-based phenology inventory’
196 shown in Wu et al., 2016) for phenology analysis at k67. The approach includes the following five
197 steps: i) we manually selected a best quality image every 6 days (i.e. overcast, near local noon, and
198 free from shadows/rain/fog) throughout the entire image time series; ii) we divided the image of
199 forest landscape into discrete regions of interest (ROIs) that corresponded to each individual well-
200 illuminated tree-crown; iii) for each selected image we surveyed all the identified tree-crowns, and
201 visually assigned each crown to one of two phenostages: leafless (leaf shedding or bare branch
202 materials accounts for around or more than 50% of the entire tree crown area) or leafy (otherwise),
203 based on their colors, textures and temporal trends of leaf color within the adjacent two weeks; iv)
204 for each selected image, we calculated a metric called ‘leafless tree-crowns fraction’ (1- ‘leafy
205 tree-crowns fraction’) by dividing the number of leafless tree-crowns by all identified tree-crowns
206 ($n=65$), and v) a mean annual cycle of monthly ‘leafless tree-crowns fraction’ (or NPV fraction)
207 was derived to indicate ecosystem-scale average phenology at k67.

208 The four-band high-resolution PlanetScope data from Planet Labs Inc. were used (Fig. 1c
209 and Table 1). Planet Labs Inc. is an American private Earth imaging company, which offers daily
210 images of global coverage, including PlanetScope (3m resolution, daily revisit cycle with near
211 nadir view) and RapidEye (5m resolution, 5.5 days revisit cycle with near nadir view) satellite
212 imagery. Researchers can access PlanetScope data through a research and education license. We
213 here used the PlanetScope data, as it has finer spatial and temporal resolutions compared with
214 RapidEye. At k67, we surveyed all available PlanetScope data for 2018 and 2019, and found no
215 good wet season images (January to May, plus December) due to heavy cloud contamination.

216 Therefore, only the six months (June to November) PlanetScope data with low cloud cover (<40%;
217 predetermined by cloud filter provided by Planet Lab Inc.) were downloaded and used, including
218 22 dates in 2018 and 16 dates in 2019. For clarity purposes, we focus on the 22 dates of
219 measurements from 2018 in the main text. Results for 2019, which are very similar as the results
220 for 2018, are in supplementary materials. At the Alter do Chão site, a total of 30 dates of
221 PlanetScope measurements in 2018 from June to December were downloaded and used. For all
222 PlanetScope data used in this study, the sensor viewing angle was less than 1.2° off nadir.

223 The coarse-resolution MODIS data was also used (Fig. 1d). Since MODIS covers the four
224 bands of PlanetScope (Table 1), it makes the integration of these two satellites possible. Here we
225 used the MODIS BRDF/Albedo model parameter product, MCD43A1 (Schaaf et al., 2002), for
226 three reasons. First, increasing evidence suggests that the BRDF effect associated with sun-sensor
227 geometry is an important confounding factor affecting satellite-detected phenology in tropical
228 forests (Galvao et al., 2011; Morton et al., 2014; Saleska et al., 2016). Second, BRDF-adjusted
229 MODIS products detect tropical forest phenology in good agreement with ground measurements
230 (Lopes et al., 2016; Wagner et al., 2016). Third, the BRDF-adjusted MODIS products from
231 MCD43A1 have been rigorously validated previously (Maeda et al., 2016; Wu et al., 2018). The
232 daily MCD43A1 data of 500m from February of 2000 to December of 2019 were downloaded.
233 We then computed the BRDF-adjusted reflectance at a simulated nadir view, 0° relative azimuth
234 angle, and 45° solar zenith angle, using the BRDF model parameters in the MCD43A1 data as
235 inputs to the semi-empirical RossThick-LiSparse reciprocal model (Wanner et al., 1995).

236 For the satellite data used in this study, there are four spatial extents involved and each has
237 a distinct purpose. First, we used a 10km×10km area of plateau forest (Fig. 1c, d) representing the
238 entire k67 site for cross-calibration by histogram matching between PlanetScope and MODIS (see
239 Section 3.2.3). This is because a significant amount of pixels after quality control are needed for
240 proper histogram matching, and the 10km×10km area provides sufficient valid MODIS pixels after
241 quality control. One PlanetScope scene often may not cover the whole area, and thus in our study,
242 multiple PlanetScope scenes of the same day were mosaicked and cropped for a full coverage of
243 the 10km×10km area. Second, we used an 8km×8km area for the Alter do Chão site (Fig. 1e, f),
244 and PlanetScope data of this site was firstly used to identify the permanent objects (i.e. buildings)
245 that are spectrally stable and then used to evaluate the robustness of our method (see Section 3.2.4).
246 Third, we used a moving window of a 5km×5km area surrounding a target pixel to firstly generate

247 the quality assurance (QA) flag (i.e. QA30; more details shown in Section 3.1) for that pixel and
248 then assessed the gap-filling procedure based on its QA time series (more details shown in Section
249 3.2.2). Fourth, we used a 3km×3km area centered on the k67 tower to calculate the seasonal trends
250 of ecosystem-scale phenology derived from PlanetScope and MODIS, and then compared them
251 with field and phenocam observations of leaf phenology.

252

253 **3. Methods**

254 In order to develop a rigorous method that integrates high-resolution PlanetScope with
255 coarse-resolution MODIS for cross-scale phenology monitoring in tropical forests, we divided the
256 work into the following four tasks: 1) acquiring and processing PlanetScope and MODIS data; 2)
257 cross-calibrating PlanetScope data using BRDF-adjusted MODIS as the reference; 3) extracting
258 reflectance spectra of the three key endmembers comprising tropical forest canopies (NPV, GV
259 and shade; also see Fig. S1 for these example endmembers shown in a WorldView-2 image at the
260 k67 site), and estimating endmember fractions for each calibrated PlanetScope image using a linear
261 spectral unmixing model; and 4) evaluating the accuracy of derived seasonal trends in NPV and
262 GV fractions from calibrated PlanetScope by comparing with ground and phenocam observations
263 of leaf phenology. The first two tasks aim to improve the data quality of PlanetScope by developing
264 consistent data processing tied to high quality data. The last two tasks aim to improve biophysical
265 interpretation of PlanetScope data, by transforming surface reflectances to fractional covers of the
266 real-world constituents of tropical forest canopy. A flow chart that summarizes the method and the
267 four key tasks is shown in Fig. 2.

268

269 **3.1 (Task 1): Acquiring and processing PlanetScope and MODIS data**

270 (1) PlanetScope. The orthorectified, near nadir view, level 3B surface reflectance product
271 was accessed at <https://www.planet.com/>, including the quality control layer. We first generated a
272 data quality mask based on the Unusable Data Mask (UDM) layer, following the instructions in
273 Planet (2019). This default cloud masking, however, did not fully detect all cloud contamination.
274 We therefore implemented a customized cloud/cloud-shadow removal algorithm that operated for
275 each PlanetScope image, following Fraser et al. (2009) and Hillger and Clark (2002). This
276 algorithm uses principal component analysis (PCA) (Chavez and Kwarteng, 1989) together with

277 Otsu thresholding (Otsu, 1979). After applying this additional clouds/cloud shadows quality
278 control, we carefully checked and manually masked any remaining clouds and cloud shadows.
279 This last step was labor intensive but developing a better automatic cloud/cloud-shadow removal
280 algorithm is beyond the scope of this paper.

281 (2) MODIS. The MCD43A1 product was accessed at <https://search.earthdata.nasa.gov/>.
282 This product provides the model parameters for removing the BRDF effect. We first adjusted the
283 MCD43A1 reflectance to a nadir view, 0° relative azimuth angle and 45° solar zenith angle,
284 following the technical guide at
285 (https://www.umb.edu/spectralmass/terra_aqua_modis/v006/introduction). We then generated
286 and applied the band-specific pixel QA layer that indicates high-quality band-specific BRDF
287 inversion results (i.e. only using quality bit index = 0/1 for full/magnitude BRDF inversions) to
288 retain as many good pixels (Schaaf et al., 2002, 2011). These procedures were applied to each
289 pixel of daily MODIS data. Following Wu et al. (2018), we also applied the QA30 flag to further
290 minimize cloud/aerosol impacts. This flag assumes that a pixel is most likely free from
291 cloud/aerosol contamination when at least 30% of the pixels within a 5km×5km area centered on
292 this pixel are also valid (i.e. passing through pixel QA control). Collectively, we applied QA30
293 when calculating the mean MODIS seasonality using all 20-year (2000-2019) data as well as
294 calculating the seasonal trends for the MODIS data in 2018 and 2019 (to match with PlanetScope
295 data for cross-calibration).

296

297 **3.2 (Task 2):** Cross-calibrating PlanetScope data using BRDF-adjusted MODIS

298 We divided this task into four sub-tasks described in sub-sections 3.2.1 through 3.2.4.

299

300 3.2.1 Retaining best quality MODIS data in 2018 and 2019

301 For this subtask, first, the band-specific quality control (including pixel QA and associated
302 QA30 flag) was generated and applied to daily MODIS data in 2018 and 2019 to filter bad pixels
303 with cloud/aerosol/cloud-shadow contaminations. Second, for each pixel of our study sites, we
304 used all the MODIS data (2000-2019) to derive an annual cycle of band-specific means and their
305 95% confidence intervals at the daily timescale. This data had minimum cloud/aerosol
306 contamination as it was sourced from a large selection of candidate pixels over time, and was then

307 used as an additional quality control to help filter any remaining bad pixels not previously removed
308 in the MODIS 2018 and 2019 data.

309

310 3.2.2 Gap-filling MODIS time series in 2018 and 2019

311 High-quality paired satellite data of the same day were needed for cross-calibration but
312 were not always available. To meet this need, we used the available best quality MODIS data for
313 2018 and 2019 (as described in Section 3.2.1 above), and developed a gap-filling method to gap
314 fill those missing values in 2018 and 2019. MODIS gap-filling for these two years was performed
315 on a daily, per band and per pixel level, following the three scenarios as follows:

316 i) if there were more than 200 valid daily measurements among a total of 365 measurements
317 for a target MODIS pixel in 2018/2019, we then fitted a cubic spline curve to the data and used
318 the daily fitted values for all days with missing values;

319 ii) if there were more than 50 but less than 200 valid daily measurements for a target
320 MODIS pixel in 2018/2019, we turned to a reference (i.e. mean valid daily measurements from a
321 neighboring 5km×5km area centered on the target pixel) to assist gap-filling, with two situations.
322 If the current-year MODIS data (2018/2019) of the neighboring 5km×5km area had sufficient
323 (≥ 200) mean valid daily measurements, the current-year data only was used to calculate the
324 reference. Otherwise, all 20 years of data (2000-2009) were used to derive the reference. With the
325 reference, we fitted its seasonal trend with a cubic spline curve to cover the full annual cycle. In
326 order to preserve any real divergence between the target pixel and the average from its neighboring
327 5km×5km area, we then shifted the shape of the fitted reference curve up and down until it best
328 matched with all the valid daily measurements of the target pixel in 2018/2019. We last
329 interpolated those missing values of the target pixel based on the shifted curve.

330 iii) if there were less than 50 valid daily measurements for a target MODIS pixel in
331 2018/2019, we assumed each day's reflectance of the target pixel was the same as from 20-year
332 daily mean of its neighboring 5km×5km area. We first fitted a cubic spline curve to the 20-year
333 daily mean, and then used the daily fitted values to fill all missing values of the target pixel in
334 2018/2019.

335 Our Fig. S2 summarized the relative abundance of these three gap-filling scenarios. The
336 first two gap-filling scenarios dominated (>99%) in our study.

337 Since a long-term (2000-2019) mean annual cycle was involved in our gap-filling
338 procedures (i.e. scenarios ii and iii, above), in order to test whether it is fine to use this long-term
339 mean annual cycle, we respectively examined the ecosystem-scale MODIS daily mean for 2018,
340 2019 and 2000-2019. As shown in Figs. 3 and S3, the curve of 2018/2019 and the curve of long-
341 term mean are very similar, suggesting that the use of long-term mean annual cycle to aid our gap-
342 filling processes is reasonable. Also in Figs. 3 and S3, the seasonal trend of gap-filled MODIS in
343 2018/2019 very well tracks that of original, non-gap-filled MODIS in 2018/2019, providing
344 additional confidence on the reliability of our gap-filling method. Admittedly, there was an
345 empirical choice of the thresholds (at 200 and 50 valid daily values per year) in determine gap-
346 filling procedures. A further sensitivity test (Fig. S4), however, suggests that our gap-filling results
347 are valid. For demonstration purpose, the original MODIS data with gaps and the corresponding
348 gap-filled MODIS data are shown in the top and bottom panels of Fig. S5, respectively.

349

350 3.2.3 Cross-calibrating PlanetScope using band-specific histogram matching

351 PlanetScope was cross-calibrated to the (gap-filled) BRDF-adjusted MODIS of the same
352 day using histogram matching—a commonly-used method for spectral cross-calibration of data
353 acquired from different sensors (Chavez and Mackinnon, 1994; Yang and Lo, 2000). The cross-
354 calibration was conducted for each band of each PlanetScope image. We first upscaled
355 PlanetScope to the MODIS spatial resolution, and then calculated the pair of band-specific
356 histograms respectively for MODIS and the upscaled PlanetScope. A Gaussian distribution was
357 then applied to fit the histograms (Fig. 4). For each reflectance band, we adjusted the upscaled
358 PlanetScope data to give the same mean and standard deviation as the corresponding band from
359 BRDF-adjusted MODIS, using a linear transformation. We recorded the band-specific adjustment
360 coefficients. We then applied these derived coefficients to the PlanetScope images at their original
361 3m resolution. Fig. S6 shows band-specific comparisons between PlanetScope data of pre and post
362 cross-calibration. The calibrated PlanetScope data show less variability over the season compared
363 to the original uncalibrated data.

364

365 3.2.4 Evaluating the robustness of the cross-calibration results

366 The robustness of our cross-calibration was evaluated by assessing the spectral consistency
367 of permanent objects, i.e. the extracted building pixels at the Alter do Chão site that should have

368 stable reflectance spectra over a season after BRDF correction. Specifically, we manually
369 identified about 180 building pixels in the PlanetScope images, and assessed the seasonal
370 variability of their reflectance spectra prior and post cross-calibration.

371

372 **3.3 (Task 3):** Estimating NPV and GV fractions using linear spectral unmixing of calibrated 373 PlanetScope

374 We hypothesized that there are three key elements within each forest canopy pixel (Fig.
375 S1): NPV (bare illuminated branches), GV (illuminated green leaves) and shade (shadow caused
376 by tall crowns and by deep narrow gaps). Our objective was to transform canopy reflectance of
377 the four PlanetScope bands into variables with clear biophysical meaning. The variables of greatest
378 interest were the NPV fraction and GV fraction. Details on the extraction of endmember-specific
379 reflectance spectra and the linear spectral unmixing model follow.

380 (1) Extracting endmember-specific reflectance spectra. We followed an existing approach
381 (Roberts et al., 1992) to extract endmember-specific reflectance spectra. It includes the following
382 four steps. First, we applied a single principle component (PC) transformation to all calibrated
383 PlanetScope data to summarize their 4-D band space in a 2-D scatter plot of PC1 and PC2, which
384 resulted in a triangular point cloud (Fig. S7). Second, we performed careful visual assessments to
385 manually identify pure pixels of each of the three endmembers, collecting at least 80 pixels per
386 endmember from the calibrated PlanetScope images covering the full dry season. Third, we plotted
387 and overlaid these manually identified pure endmembers on the scatter plot as in Fig. S7, to see if
388 they coincided with the point cloud vertices, as required for unmixing. Also in Fig. S7, we then
389 delimited the rectangle for each endmember based on the mean and two standard deviations of
390 each PCA axis derived from those manually identified. Last, we calculated the average reflectance
391 per band for all image pixels found within the three rectangles to derive reflectance spectra for the
392 three endmembers. We compared these endmember-specific reflectance spectra to spectra derived
393 from the smaller number of visually sampled pixels and the results were very comparable (Fig.
394 S8).

395 (2) Estimating pixel-level NPV and GV fractions using a linear spectral unmixing model.
396 With the derived endmember-specific reflectance spectra, we then applied a linear spectral
397 unmixing model (Keshava and Mustard, 2002) to estimate the fractional cover of each endmember
398 on a pixel-by-pixel basis for each PlanetScope image, assuming that the three endmembers

399 contribute to surface reflectance in a weight that is linearly proportional to their fractional cover
 400 within a pixel. The linear spectral unmixing model can be written as below:

$$401 \quad x_i = \sum_{k=1}^M p_{ik} e_k + \epsilon_i, i = 1, \dots, N \quad (1)$$

402 where p_{ik} is the fraction of endmember k in pixel i , e_k is the reflectance spectra of the k th
 403 endmember, M is the number of endmembers ($M=3$), ϵ_i is an error term, x_i is the reflectance spectra
 404 of pixel i , and N is the total number of pixels of a given PlanetScope image. The fractional values
 405 of this model satisfy the constraints

$$406 \quad p_{ik} \geq 0 \forall k = 1, \dots, M; \sum_{k=1}^M p_{ik} = 1. \quad (2)$$

407 After estimating endmember-specific fractions, including the fractional covers for $k=1, 2,$
 408 3 , respectively representing NPV, GV, and shade, we then re-assigned the shade fraction of each
 409 pixel to its NPV and GV components (see Eqns. 3 and 4), according to their initial, estimated
 410 fraction values in Eqn. 1. This is based on an assumption that shade effect happens equally to the
 411 NPV and GV elements.

$$412 \quad p_{i1}' = p_{i1} + \frac{p_{i1}}{p_{i1}+p_{i2}} \times p_{i3} \quad (3)$$

$$413 \quad p_{i2}' = p_{i2} + \frac{p_{i2}}{p_{i1}+p_{i2}} \times p_{i3} \quad (4)$$

414 where p_{i1}, p_{i2} and p_{i1}', p_{i2}' are the fractions of pure endmembers of NPV and GV in pixel i before
 415 and after reassignment of the shade fraction (p_{i3}). After this attribution of shade, we were left with
 416 two endmembers, NPV and GV, whose fractional contribution to each PlanetScope pixel sums to
 417 1.0. We also applied the same linear spectral unmixing method, including endmember-specific
 418 reflectance spectra derived from calibrated PlanetScope, to the BRDF-adjusted MODIS data.

419
 420 **3.4 (Task 4):** Evaluating the accuracy of PlanetScope-derived seasonal trends in NPV and GV
 421 fractions by comparing with field and phenocam measurements of leaf phenology

422 We evaluated linear regressions between ecosystem-scale PlanetScope-derived NPV and
 423 GV fractions over the six months (June to November) in 2018/2019 and three local phenology
 424 measurements, including: (i) field measurements of LAI, (ii) phenocam-based leafless tree-crown
 425 fraction, and (iii) phenocam-based leafy tree-crown fraction (=1-leafless tree crown fraction).

426
 427 **3.5 (Sensitivity analysis):** Assessing the effects of non-matching time span or spatial coverage on
 428 the derived ecosystem-scale leaf phenology patterns

429 There are multiple datasets involved in this study. These vary in the year of measurement,
430 the multi-year duration of measurement and the spatial extent. Therefore, it is important to assess
431 whether such variations would impact the derived ecosystem-scale leaf phenology patterns. For
432 this purpose, we assessed the effects of temporal duration/offset and of spatial extent, respectively,
433 summarized as below:

434 (1) Evaluating the effect of temporal duration and temporal offset on the derived
435 ecosystem-scale phenology. In this study, four datasets were used, including phenocam
436 observations (daily in years 2010-2011), field LAI measurements (monthly in years 2000-2005),
437 MODIS data (daily in years 2000-2019), and PlanetScope data (22 dates of measurements in 2018
438 and 16 dates in 2019). Based on these datasets, we first tested whether the measurements from
439 different years and multi-year durations would affect ecosystem-scale phenology using daily raw,
440 non-gap-filled MODIS. The sensitivity analysis shows that the (average) ecosystem-scale
441 phenology trends in 2000-2005, 2010-2011, 2018 and 2019 are very similar, and are all close to
442 the 20-year mean seasonality at our study site (Fig. S9a, b). The analysis thus suggests that our
443 local observations, though collected in different time periods, are still very useful to help evaluate
444 ecosystem-scale phenology in 2018 and 2019.

445 (2) Evaluating the effect of spatial extent on the derived ecosystem-scale phenology.
446 Spatial extent also differs in our datasets, including phenocam (in an about 200m×300m area),
447 field LAI measurements (in a 100m×100m area), MODIS (in a 10km×10km area) and PlanetScope
448 (in a 10km×10km area). To test the effect of spatial extent, we used the calibrated PlanetScope
449 data and respectively derived the ecosystem-scale average seasonal trend for windows of different
450 sizes centered on the k67 eddy covariance tower: 100m×100m, 200m×300m, 500m×500m,
451 1km×1km, 3km×3km, 5km×5km, and 10km×10km. The sensitivity analysis shows that regardless
452 of the spatial extent, the (average) ecosystem-scale seasonal trends are quite similar and
453 comparable (Fig. S9c, d). This also suggests that local observations from phenocam and LAI have
454 sufficient spatial coverage to represent ecosystem-scale phenology at k67.

455

456 **4. Results**

457 **4.1 Cross-calibrating PlanetScope: robustness of the method and the seasonal trend**

458 The robustness of our cross-calibration results was evaluated through the following three
459 types of assessments: 1) at the forest ecosystem scale, the average seasonal variability of all bands

460 and of two vegetation indices (VIs), the normalized difference vegetation index (NDVI) and the
461 enhanced vegetation index (EVI) (Figs. 5 and S10); 2) at the fine spatial scale, the seasonal trend
462 in reflectance spectra of permanent objects (i.e. buildings) (Fig. 6); and 3) also at the fine scale,
463 the spectral reflectance of three pure endmembers (NPV, GV, and shade) (Fig. 7).

464 In the first assessment, our results show that calibrated PlanetScope captures the same
465 ecosystem-scale seasonal trends from June to November as BRDF-adjusted MODIS, a consistent
466 pattern throughout all four spectral bands and both VIs (blue, green, red, NIR, NDVI, and EVI;
467 Figs. 5 and S10). Uncalibrated PlanetScope also shows similar dry season increasing trends across
468 four reflectance bands and two VIs, but is not in perfect agreement with that of MODIS, possibly
469 due to a BRDF effect related to dry-season variation in solar elevation as well as the inconsistency
470 in the DN scaling among different PlanetScope sensors. As a result, the uncalibrated PlanetScope
471 reflectance values and their seasonal ranges are both higher than those of BRDF-adjusted MODIS
472 (Figs. 5 and S10). For example, the dry season range of blue reflectance in BRDF-adjusted MODIS
473 is only 0.003-0.035, but it is 0.014-0.098 for PlanetScope. Similarly, ranges of 0.017-0.06, 0.007-
474 0.045, and 0.20-0.40 are found in BRDF-adjusted MODIS green, red and NIR reflectances
475 respectively, while 0.020-0.110, 0.020-0.093, and 0.23-0.42 are found in the corresponding
476 uncalibrated PlanetScope. The cross-calibration also reduces the PlanetScope fluctuation that
477 departs from the overall uncalibrated seasonal trend. Such cross comparisons thus suggest that the
478 PlanetScope surface reflectance data indeed requires cross-calibration, and the proposed method
479 effectively cross-calibrates the PlanetScope data, resulting in the same seasonal trend as MODIS
480 at the ecosystem scale.

481 In addition to the ecosystem-scale consistency, our two additional assessments also
482 demonstrate that calibrated PlanetScope rigorously captures the seasonal trend in surface
483 reflectance at the fine scale. The assessment from the permanent objects demonstrates that the
484 cross-calibration stabilizes the reflectance spectra variability of the buildings at the fine scale. As
485 shown in Fig. 6, after the cross-calibration, the buildings have nearly constant reflectance spectra
486 across the full dry season from June to November. In contrast, prior to calibration, their reflectance
487 spectra show large inter-month variability. Another assessment from endmember-specific
488 reflectance spectra also suggests that the post-calibration reflectance spectra for each pure
489 endmember (of 3m spatial scale of canopy surface) show small variation within each endmember,

490 while displaying large inter-endmember differences for all four reflectance bands (Fig. 7),
491 providing further confidence on the robustness of our calibrated results at the fine scale.

492 Despite MODIS and calibrated PlanetScope having the same seasonal trend in ecosystem-
493 scale reflectance as above, our results suggest that calibrated PlanetScope also provides rich and
494 detailed phenological variations at the pixel level of 3m (Fig. 8). Particularly, PlanetScope captures
495 large variability in reflectance dynamics at the fine scale across both space and time. Such observed
496 high seasonal variability at the fine-scale might result from the fact that the tropical forest at k67
497 harbors high plant diversity, and tree individuals of different species or the same species but with
498 different growth environments vary in their phenological events including both timing and
499 magnitude.

500

501 **4.2 Evaluating the seasonal trends in NPV and GV fractions derived from the calibrated** 502 **PlanetScope data**

503 We compared the PlanetScope-derived NPV fraction (and its complement, GV fraction)
504 with phenocam and field LAI observations. Our results demonstrate that at the ecosystem-scale,
505 the estimated seasonal trends in NPV fraction agree well with phenocam observations of leafless
506 tree-crown fraction ($R^2=0.82$, $p=0.014$ for PlanetScope 2018 in Fig. 9a; $R^2=0.73$, $p=0.030$ for
507 PlanetScope 2019 in Fig. S11a), all of which show a decreasing trend in NPV fraction throughout
508 the full dry season. The absolute dry-season change in NPV fraction is also similar across all
509 indicators from both phenocam and PlanetScope: about 10% decrease from June to November.
510 Meanwhile, we also observed a modest absolute value difference between the two approaches,
511 with the NPV fraction derived from PlanetScope consistently having ~5-10% higher fraction
512 values than the phenocam observations. The alternative metric, GV fraction from calibrated
513 PlanetScope, also shows the expected complimentary dry-season increasing trends with phenocam
514 observations ($R^2=0.82$, $p=0.014$ for PlanetScope 2018 in Fig. 9b; $R^2=0.73$, $p=0.030$ for
515 PlanetScope 2019 in Fig. S11b) and field LAI measurements ($R^2=0.81$, $p=0.015$ for PlanetScope
516 2018 in Fig. 9b; $R^2=0.72$, $p=0.034$ for PlanetScope 2019 in Fig. S11b). Additionally, the NPV
517 fraction extracted from BRDF-adjusted MODIS using the same linear spectral unmixing model
518 also shows the similar pattern as that from calibrated PlanetScope in 2018 ($R^2=0.78$, $p=0.019$; Fig.
519 S12) but having closer relationships with phenocam observations of the same year in 2010-2011
520 ($R^2=0.96$, $p=0.001$; Fig. S12).

521 In addition, we assessed pixel-level (i.e. 3m resolution) seasonal variability in NPV
522 fraction extracted from calibrated PlanetScope. Our results in Fig. 10 show that there are large
523 seasonal variations in NPV fraction at the pixel level, with some pixels exhibiting a similar dry-
524 season decreasing trend compared with the ecosystem-scale average pattern as shown in Fig. 9a
525 but with large differences in change magnitude across pixels, while other pixels exhibit no trend
526 or an increasing seasonal trend. Meanwhile, Fig. 11 provides the dry-season change rate in NPV
527 fraction at the pixel level with a much greater decreasing trend than increasing trend in NPV
528 fraction (84.6% vs. 15.4%) in the same area. Further, by assessing the seasonal changing trend
529 across all the PlanetScope pixels at the k67 site, our results in Fig. S13 suggest that there are 71.2%
530 (and 74.4%) of all pixels showing a dry-season decreasing trend in NPV fraction (i.e. a green-up)
531 while 28.8% (and 25.6%) of all pixels showing a dry-season increasing trend (i.e. a brown-down)
532 in 2018 (and 2019).

533

534 **5. Discussion**

535 Understanding patterns of plant phenology from individual tree-crowns up to ecosystems
536 remains a critical challenge in plant ecology in general (Berra et al., 2019; Hufkens et al., 2012)
537 and ecology of tropical evergreen forests in particular (Albert et al., 2018; Lopes et al., 2016; Park
538 et al., 2019). In this study, we demonstrated that an integration of high-resolution PlanetScope
539 with coarse-resolution MODIS improves characterization of dry-season phenostages and green-up
540 of tropical evergreen forests across a wide range of spatial scales from a pixel of 3m (i.e. the scale
541 of an individual tree-crown or below) up to ecosystems. Combined with a linear spectral unmixing
542 model, such cross-satellite integration quantitatively differentiates GV from NPV, which is
543 superior to conventional phenology monitoring using reflectance or a vegetation index because it
544 has improved biophysical meaning. Our work thus represents a significant step forward in our
545 ability to improve characterization of dry-season leaf phenology pattern in tropical evergreen
546 forests, ranging from tree-crown scales to ecosystems and from conventional metrics of reflectance
547 or vegetation index to GV and NPV fractions.

548 Our proposed PlanetScope-MODIS integration is similar in concept to previous cross-
549 sensor fusion/calibration work, but with advances. As in prior fusion/calibration work (Gao et al.,
550 2006; Houborg & McCabe, 2018a, 2018b), we used an orbital sensor of coarse spatial resolution
551 that is frequent, accurate and corrected for BRDF effects as the benchmark to cross-calibrate a

552 high spatial resolution sensor with lower accuracy and uncorrected BRDF. We followed the
553 approach of Luo et al. (2018) to gap fill missing days for each pixel in the MODIS timeline
554 according to its own seasonal trend or the seasonal trends from adjacent pixels. Different from
555 previous fusion/calibration work, we did not gap fill the high spatial resolution sensor timeline,
556 because PlanetScope provides high frequency nadir-view coverage across the full dry season for
557 our site. Beyond these similarities and differences, our approach includes three major
558 advancements.

559 First, to our knowledge, it is the first study to integrate multiple sensors and orbital
560 platforms to improve fine-scale leaf phenology studies of tropical evergreen forest ecosystems.
561 Cross-scale multi-satellite integration has been challenging in tropical evergreen forests, due to the
562 frequent cloud cover over the annual cycle. Consequently, most satellite fusion/calibration
563 techniques have been developed and applied in other biomes (Liao et al., 2019; Semmens et al.,
564 2016; Walker et al., 2012; Yang et al., 2017) but has been less used in tropical biomes both in
565 techniques and mechanism for phenology monitoring (Viennois et al., 2013; Zeng et al., 2018).
566 Here we demonstrated the feasibility of integrating PlanetScope with MODIS for cross-scale
567 detection of tropical forest leaf phenology (Figs. 5 and 8). In contrast, the coarse-resolution
568 MODIS sensors alone can detect ecosystem-scale but not fine-scale leaf phenology dynamics.
569 Similarly, the use of the PlanetScope constellation alone is unsuccessful due to poor calibration
570 (Houborg & McCabe, 2018a, 2018b) and seasonally varying solar elevation, leading to noisy and
571 biased reflectance values over the season, impeding leaf phenology monitoring at both ecosystem
572 and tree-crown scales (Figs. 5 and 6). Only with the PlanetScope-MODIS integration, we detected
573 the dry season leaf phenology dynamics at both ecosystem (i.e. an overall dry-season green-up
574 pattern) and individual tree-crown (i.e. pronounced phenological diversity among individuals)
575 scales. These remotely detected phenology patterns agree with many previous findings from field
576 (e.g. Brando et al., 2010), tower-phenocam (e.g. Wu et al., 2016), and satellite (e.g. Huete et al.,
577 2006; Saleska et al., 2016) observations. For example, we confirm that Central Amazon evergreen
578 forests undergo leaf turnover (as indicated by many pre-flush leafless phenostage crowns)
579 followed by ecosystem-scale green-up (due to post-flush leaf maturation) in the dry season period
580 of high sunlight and reduced rainfall (Wu et al., 2018). This suggests that these forests are not
581 water limited and are more likely light limited (Guan et al., 2015; Huete et al., 2006). However, it
582 remains mechanistically unclear and awaits more in-depth future exploration regarding why there

583 is such high inter-crown phenological diversity during the long dry-season (Figs. 10-11), despite
584 an overall ecosystem-scale green-up pattern (Fig. 9).

585 The success of our multi-sensor integration relies, first of all, on several conditions: i)
586 MODIS has long-term frequent measurements, which provides sufficient observations to obtain
587 cloud-free samples over annual cycles. Thus it is feasible to use the mean seasonal trend to help
588 interpolate missing daily values due to cloud/aerosol/cloud shading contamination; ii) MODIS
589 BRDF-adjusted products have been rigorously validated previously (Maeda et al., 2016; Wagner
590 et al., 2016; Wu et al., 2018), and thus can serve as a good reference for benchmarking other
591 satellites, such as PlanetScope shown in this study (Fig. 6); and iii) PlanetScope has frequent
592 measurements over the annual cycles, e.g. nearly daily revisit cycle, which also makes it feasible
593 to obtain frequent cloud-free data, especially during the less cloudy dry season (Fig. 5). The
594 success of our integration also suggests that the same approach might be extendable to Sentinel-2
595 (with 5day interval, 10m resolution) (Drusch et al., 2012) and other satellites with both frequent
596 revisit and high spatial resolution (e.g. GeoEye-1, GaoFen-2, VENUS and Pleiades) (Dedieu et al,
597 2006; Dribault et al., 2012; Gu and Tong, 2015; Pu et al., 2018). We recommend BRDF-adjusted
598 MODIS be used as a calibration reference for such multi-sensor integration.

599 Second, we applied rigorous assessments to ensure such PlanetScope-MODIS integration
600 worked consistently well across all scales. The calibrated PlanetScope data exhibited the same
601 seasonal pattern as MODIS at the ecosystem scale (e.g. green lines in Fig. 5). However, this alone
602 does not prove the cross-calibration also works at the fine spatial scale. For validation of fine-scale
603 we performed one additional assessment which is the post-calibration spectral stability over the
604 entire dry season for permanent objects (i.e. buildings) (Fig. 6b). The other is the assessment of
605 endmember-specific reflectance spectra (NPV, GV, and shade) extracted from the calibrated
606 PlanetScope data (Fig. 7). These endmember-specific reflectance spectra agree well with previous
607 findings based on field measurements of reflectance spectra of the three canopy materials (Asner,
608 1998; Clark and Roberts, 2012), with other high-resolution satellite data (Feret et al., 2015), and
609 with process-based model simulations (Wu et al., 2018). In summary, a multitude of validation
610 assessments suggest the proposed PlanetScope-MODIS integration works consistently well at both
611 fine and ecosystem scales. Additionally, because stable building reflectance spectra and NPV and
612 GV fractions were extracted on the PlanetScope pixel level of 3m, we conclude the cross-
613 calibration allows detection of tree-crown scale phenostages.

614 Third, our approach provides a metric with clear biophysical meaning, the NPV fraction or
615 its complement, the GV fraction, to aid quantitative measurements of tropical leaf phenology.
616 Satellite remote sensing has been powerful to monitor land surface phenology over large areas
617 (Moulin et al., 1997; White et al., 2009), but lacks clear biophysical meaning if expressed as
618 canopy reflectance, or even as a vegetation index (Samanta et al., 2012; Wu et al., 2018). The
619 timing of massive leaf flush and of complete or partial loss are important phenostages at the tree-
620 crown scale and are detectable using canopy leaf fractional cover (Lopes et al., 2016; Richardson
621 et al., 2018). Based on this idea, we derived the NPV fraction to represent fractional cover of non-
622 photosynthetic vegetation within a PlanetScope pixel using linear spectral unmixing. Our derived
623 dry-season NPV trends demonstrate strong ecosystem-scale agreement with phenocam
624 observations (Fig. 9) while also characterize large inter-crown variance at the fine-scale (Figs. 10-
625 11), highlighting the effectiveness of our approach.

626 It is also worthy to note that there is some consistent seasonal mismatch in the absolute
627 value of NPV fraction between phenocam and PlanetScope (Fig. 9). We hypothesize two main
628 reasons for this mismatch. First, the NPV endmember spectra derived from 3-m PlanetScope data
629 might differ from laboratory spectra of pure bare branch material. As a result, green leaves of
630 shorter crowns (i.e. understory) are in the background of a single bare crown and the leaves of
631 surrounding green crowns strongly transmit and reflect NIR preferentially onto an isolated bare
632 crown, raising apparent NIR reflectance in the PlanetScope-derived NPV endmember (Eriksson et
633 al., 2006). This leads to an overestimate of NPV fraction as part of green leaf information is
634 assigned to the NPV category in a linear spectral unmixing. Second, though the seasonal trend in
635 bare crown exposure was well correlated between the phenocam years of 2010-2011 and the
636 PlanetScope year of 2018/2019, there could be a difference in magnitude between these two time
637 periods (e.g. Fig. S12). Therefore, a further detailed monitoring and validation of tree-crown scale
638 leaf phenology (e.g. using drones; Park et al., 2019) is still needed, but beyond the scope of this
639 paper. Dry-season flowering in tropical trees (e.g. Borchert et al., 2005; Carvalho et al., 2013)
640 might also affect the estimation of NPV fraction. Whether the flowering would lead to an
641 overestimate of NPV fraction remains unknown, so more in-situ measurements of both flower
642 phenology and canopy reflectance of flowering canopies are still needed to help quantify the
643 flowering impacts. Nonetheless, flowers in the crowns of most Central Amazon trees occupy a
644 small fraction of crown area, and a recent study (Lopes et al., 2016) using a tower-mounted

645 phenocam in a Central Amazon forest near Manaus, Brazil found that flowers have little effect on
646 the seasonal change in ecosystem-scale “greenness”.

647 The proposed PlanetScope-MODIS integration for assessments of seasonal and spatial
648 dynamics in NPV fraction at the tree-crown scale also brings new opportunities to advance plant
649 ecology studies. First, it can improve our understanding of phenological scaling from individuals
650 to ecosystems (Nijland et al., 2016; Vrieling et al., 2017). Since ecosystem-scale phenology
651 emerges from the phenology of a community of tree species and individuals, several recent studies
652 have shown that the diversity in plant phenology at the fine scale can significantly affect the
653 ecosystem-scale phenology extracted, including the timing of key phenological events (e.g. leaf
654 on and off) (Chen et al., 2018) and the magnitude of seasonal fluctuations (e.g. Lopes et al., 2016;
655 Saleska et al., 2016). This not only applies to the temperate biomes, where ecosystem-scale
656 phenology shows large sensitivity to global climate change (Jeong et al., 2011; Körner and Basler,
657 2010; Thackeray et al., 2016), but is also important for the tropical biomes, where phenological
658 dynamics at the tree-crown level dominantly determine tropical forests’ ability to interact with the
659 climate system (e.g. Albert et al., 2018; Wright et al., 2017; Wu et al., 2016). The improved fine-
660 scale phenology monitoring as shown here thus offers a great opportunity to revisit these scaling
661 issues. Second, fine-scale NPV assessments also provide an important dataset to help parameterize,
662 constrain, and evaluate process-based models. Leaf phenology has been an important component
663 for process-based models to simulate large-scale climate-vegetation interactions (Fisher et al.,
664 2015; Restrepo-Coupe et al., 2017; Richardson et al., 2012). Yet the patterns and mechanisms of
665 leaf phenology over large scales remain poorly understood (Richardson et al., 2010; Xu et al., 2016,
666 2017). Once leaf phenology patterns have been derived at both fine and ecosystem scales, it
667 becomes possible to evaluate the competing mechanisms underlying current phenology models
668 and to parameterize process-based models for cross-scale simulations of carbon and water fluxes.

669 Despite these promising implications, our study also identifies four important next steps
670 that need to be considered for future advances. First, the robustness of this method in the wet
671 season is not yet assessed due to the frequent cloud covers in the high rainfall wet season of our
672 study site. There might be even fewer or no valid pixels for both MODIS and PlanetScope
673 measurements, resulting in higher uncertainty for MODIS gap-filling results in the wet season and
674 insufficient PlanetScope pixels for spectral cross-calibration using the histogram matching
675 approach (Fig.3). To (partly) resolve this issue, we recommend a stricter quality control for

676 MODIS and an improved cloud removal algorithm for PlanetScope (e.g. Planet’s UDM2
677 classification approach; Shendryk et al., 2019) to retain as many valid pixels and/or PlanetScope
678 in a monthly composite (assuming unchanged leaf phenology within a month) be needed to ensure
679 sufficient valid pixels for spectral cross-calibration. Second, the topography effects are not yet
680 considered in this study. Across large tropical areas, there are large variations in topography
681 (Jucker et al., 2018; Schwartz et al., 2019). Slope and aspect of the land relative to view and
682 illumination angles exert large effects on apparent land surface reflectance (Matsushita et al., 2007;
683 Wu et al., 2019a, 2019b). To avoid these complications, we focused on a large flat plateau site in
684 the current study. But it is thus important to explore whether the same method can be extended to
685 other regions with more accentuated topographic variation. Third, we used a fixed PlanetScope-
686 derived, endmember-specific reflectance spectra in linear spectral unmixing. Other unmixing
687 models (Asner et al., 2009; Roberts et al., 1998) accommodate variation in the reflectance spectra
688 of each endmember. Allowing for such variation is important for deriving a more broadly
689 applicable approach across large tropical areas. Fourth, our multi-sensor integration can enable
690 high-resolution monitoring of dry season dynamics in canopy-surface NPV and GV fractions.
691 However, it remains difficult to separate each individual tree crowns. Therefore, any approach to
692 enable tree-crown segregation or to combine other high-resolution orthorectified images (e.g.
693 drone or aerial photos) for tree-crown segregation (Klosterman et al., 2018; Park et al., 2019) will
694 make the derived fine-scale phenology metrics more useful.

695

696 **6. Conclusions**

697 This study develops a method to integrate PlanetScope with BRDF-adjusted MODIS to
698 enable cross-scale phenology monitoring in a Central Amazon tropical evergreen forest. The
699 method shares a similar concept as previous satellite image cross-sensor fusion/calibration work,
700 but also has three major advancements. First, it represents the first study in tropical evergreen
701 forests to integrate multi-satellites to enable fine-scale phenology monitoring. Second, we adopted
702 rigorous validation assessments to ensure that PlanetScope-MODIS integration worked
703 consistently well across all spatial scales. Third, the method also offers a metric with clear
704 biophysical meaning, i.e. the NPV fraction, to aid quantification of tropical leaf phenology.
705 Compared with other phenology monitoring methods, such as tower-mounted phenocams and
706 frequent drone flights, our integration not only aids detection of tree-crown scale leaf phenology

707 with high accuracy ($R^2=0.82$; Fig. 9), but also allows for leaf phenology monitoring to much larger
708 areas. These advantages make our method can be extended to other high resolution satellites and/or
709 other regions, advancing our ability to monitor land plant phenology and associated vegetation
710 dynamics in the context of global change.

711

712

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721

722 **References:**

- 723 Albert, L.P., Restrepo-Coupe, N., Smith, M.N., Wu, J., Chavana-Bryant, C., Prohaska, N., Taylor, T.C.,
724 Martins, G.A., Ciais, P., Mao, J.F., Arain, M.A., Li, W., Shi, X.Y., Ricciuto, D.M., Huxman, T.E.,
725 McMahon, S.M., Saleska, S.R., 2019. Cryptic phenology in plants: Case studies, implications, and
726 recommendations. *Global change biology*, 25, 3591-3608.
- 727 Albert, L.P., Wu, J., Prohaska, N., de Camargo, P.B., Huxman, T.E., Tribuzy, E.S., Ivanov, V.Y., Oliveira,
728 R.S., Garcia, S., Smith, M.N., Oliveira, R.C., Restrepo-Coupe, N., da Silva, R., Stark, S.C., Martins,
729 G.A., Penha, D.V., Saleska, S.R., 2018. Age-dependent leaf physiology and consequences for crown-
730 scale carbon uptake during the dry season in an Amazon evergreen forest. *New Phytologist*, 219, 870-
731 884.
- 732 Alberton, B., Torres, R.D., Cancian, L.F., Borges, B.D., Almeida, J., Mariano, G.C., dos Santos, J.,
733 Morellato, L.P.C., 2017. Introducing digital cameras to monitor plant phenology in the tropics:
734 Applications for conservation. *Perspectives in Ecology and Conservation*, 15, 82-90.
- 735 Asner, G.P., 1998. Biophysical and biochemical sources of variability in canopy reflectance. *Remote*
736 *Sensing of Environment*, 64, 234-253.
- 737 Asner, G. P., Knapp, D. E., Balaji, A., Páez-Acosta, G., 2009. Automated mapping of tropical deforestation
738 and forest degradation: CLASlite. *Journal of Applied Remote Sensing*, 3, 033543.
739 <http://doi.org/10.1117/1.3223675>.
- 740 Berra, E.F., Gaulton, R., Barr, S., 2019. Assessing spring phenology of a temperate woodland: A multiscale
741 comparison of ground, unmanned aerial vehicle and Landsat satellite observations. *Remote Sensing*
742 *of Environment*, 223, 229-242.
- 743 Blanchard, E., Birnbaum, P., Ibanez, T., Boutreux, T., Antin, C., Ploton, P., Vincent, G., Pouteau, R.,
744 Vandrot, H., Hequet, V., Barbier, N., Droissart, V., Sonke, B., Texier, N., Kamdem, N.G., Zebaze, D.,
745 Libalah, M., Coutron, P., 2016. Contrasted allometries between stem diameter, crown area, and tree
746 height in five tropical biogeographic areas. *Trees-Structure and Function*, 30, 1953-1968.

747 Bonan, G. B., 2008. Forests and climate change: forcings, feedbacks, and the climate benefits of forests.
748 science, 320, 1444-1449.

749 Borchert, R., Renner, S.S., Calle, Z., Navarrete, D., Tye, A., Gautier, L., Spichiger, R., von Hildebrand, P.,
750 2005. Photoperiodic induction of synchronous flowering near the Equator. *Nature*, 433, 627–629.

751 Brando, P.M., Goetz, S.J., Baccini, A., Nepstad, D.C., Beck, P.S.A., Christman, M.C., 2010. Seasonal and
752 interannual variability of climate and vegetation indices across the Amazon. *Proceedings of the*
753 *National Academy of Sciences of the United States of America*, 107, 14685-14690.

754 Carvalho, S., Schlerf, M., van Der Putten, W. H., and Skidmore, A. K., 2013. Hyperspectral reflectance of
755 leaves and flowers of an outbreak species discriminates season and successional stage of
756 vegetation. *International journal of applied earth observation and geoinformation*, 24, 32-41.

757 Chavez, P.S., Kwarteng, A.Y., 1989. Extracting spectral contrast in Landsat Thematic Mapper image data
758 using selective principal component analysis. *Photogrammetric Engineering and Remote Sensing*, 55,
759 339-348.

760 Chavez, P.S., Mackinnon, D.J., 1994. Automatic detection of vegetation changes in the southwestern
761 United States using remotely sensed images. *Photogrammetric Engineering and Remote Sensing*, 60,
762 571-583.

763 Chen, X., Wang, D.W., Chen, J., Wang, C., Shen, M.G., 2018. The mixed pixel effect in land surface
764 phenology: A simulation study. *Remote Sensing of Environment*, 211, 338-344.

765 Clark, M.L., Roberts, D.A., 2012. Species-level differences in hyperspectral metrics among tropical
766 rainforest trees as determined by a tree-based classifier. *Remote Sensing*, 4, 1820-1855.

767 de Moura, Y.M., Galvao, L.S., Hilker, T., Wu, J., Saleska, S., do Amaral, C.H., Nelson, B.W., Lopes, A.P.,
768 Wiedeman, K.K., Prohaska, N., de Oliveira, R.C., Machado, C.B., Aragao, L.E.O.C., 2017. Spectral
769 analysis of amazon canopy phenology during the dry season using a tower hyperspectral camera and
770 modis observations. *ISPRS Journal of Photogrammetry and Remote Sensing*, 131, 52-64.

771 Dedieu G., Karnieli A., Hagolle O., Jeanjean H., Cabot F., Ferrier P., Yaniv Y., 2006. Venüs : A joint
772 Israel-French Earth Observation scientific mission with high spatial and temporal resolution
773 capabilities. In: *The 2nd International Symposium on Recent Advances in Qualitative Remote Sensing*,
774 pp. 517-521.

775 Detto, M., Wright, S.J., Calderon, O., Muller-Landau, H.C., 2018. Resource acquisition and reproductive
776 strategies of tropical forest in response to the El Niño-Southern Oscillation. *Nature Communications*,
777 9, 913. <http://doi.org/10.1038/s41467-018-03306-9>.

778 Dribault, Y., Chokmani, K., Bernier, M., 2012. Monitoring seasonal hydrological dynamics of
779 minerotrophic peatlands using multi-date GeoEye-1 very high resolution imagery and object-based
780 classification. *Remote Sensing*, 4, 1887-1912.

781 Drusch, M., Del Bello, U., Carlier, S., Colin, O., Fernandez, V., Gascon, F., Hoersch, B., Isola, C., Laberinti,
782 P., Martimort, P., Meygret, A., Spoto, F., Sy, O., Marchese, F., Bargellini, P., 2012. Sentinel-2: ESA's
783 Optical High-Resolution Mission for GMES Operational Services. *Remote Sensing of Environment*,
784 120, 25-36.

785 Eamus, D., 1999. Ecophysiological traits of deciduous and evergreen woody species in the seasonally dry
786 tropics. *Trends in Ecology & Evolution*, 14, 11-16.

787 Eriksson, H.M., Eklundh, L., Kuusk, A., Nilson, T., 2006. Impact of understory vegetation on forest canopy
788 reflectance and remotely sensed LAI estimates. *Remote Sensing of Environment*. 103. 408-418.
789 <http://doi.org/10.1016/j.rse.2006.04.005>.

790 Feret, J.B., Corbane, C., Alleaume, S., 2015. Detecting the phenology and discriminating Mediterranean
791 natural habitats with multispectral sensors-An analysis based on multiseasonal field spectra. *IEEE*
792 *Journal of Selected Topics in Applied Earth Observations and Remote Sensing*, 8, 2294-2305.

793 Fisher, R.A., Muszala, S., Versteinstein, M., Lawrence, P., Xu, C., McDowell, N.G., Knox, R.G., Koven,
794 C., Holm, J., Rogers, B.M., Spessa, A., Lawrence, D., Bonan, G., 2015. Taking off the training wheels:
795 The properties of a dynamic vegetation model without climate envelopes, CLM4.5(ED). *Geoscientific*
796 *Model Development*, 8, 3593-3619.

797 Fraser, A. D., Massom, R. A., Michael, K. J., 2009. A method for compositing polar MODIS satellite
798 images to remove cloud cover for landfast sea-ice detection. *IEEE transactions on geoscience and*
799 *remote sensing*, 47, 3272-3282.

800 Galvao, L.S., dos Santos, J.R., Roberts, D.A., Breunig, F.M., Toomey, M., de Moura, Y.M., 2011. On intra-
801 annual EVI variability in the dry season of tropical forest: A case study with MODIS and hyperspectral
802 data. *Remote Sensing of Environment*, 115, 2350-2359.

803 Gao, F., Masek, J., Schwaller, M., Hall, F., 2006. On the blending of the Landsat and MODIS surface
804 reflectance: Predicting daily Landsat surface reflectance. *IEEE Transactions on Geoscience and*
805 *Remote Sensing*, 44, 2207-2218.

806 Gonçalves, N. B., Lopes, A. P., Dalagnol, R., Wu, J., Pinho, D. M., Nelson, B. W., 2020. Both near-surface
807 and satellite remote sensing confirm drought legacy effect on tropical forest leaf phenology after
808 2015/2016 ENSO drought. *Remote Sensing of Environment*, 237, 111489.
809 <http://doi.org/10.1016/j.rse.2019.111489>.

810 Gu, X.F., Tong, X.D., 2015. Overview of China Earth Observation Satellite Programs. *IEEE Geoscience*
811 *and Remote Sensing Magazine*, 3, 113-129.

812 Guan, K.Y., Pan, M., Li, H.B., Wolf, A., Wu, J., Medvigy, D., Caylor, K.K., Sheffield, J., Wood, E.F.,
813 Malhi, Y., Liang, M.L., Kimball, J.S., Saleska, S.R., Berry, J., Joiner, J., Lyapustin, A.I., 2015.
814 Photosynthetic seasonality of global tropical forests constrained by hydroclimate. *Nature Geoscience*,
815 8, 284-289.

816 Hillger, D. W., Clark, J. D., 2002. Principal component image analysis of MODIS for volcanic ash. Part I:
817 Most important bands and implications for future GOES imagers. *Journal of Applied Meteorology*, 41,
818 985-1001.

819 Houborg, R., McCabe, M. F., 2018a. A cubesat enabled spatio-temporal enhancement method (cestem)
820 utilizing planet, landsat and modis data. *Remote Sensing of Environment*, 209, 211-226.

821 Houborg, R., McCabe, M. F., 2018b. Daily Retrieval of NDVI and LAI at 3 m Resolution via the Fusion
822 of CubeSat, Landsat, and MODIS Data. *Remote Sensing*, 10, 890. <https://doi.org/10.3390/rs10060890>.

823 Huete, A.R., Didan, K., Miura, T., Rodriguez, E.P., Gao, X., Ferreira, L.G., 2002. Overview of the
824 radiometric and biophysical performance of the MODIS vegetation indices. *Remote Sensing of*
825 *Environment*, 83, 195-213.

826 Huete, A.R., Didan, K., Shimabukuro, Y.E., Ratana, P., Saleska, S.R., Hutyra, L.R., Yang, W.Z., Nemani,
827 R.R., Myneni, R., 2006. Amazon rainforests green-up with sunlight in dry season. *Geophysical*
828 *Research Letters*, 33, L06405. <https://doi.org/10.1029/2005GL025583>.

829 Hufkens, K., Friedl, M., Sonnentag, O., Braswell, B.H., Milliman, T., Richardson, A.D., 2012. Linking
830 near-surface and satellite remote sensing measurements of deciduous broadleaf forest phenology.
831 *Remote Sensing of Environment*, 117, 307-321.

832 Hutyra, L.R., Munger, J.W., Saleska, S.R., Gottlieb, E., Daube, B.C., Dunn, A.L., Amaral, D.F., de
833 Camargo, P.B., Wofsy, S.C., 2007. Seasonal controls on the exchange of carbon and water in an
834 Amazonian rain forest. *Journal of Geophysical Research*, 112,
835 G03008. <https://doi.org/10.1029/2006JG000365>.

836 Jeong, S.J., Ho, C.H., Gim, H.J., Brown, M.E., 2011. Phenology shifts at start vs. end of growing season in
837 temperate vegetation over the Northern Hemisphere for the period 1982-2008. *Global Change Biology*,
838 17, 2385-2399.

839 Jucker, T., Bongalov, B., Burslem, D.F.R.P., Nilus, R., Dalponte, M., Lewis, S.L., Phillips, O.L., Qie, L.,
840 Coomes, D.A., 2018. Topography shapes the structure, composition and function of tropical forest
841 landscapes. *Ecology Letters*, 21, 989-1000.

842 Jung, M., Schwalm, C., Migliavacca, M., Walther, S., Camps-Valls, G., Koirala, S., Anthoni, P., Besnard,
843 S., Bodesheim, P., Carvalhais, N., Chevallier, F., Gans, F., Groll, D. S., Haverd, V., Ichii, K., Jain, A.
844 K., Liu, J., Lombardozzi, D., Nabel, J. E. M. S., Nelson, J. A., Pallandt, M., Papale, D., Peters, W.,
845 Pongratz, J., Rödenbeck, C., Sitch, S., Tramontana, G., Weber, U., Reichstein, M., Koehler, P.,
846 O'Sullivan, M., and Walker, A., 2019. Scaling carbon fluxes from eddy covariance sites to globe:

847 Synthesis and evaluation of the FLUXCOM approach, *Biogeosciences Discuss.*
848 <https://doi.org/10.5194/bg-2019-368>.

849 Keshava, N., Mustard, J.F., 2002. Spectral unmixing. *IEEE Signal Processing Magazine*, 19, 44-57.

850 Klosterman, S., Melaas, E., Wang, J.A., Martinez, A., Frederick, S., O'Keefe, J., Orwig, D.A., Wang, Z.S.,
851 Sun, Q.S., Schaaf, C., Friedl, M., Richardson, A.D., 2018. Fine-scale perspectives on landscape
852 phenology from unmanned aerial vehicle (UAV) photography. *Agricultural and Forest Meteorology*,
853 248, 397-407.

854 Körner, C., Basler, D., 2010. Phenology Under Global Warming. *Science*, 327, 1461-1462.

855 Liao, C., Wang, J., Dong, T., Shang, J., Liu, J., Song, Y., 2019. Using spatio-temporal fusion of Landsat-8
856 and MODIS data to derive phenology, biomass and yield estimates for corn and soybean. *Science of*
857 *The Total Environment*, 650, 1707-1721.

858 Lopes, A.P., Nelson, B.W., Wu, J., Graca, P.M.L.D., Tavares, J.V., Prohaska, N., Martins, G.A., Saleska,
859 S.R., 2016. Leaf flush drives dry season green-up of the Central Amazon. *Remote Sensing of*
860 *Environment*, 182, 90-98.

861 Luo, Y.N., Guanb, K.Y., Pen, J., 2018. STAIR: A generic and fully-automated method to fuse multiple
862 sources of optical satellite data to generate a high-resolution, daily and cloud-/gap-free surface
863 reflectance product. *Remote Sensing of Environment*, 214, 87-99.

864 Maeda, E.E., Moura, Y.M., Wagner, F., Hilker, T., Lyapustin, A.I., Wang, Y.J., Chave, J., Mottus, M.,
865 Aragao, L.E.O.C., Shimabukuro, Y., 2016. Consistency of vegetation index seasonality across the
866 Amazon rainforest. *International Journal of Applied Earth Observation and Geoinformation*, 52, 42-
867 53.

868 Matsushita, B., Yang, W., Chen, J., Onda, Y., Qiu, G.Y., 2007. Sensitivity of the Enhanced Vegetation
869 Index (EVI) and Normalized Difference Vegetation Index (NDVI) to topographic effects: A case study
870 in high-density cypress forest. *Sensors*, 7, 2636-2651.

871 Moore, C.E., Brown, T., Keenan, T.F., Duursma, R.A., van Dijk, A.I.J.M., Beringer, J., Culvenor, D., Evans,
872 B., Huete, A., Hutley, L.B., Maier, S., Restrepo-Coupe, N., Sonnentag, O., Specht, A., Taylor, J.R.,
873 van Gorsel, E., Liddell, M.J., 2016. Reviews and syntheses: Australian vegetation phenology: New
874 insights from satellite remote sensing and digital repeat photography. *Biogeosciences*, 13, 5085-5102.

875 Morton, D.C., Nagol, J., Carabajal, C.C., Rosette, J., Palace, M., Cook, B.D., Vermote, E.F., Harding, D.J.,
876 North, P.R.J., 2014. Amazon forests maintain consistent canopy structure and greenness during the
877 dry season. *Nature*, 506, 221-224.

878 Moulin, S., Kergoat, L., Viovy, N., Dedieu, G., 1997. Global-scale assessment of vegetation phenology
879 using NOAA/AVHRR satellite measurements. *Journal of Climate*, 10, 1154-1170.

880 Nijland, W., Bolton, D.K., Coops, N.C., & Stenhouse, G., 2016. Imaging phenology; scaling from camera
881 plots to landscapes. *Remote Sensing of Environment*, 177, 13-20.

882 Otsu, N., 1979. Threshold selection method from gray-level histograms. *IEEE Transactions on Systems*
883 *Man and Cybernetics*, 9, 62-66.

884 Park, J.Y., Muller-Landau, H.C., Lichstein, J.W., Rifai, S.W., Dandois, J.P., Bohlman, S.A., 2019.
885 Quantifying leaf phenology of individual trees and species in a tropical forest using unmanned aerial
886 vehicle (UAV) Images. *Remote Sensing*, 11, 1534. <https://doi.org/10.3390/rs11131534>.

887 Planet, 2019. Planet Imagery Product Specification [WWW Document].
888 <https://assets.planet.com/docs/combined-imagery-product-spec-april-2019.pdf> (accessed 03 March
889 2020).

890 Planet Team, 2018. Planet Application Program Interface: In Space for Life on Earth. In. San Francisco,
891 CA. <https://api.planet.com> (accessed 02 March 2020).

892 Pu, R.L., Landry, S., Yu, Q.Y., 2018. Assessing the potential of multi-seasonal high resolution Pleiades
893 satellite imagery for mapping urban tree species. *International Journal of Applied Earth Observation*
894 *and Geoinformation*, 71, 144-158.

895 Reich, P. B., 1995. Phenology of tropical forests: patterns, causes, and consequences. *Canadian Journal of*
896 *Botany*, 73, 164-174. <https://doi.org/10.1139/b95-020>.

897 Restrepo-Coupe, N., da Rocha, H.R., Hutyrá, L.R., da Araujo, A.C., Borma, L.S., Christoffersen, B., Cabral,
898 O.M.R., de Camargo, P.B., Cardoso, F.L., da Costa, A.C.L., Fitzjarrald, D.R., Goulden, M.L., Kruijft,
899 B., Maia, J.M.F., Malhi, Y.S., Manzi, A.O., Miller, S.D., Nobre, A.D., von Randow, C., Sa, L.D.A.,
900 Sakai, R.K., Tota, J., Wofsy, S.C., Zanchi, F.B., Saleska, S.R., 2013. What drives the seasonality of
901 photosynthesis across the Amazon basin? A cross-site analysis of eddy flux tower measurements from
902 the Brasil flux network. *Agricultural and Forest Meteorology*, 182, 128-144.

903 Restrepo-Coupe, N., Levine, N.M., Christoffersen, B.O., Albert, L.P., Wu, J., Costa, M.H., Galbraith, D.,
904 Imbuzeiro, H., Martins, G., da Araujo, A.C., Malhi, Y.S., Zeng, X.B., Moorcroft, P., Saleska, S.R.,
905 2017. Do dynamic global vegetation models capture the seasonality of carbon fluxes in the Amazon
906 basin? A data-model intercomparison. *Global Change Biology*, 23, 191-208.

907 Rice, A.H., Pyle, E.H., Saleska, S.R., Hutyrá, L., Palace, M., Keller, M., de Camargo, P.B., Portilho, K.,
908 Marques, D.F., Wofsy, S.C., 2004. Carbon balance and vegetation dynamics in an old-growth
909 Amazonian forest. *Ecological Applications*, 14, 55-71.

910 Richardson, A.D., Anderson, R.S., Arain, M.A., Barr, A.G., Bohrer, G., Chen, G.S., Chen, J.M., Ciais, P.,
911 Davis, K.J., Desai, A.R., Dietze, M.C., Dragoni, D., Garrity, S.R., Gough, C.M., Grant, R., Hollinger,
912 D.Y., Margolis, H.A., McCaughey, H., Migliavacca, M., Monson, R.K., Munger, J.W., Poulter, B.,
913 Raczka, B.M., Ricciuto, D.M., Sahoo, A.K., Schaefer, K., Tian, H.Q., Vargas, R., Verbeeck, H., Xiao,
914 J.F., Xue, Y.K., 2012. Terrestrial biosphere models need better representation of vegetation phenology:
915 Results from the North American Carbon Program Site Synthesis. *Global Change Biology*, 18, 566-
916 584.

917 Richardson, A.D., Black, T.A., Ciais, P., Delbart, N., Friedl, M.A., Gobron, N., Hollinger, D.Y., Kutsch,
918 W.L., Longdoz, B., Luysaert, S., Migliavacca, M., Montagnani, L., Munger, J.W., Moors, E., Piao,
919 S.L., Rebmann, C., Reichstein, M., Saigusa, N., Tomelleri, E., Vargas, R., Varlagin, A., 2010.
920 Influence of spring and autumn phenological transitions on forest ecosystem productivity.
921 *Philosophical Transactions of the Royal Society B*, 365, 3227-3246.

922 Richardson, A.D., Hufkens, K., Milliman, T., Aubrecht, D.M., Chen, M., Gray, J.M., Johnston, M.R.,
923 Keenan, T.F., Klosterman, S.T., Kosmala, M., Melaas, E.K., Friedl, M.A., Frohling, S., 2018. Tracking
924 vegetation phenology across diverse North American biomes using PhenoCam imagery. *Scientific*
925 *Data*, 5, 180028. <https://doi.org/10.1038/sdata.2018.28>.

926 Roberts, D. A., Gardner, M., Church, R., Ustin, S., Scheer, G., Green, R. O., 1998. Mapping chaparral in
927 the Santa Monica Mountains using Multiple Endmember Spectral Mixture models. *Remote Sensing*
928 *of Environment*, 65, 267–279.

929 Roberts, D. A., Smith, M. O., Sabol, D. E., Adams, J. B., and Ustin, S. L., 1992. Mapping the spectral
930 variability in photosynthetic and non-photosynthetic vegetation, soils, and shade using AVIRIS. In:
931 *Summaries 3rd Annual JPL Airborne Geoscience Workshop*, pp 38–40.

932 Saleska, S.R., Wu, J., Guan, K.Y., Araujo, A.C., Huete, A., Nobre, A.D., Restrepo-Coupe, N., 2016. Dry-
933 season greening of Amazon forests. *Nature*, 531, E4-E5. <https://doi.org/10.1038/nature16457>.

934 Samanta, A., Ganguly, S., Hashimoto, H., Devadiga, S., Vermote, E., Knyazikhin, Y., Nemani, R.R.,
935 Myneni, R.B., 2010. Amazon forests did not green-up during the 2005 drought. *Geophysical Research*
936 *Letters*, 37. <http://doi.org/10.1029/2009GL042154>.

937 Samanta, A., Ganguly, S., Vermote, E., Nemani, R.R., Myneni, R.B., 2012. Interpretation of variations in
938 MODIS-measured greenness levels of Amazon forests during 2000 to 2009. *Environmental Research*
939 *Letters*, 7, 024018. <https://doi.org/10.1088/1748-9326/7/2/024018>.

940 Schaaf, C.B., Gao, F., Strahler, A.H., Lucht, W., Li, X.W., Tsang, T., Strugnell, N.C., Zhang, X.Y., Jin,
941 Y.F., Muller, J.P., Lewis, P., Barnsley, M., Hobson, P., Disney, M., Roberts, G., Dunderdale, M., Doll,
942 C., d'Entremont, R.P., Hu, B.X., Liang, S.L., Privette, J.L., Roy, D., 2002. First operational BRDF,
943 albedo nadir reflectance products from MODIS. *Remote Sensing of Environment*, 83, 135-148.

944 Schaaf, C.B., Liu, J., Gao, F., Strahler, A.H., 2011. MODIS albedo and reflectance anisotropy products
945 from Aqua and Terra. *Land Remote Sensing and Global Environmental Change: NASA's Earth*
946 *Observing System and the Science of ASTER and MODIS*, 11, 549-561.

947 Schwartz, N.B., Budsock, A.M., Uriarte, M., 2019. Fragmentation, forest structure, and topography
948 modulate impacts of drought in a tropical forest landscape. *Ecology*, 100, e02677.
949 <http://doi.org/10.1002/ecy.2677>.

950 Semmens, K.A., Anderson, M.C., Kustas, W.P., Gao, F., Alfieri, J.G., McKee, L., Prueger, J.H., Hain, C.R.,
951 Cammalleri, C., Yang, Y., Xia, T., Sanchez, L., Alsina, M.M., Velez, M., 2016. Monitoring daily
952 evapotranspiration over two California vineyards using Landsat 8 in a multi-sensor data fusion
953 approach. *Remote Sensing of Environment*, 185, 155-170.

954 Shendryk, Y., Rist, Y., Ticehurst, C., Thorburn, P., 2019. Deep learning for multi-modal classification of
955 cloud, shadow and land cover scenes in PlanetScope and Sentinel-2 imagery. *ISPRS Journal of*
956 *Photogrammetry and Remote Sensing*, 157, 124-136.

957 Thackeray, S.J., Henrys, P.A., Hemming, D., Bell, J.R., Botham, M.S., Burthe, S., Helaouet, P., Johns,
958 D.G., Jones, I.D., Leech, D.I., Mackay, E.B., Massimino, D., Atkinson, S., Bacon, P.J., Brereton, T.M.,
959 Carvalho, L., Clutton-Brock, T.H., Duck, C., Edwards, M., Elliott, J.M., Hall, S.J.G., Harrington, R.,
960 Pearce-Higgins, J.W., Hoye, T.T., Kruuk, L.E.B., Pemberton, J.M., Sparks, T.H., Thompson, P.M.,
961 White, I., Winfield, I.J., Wanless, S., 2016. Phenological sensitivity to climate across taxa and trophic
962 levels. *Nature*, 535, 241-245.

963 Viennois, G., Barbier, N., Fabre, I., Couteron, P., 2013. Multiresolution quantification of deciduousness in
964 West-Central African forests. *Biogeosciences*, 10, 6957-6967.

965 Vrieling, A., Skidmore, A.K., Wang, T.J., Meroni, M., Ens, B.J., Oosterbeek, K., O'Connor, B.,
966 Darvishzadeh, R., Heurich, M., Shepherd, A., & Paganini, M., 2017. Spatially detailed retrievals of
967 spring phenology from single-season high-resolution image time series. *International Journal of*
968 *Applied Earth Observation and Geoinformation*, 59, 19-30.

969 Wagner, F.H., Herault, B., Bonal, D., Stahl, C., Anderson, L.O., Baker, T.R., Becker, G.S., Beeckman, H.,
970 Souza, D.B., Botosso, P.C., Bowman, D.M.J.S., Brauning, A., Brede, B., Brown, F.I., Camarero, J.J.,
971 Camargo, P.B., Cardoso, F.C.G., Carvalho, F.A., Castro, W., Chagas, R.K., Chave, J., Chidumayo,
972 E.N., Clark, D.A., Costa, F.R.C., Couralet, C., Mauricio, P.H.D., Dalitz, H., de Castro, V.R., Milani,
973 J.E.D., de Oliveira, E.C., Arruda, L.D.S., Devineau, J.L., Drew, D.M., Dunisch, O., Durigan, G.,
974 Elifuraha, E., Fedele, M., Fedele, L.F., Figueiredo, A., Finger, C.A.G., Franco, A.C., Freitas, J.L.,
975 Galvao, F., Gebrekirstos, A., Gliniars, R., Graca, P.M.L.D., Griffiths, A.D., Grogan, J., Guan, K.,
976 Homeier, J., Kanieski, M.R., Kho, L.K., Koenig, J., Kohler, S.V., Krepkowski, J., Lemos, J.P.,
977 Lieberman, D., Lieberman, M.E., Lisi, C.S., Santos, T.L., Ayala, J.L.L., Maeda, E.E., Malhi, Y., Maria,
978 V.R.B., Marques, M.C.M., Marques, R., Chamba, H.M., Mbwambo, L., Melgaco, K.L.L., Mendivelso,
979 H.A., Murphy, B.P., O'Brien, J.J., Oberbauer, S.F., Okada, N., Pelissier, R., Prior, L.D., Roig, F.A.,
980 Ross, M., Rossatto, D.R., Rossi, V., Rowland, L., Rutishauser, E., Santana, H., Schulze, M., Selhorst,
981 D., Silva, W.R., Silveira, M., Spann, S., Swaine, M.D., Toledo, J.J., Toledo, M.M., Toledo, M., Toma,
982 T., Tomazello, M., Hernandez, J.I.V., Verbesselt, J., Vieira, S.A., Vincent, G., de Castilho, C.V.,
983 Volland, F., Worbes, M., Zanon, M.L.B., Aragao, L.E.O.C., 2016. Climate seasonality limits leaf
984 carbon assimilation and wood productivity in tropical forests. *Biogeosciences*, 13, 2537-2562.

985 Walker, J.J., de Beurs, K.M., Wynne, R.H., Gao, F., 2012. Evaluation of Landsat and MODIS data fusion
986 products for analysis of dryland forest phenology. *Remote Sensing of Environment*, 117, 381-393.

987 Wanner, W., Li, X., Strahler, A. H., 1995. On the derivation of kernels for kernel-driven models of
988 bidirectional reflectance. *Journal of Geophysical Research: Atmospheres*, 100, 21077-21089.

989 White, M.A., de Beurs, K.M., Didan, K., Inouye, D.W., Richardson, A.D., Jensen, O.P., O'Keefe, J., Zhang,
990 G., Nemani, R.R., van Leeuwen, W.J.D., Brown, J.F., de Wit, A., Schaepman, M., Lin, X.M.,
991 Dettinger, M., Bailey, A.S., Kimball, J., Schwartz, M.D., Baldocchi, D.D., Lee, J.T., Lauenroth, W.K.,
992 2009. Intercomparison, interpretation, and assessment of spring phenology in North America
993 estimated from remote sensing for 1982-2006. *Global Change Biology*, 15, 2335-2359.

994 Wright, J.S., Fu, R., Worden, J. R., Chakraborty, S., Clinton, N. E., Risi, C., Sun, Y., Yin, L., 2017.
995 Rainforest-initiated wet season onset over the southern Amazon. *Proceedings of the National Academy*
996 *of Sciences*, 114, 8481-8486.

997 Wu, J., Albert, L.P., Lopes, A.P., Restrepo-Coupe, N., Hayek, M., Wiedemann, K.T., Guan, K.Y., Stark,
 998 S.C., Christoffersen, B., Prohaska, N., Tavares, J.V., Marostica, S., Kobayashi, H., Ferreira, M.L.,
 999 Campos, K.S., da Silva, R., Brando, P.M., Dye, D.G., Huxman, T.E., Huete, A.R., Nelson, B.W.,
 1000 Saleska, S.R., 2016. Leaf development and demography explain photosynthetic seasonality in Amazon
 1001 evergreen forests. *Science*, 351, 972-976.
 1002 Wu, J., Kobayashi, H., Stark, S.C., Meng, R., Guan, K.Y., Tran, N.N., Gao, S.C., Yang, W., Restrepo-
 1003 Coupe, N., Miura, T., Oliviera, R.C., Rogers, A., Dye, D.G., Nelson, B.W., Serbin, S.P., Huete, A.R.,
 1004 Saleska, S.R., 2018. Biological processes dominate seasonality of remotely sensed canopy greenness
 1005 in an Amazon evergreen forest. *New Phytologist*, 217, 1507-1520.
 1006 Wu, S.B., Wen, J.G., Gastellu-Etchegorry, J.P., Liu, Q.H., You, D.Q., Xiao, Q., Hao, D.L., Lin, X.W., Yin,
 1007 T.G., 2019a. The definition of remotely sensed reflectance quantities suitable for rugged terrain.
 1008 *Remote Sensing of Environment*, 225, 403-415.
 1009 Wu, S.B., Wen, J.G., Lin, X.W., Hao, D.L., You, D.Q., Xiao, Q., Liu, Q.H., Yin, T.G., 2019b. Modeling
 1010 discrete forest anisotropic reflectance over a sloped surface with an extended GOMS and SAIL model.
 1011 *IEEE Transactions on Geoscience and Remote Sensing*, 57, 944-957.
 1012 Xu, X.T., Medvigy, D., Powers, J.S., Becknell, J.M., Guan, K.Y., 2016. Diversity in plant hydraulic traits
 1013 explains seasonal and inter-annual variations of vegetation dynamics in seasonally dry tropical forests.
 1014 *New Phytologist*, 212, 80-95.
 1015 Xu, X.T., Medvigy, D., Wright, S.J., Kitajima, K., Wu, J., Albert, L.P., Martins, G.A., Saleska, S.R., Pacala,
 1016 S.W., 2017. Variations of leaf longevity in tropical moist forests predicted by a trait-driven carbon
 1017 optimality model. *Ecology Letters*, 20, 1097-1106.
 1018 Yang, X.J., Lo, C.P., 2000. Relative radiometric normalization performance for change detection from
 1019 multi-date satellite images. *Photogrammetric Engineering and Remote Sensing*, 66, 967-980.
 1020 Yang, Y., Anderson, M.C., Gao, F., Hain, C.R., Semmens, K.A., Kustas, W.P., Noormets, A., Wynne, R.H.,
 1021 Thomas, V.A., Sun, G., 2017. Daily Landsat-scale evapotranspiration estimation over a forested
 1022 landscape in North Carolina, USA, using multi-satellite data fusion. *Hydrology and Earth System
 1023 Sciences*, 21, 1017-1037.
 1024 Zeng, Z.Z., Estes, L., Ziegler, A.D., Chen, A.P., Searchinger, T., Hua, F.Y., Guan, K.Y., Jintrawet, A.,
 1025 Wood, E.F., 2018. Highland cropland expansion and forest loss in Southeast Asia in the twenty-first
 1026 century. *Nature Geoscience*, 11, 556-562.
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Tables and Figure captions

1030 **Table 1.** Spatial resolutions, accessed data time ranges, and spectral bands and band-specific
1031 wavelength ranges of PlanetScope and MODIS data used at the k67 site.

1032

Satellite	Spatial resolution (m)	Accessed data time range	Spectral band and wavelength range (nm)			
			Blue	Green	Red	NIR
PlanetScope	3	06/2018-11/2018 06/2019-11/2019	455-515	500-590	590-670	780-860
MODIS	500	02/2000-12/2019	459-479	545-565	620-670	841-876

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1035 **Figure captions**

1036 **Figure 1.** Locations and multi-scale observations at the k67 tower site and Alter do Chão site in a
1037 Central Amazon evergreen forest in Brazil. (a) The locations of the study sites, including the red
1038 pentagram for the k67 site and green square for the Alter do Chão site; and the multi-scale
1039 observations include (b) tower-mounted phenocam (temporal coverage: 2010-2011; spatial
1040 coverage: about 200m×300m) at the k67 site, which was equipped with a 3-band (NIR, red, and
1041 green) Tetracam Agricultural Digital Camera, and thus green vegetation in the camera image (false
1042 red composited by RGB=NIR-red-green) looks red; (c) PlanetScope data of near daily nadir
1043 coverage at a 3m spatial resolution at the k67 site (temporal coverage: dry season of 2018 and
1044 2019; spatial coverage: 10km×10km); (d) daily MODIS data of a 500m spatial resolution at the
1045 k67 site (temporal coverage: 2000-2019; spatial coverage: 10km×10km); (e) PlanetScope data at
1046 the Alter do Chão site (temporal coverage: dry season of 2018; spatial coverage: 8km×8km); and
1047 (f) MODIS data at the Alter do Chão site (temporal coverage: 2000-2019; spatial coverage:
1048 8km×8km). The background figure in panel (a) is adapted from National Geographic, ESRI;
1049 phenocam data can be accessed from Wu et al. (2016); satellite data of the two sites in panel (c)-
1050 (f) are displayed in the same false red composite as phenocam; the k67 site is used to evaluate the
1051 multi-scale approach for tropical phenology monitoring, and the Alter do Chão site is used to
1052 evaluate robustness of the approach.

1053 **Figure 2.** Flowchart of the method. It includes four major tasks: 1) acquiring and processing the
1054 PlanetScope and MODIS data, 2) cross-calibrating the PlanetScope data using BRDF-adjusted
1055 MODIS, 3) extracting reflectance spectra of the three key endmembers comprising tropical forest
1056 canopies and estimating the fractions of non-photosynthetic vegetation (NPV) and green
1057 vegetation (GV) of each pixel in the calibrated PlanetScope images, and 4) evaluating the accuracy
1058 of PlanetScope-derived seasonal trends in NPV and GV fractions by comparing with ground-based
1059 measurements of leaf phenology.

1060 **Figure 3.** Example demonstration of band-specific outlier detection and gap-filling for BRDF-
1061 adjusted MODIS seasonality in 2018, including reflectance bands of (a) blue, (b) green, (c) red,
1062 and (d) NIR. The 20-year (i.e. 2000-2019) mean MODIS seasonality is displayed with the mean
1063 values in black lines, and grey shading for 95% confidence interval (i.e. 2.5 percentile for bottom
1064 and 97.5 percentile for top); the 2018 seasonality after quality control is displayed with valid data
1065 points (blue crosses, within the 95% confidence interval of 20-year mean seasonality) and outlier
1066 data points (blue circles, outside the 95% confidence interval); the gap-filling results in 2018 are
1067 displayed as red lines for daily means and red shadings for 95% confidence interval. A 5km×5km
1068 area centered on the k67 tower site is used here for demonstration purposes. Light grey shading
1069 indicates the dry season of the k67 site.

1070 **Figure 4.** Example demonstration of band-specific histograms and the fitted Gaussian distribution
1071 curves for BRDF-adjusted MODIS (shown as red) and upscaled PlanetScope (shown as green) on
1072 November 21, 2018, including reflectance bands of (a) blue, (b) green, (c) red, and (d) NIR. The
1073 spatial extent used here includes a 10km×10km area centered on the k67 tower site. BRDF-
1074 adjusted MODIS has a spatial resolution of 500m; upscaled PlanetScope refers to upscaling the
1075 original PlanetScope of a 3m spatial resolution to the same spatial resolution as that of MODIS;

1076 the probability distribution function (PDF) is used to describe the fitted Gaussian distribution
1077 curves.

1078 **Figure 5.** Ecosystem-scale seasonality of BRDF-adjusted MODIS (20-year mean in black and its
1079 95% confidence interval in grey shading, and 2018 gap-filled in red) and 2018 PlanetScope
1080 (uncalibrated in blue and calibrated in green), including reflectance bands of (a) blue, (b) green,
1081 (c) red, and (d) NIR, and vegetation indices of (e) Normalized Difference Vegetation Index
1082 (NDVI), and (f) Enhanced Vegetation Index (EVI). BRDF-adjusted MODIS including both 20-
1083 year mean and 2018 gap-filled are displayed as background information; the
1084 uncalibrated/calibrated PlanetScope is based on the histogram matching analysis as shown in Fig.
1085 4; and a 3km×3km area centered on the k67 tower site is used here to calculate ecosystem-scale
1086 seasonality. Light grey shading indicates the dry season of the k67 site.

1087 **Figure 6.** Seasonal variation in PlanetScope-derived reflectance spectra of stable permanent
1088 objects (urban buildings) (a) prior and (b) post cross-calibration. The 180 building pixels were
1089 carefully and manually extracted from an area of 8km×8km centered on the town of Alter do Chão,
1090 Brazil, which is 37 km from the k67 tower site (see Fig. 1 for more details). Error bars indicate
1091 one standard deviation for the reflectance spectra among all building pixels; each colored line
1092 indicates one of nine selected dates of PlanetScope measurements in 2018; these selected dates
1093 cover the full dry season at the k67 site and the Alter do Chão site.

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1095 **Figure 7.** The extracted mean (color lines) and one standard deviation (error bars) of reflectance
1096 spectra for three key endmembers at the k67 site using all the calibrated PlanetScope data from
1097 June to November in 2018 and 2019. These three endmembers are pixels of completely leafless
1098 tree-crowns for pure non-photosynthetic vegetation (NPV), leafy tree-crowns for pure green
1099 vegetation (GV), and deep shade/shadow portions of the canopy (shade).

1100 **Figure 8.** Spatial and temporal variations in canopy reflectance for calibrated PlanetScope (top
1101 panel) and corresponding magnified PlanetScope (bottom panel) (composited as RGB=NIR-red-
1102 blue). For demonstration purposes, we selected 6 dates among all 22 dates of PlanetScope
1103 measurements during the dry season of 2018, including one date per month from June to November
1104 (i.e. June 15, July 06, August 20, September 20, October 15, and November 01 from left to right);
1105 each image subset (shown in top panel) centered on the k67 tower site has a spatial coverage of
1106 500m×500m (=one MODIS pixel).

1107
1108 **Figure 9.** Comparisons of PlanetScope-derived phenology metrics and ground-based phenology
1109 measurements, including a) PlanetScope-derived and tower-phenocam measurements of NPV
1110 fraction, and b) PlanetScope-derived and tower-phenocam measurements of GV fraction and field
1111 LAI measurements. The calibrated PlanetScope data in 2018 are used here, and the PlanetScope-
1112 derived phenology metrics represent an average of a 3km×3km area centered on the k67 site; error
1113 bars indicate one standard deviation. Tower-phenocam measurements in 2010-2011 (conducted in
1114 an about 200m×300m area centered on the k67 site) and field LAI measurements in 2000-2005 (a
1115 100m×100m plot, ~5 km apart from the k67 site) are based on the literature values (Brando et al.,
1116 2010; Wu et al., 2016; see methods for details).

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1118 **Figure 10.** Spatial and temporal variations in PlanetScope-derived NPV fraction in an area of
1119 500m×500m centered on the k67 tower site (top panel) and in eight pixels located in the same area
1120 (bottom panel). The spatial extent shown in the top panel is the same as Fig. 8, and 6 dates among
1121 a total of 22 dates in the dry season of 2018 are selected, including one date per month from June
1122 to November (June 15, July 06, August 20, September 20, October 15, and November 01 from left
1123 to right). The gray bar with a range from 0 to 1 represents an increasing NPV fraction from 0 to
1124 100% within an image pixel. For demonstration purposes, spatial and temporal dynamics of NPV
1125 fraction in eight pixels (A to H) are shown in the bottom panel, with the ecosystem-scale average
1126 NPV fraction of a 3km×3km area centered on the k67 site shown in black line; error bars indicate
1127 one standard deviation.

1128
1129 **Figure 11.** Assessing the dry-season change rate in NPV fraction derived from the calibrated
1130 PlanetScope data in 2018, including (a) a map of dry-season change rate in NPV fraction for an
1131 area of 500m×500m centered on the k67 tower site (the same as Fig. 10), and (b) statistical
1132 summary on frequency distribution of dry-season change rate for all pixels shown in panel (a). The
1133 dry-season change rate in NPV fraction is assessed through a linear regression analysis between
1134 dry-season change in NPV fraction and day of year in the dry season. The dry-season change rate
1135 in NPV fraction has a range from -0.005 to 0.005, with a negative value indicating dry-season
1136 decrease in NPV fraction (or green-up) and a positive value rate indicating dry-season increase in
1137 NPV fraction (or brown-down). In this area, there are 84.6% of pixels showing a dry-season green-
1138 up trend and 15.4% showing a brown-down trend.

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Figures and Tables

Figure 1. Locations and multi-scale observations at the k67 tower site and Alter do Chão site in a Central Amazon evergreen forest in Brazil. (a) The locations of the study sites, including the red pentagram for the k67 site and green square for the Alter do Chão site; and the multi-scale observations include (b) tower-mounted phenocam (temporal coverage: 2010-2011; spatial coverage: about 200m×300m) at the k67 site, which was equipped with a 3-band (NIR, red, and green) Tetracam Agricultural Digital Camera, and thus green vegetation in the camera image (false red composited by RGB=NIR-red-green) looks red; (c) PlanetScope data of near daily nadir coverage at a 3m spatial resolution at the k67 site (temporal coverage: dry season of 2018 and 2019; spatial coverage: 10km×10km); (d) daily MODIS data of a 500m spatial resolution at the k67 site (temporal coverage: 2000-2019; spatial coverage: 10km×10km); (e) PlanetScope data at the Alter do Chão site (temporal coverage: dry season of 2018; spatial coverage: 8km×8km); and (f) MODIS data at the Alter do Chão site (temporal coverage: 2000-2019; spatial coverage: 8km×8km). The background figure in panel (a) is adapted from National Geographic, ESRI; phenocam data can be accessed from Wu et al. (2016); satellite data of the two sites in panel (c)-(f) are displayed in the same false red composite as phenocam; the k67 site is used to evaluate the multi-scale approach for tropical phenology monitoring, and the Alter do Chão site is used to evaluate robustness of the approach.

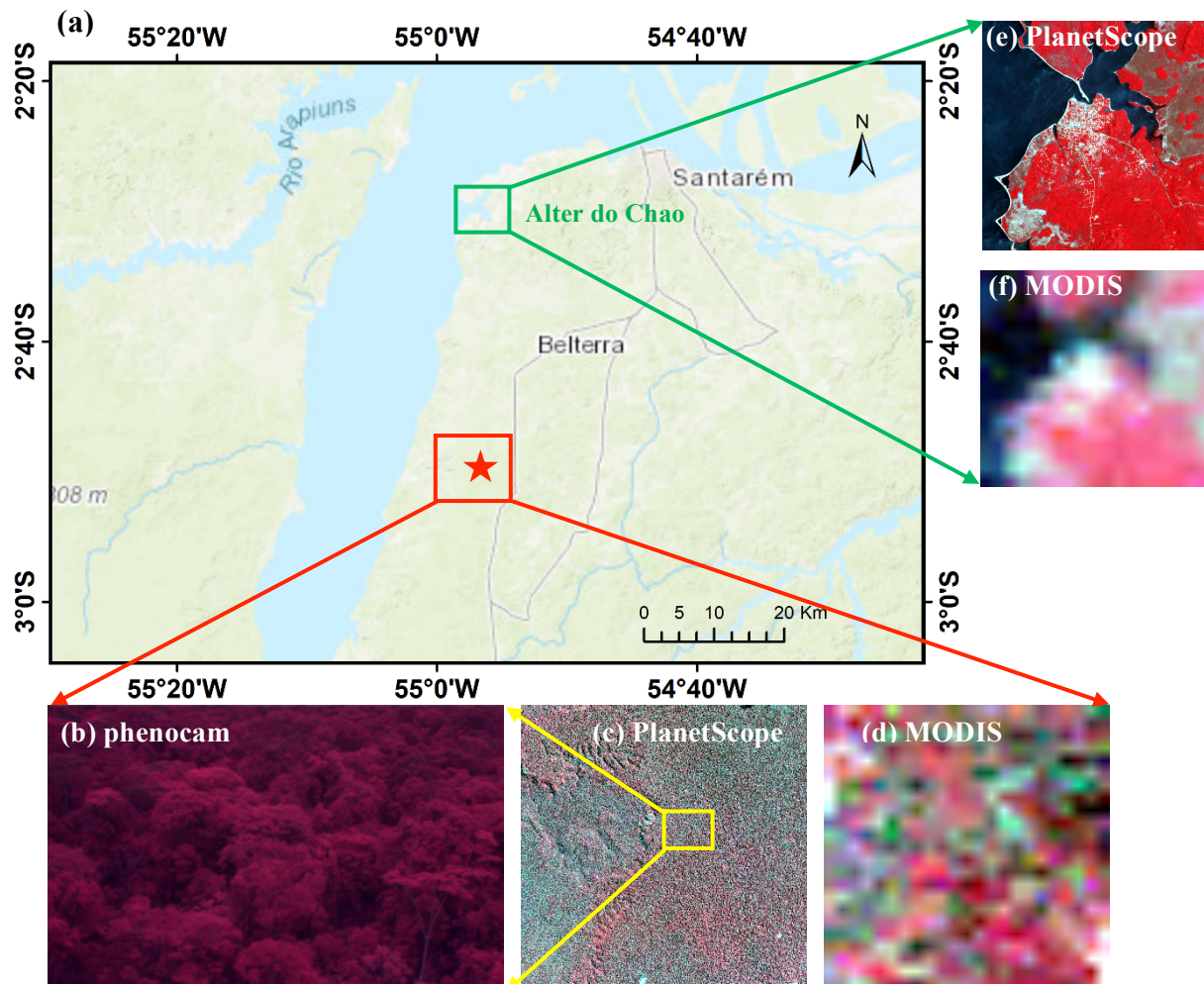


Figure 2. Flowchart of the method. It includes four major tasks: 1) acquiring and processing the PlanetScope and MODIS data, 2) cross-calibrating the PlanetScope data using BRDF-adjusted MODIS, 3) extracting reflectance spectra of the three key endmembers comprising tropical forest canopies and estimating the fractions of non-photosynthetic vegetation (NPV) and green vegetation (GV) of each pixel in the calibrated PlanetScope images, and 4) evaluating the accuracy of PlanetScope-derived seasonal trends in NPV and GV fractions by comparing with ground-based measurements of leaf phenology.

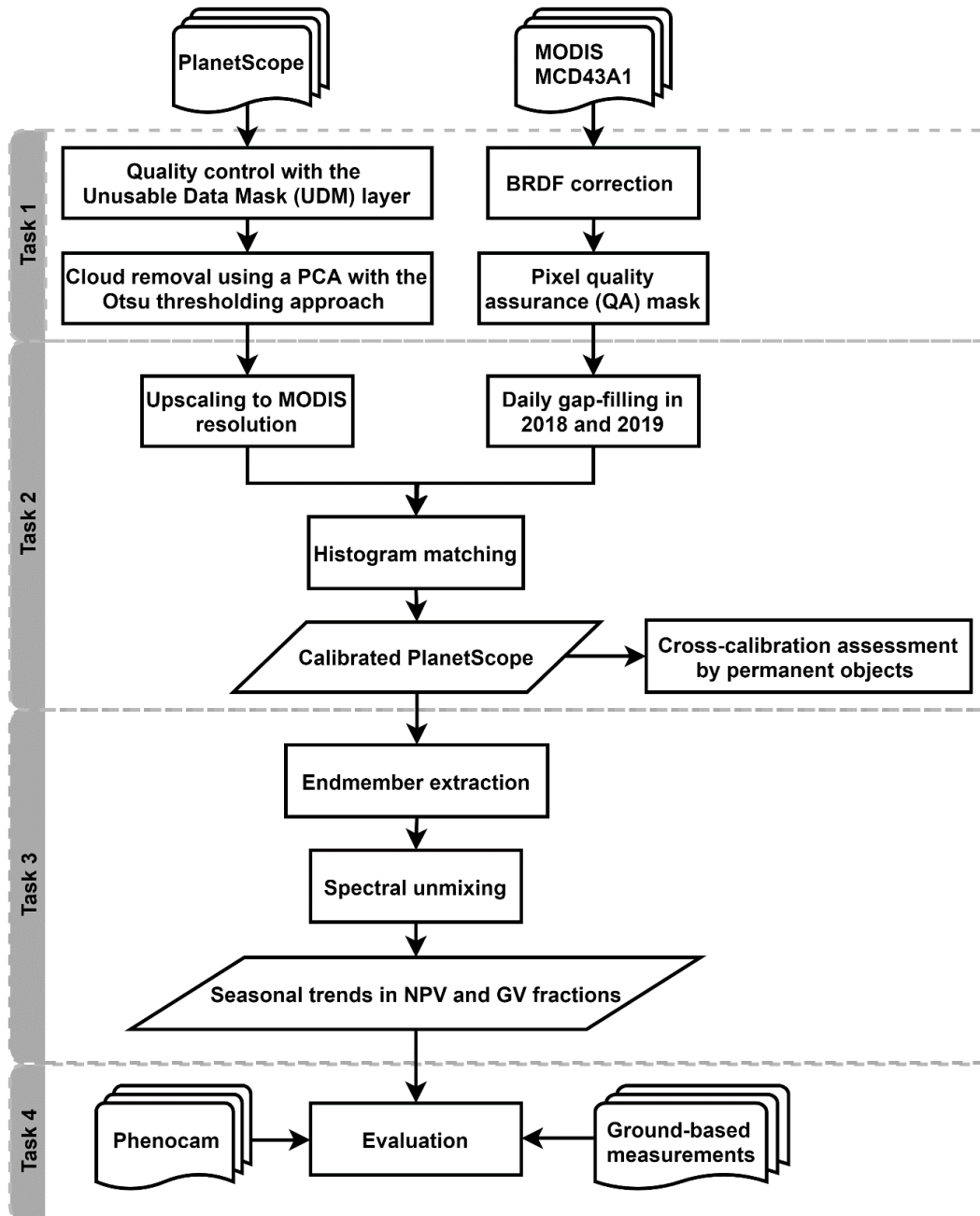


Figure 3. Example demonstration of band-specific outlier detection and gap-filling for BRDF-adjusted MODIS seasonality in 2018, including reflectance bands of (a) blue, (b) green, (c) red, and (d) NIR. The 20-year (i.e. 2000-2019) mean MODIS seasonality is displayed with the mean values in black lines, and grey shading for 95% confidence interval (i.e. 2.5 percentile for bottom and 97.5 percentile for top); the 2018 seasonality after quality control is displayed with valid data points (blue crosses, within the 95% confidence interval of 20-year mean seasonality) and outlier data points (blue circles, outside the 95% confidence interval); the gap-filling results in 2018 are displayed as red lines for daily means and red shadings for 95% confidence interval. A 5km×5km area centered on the k67 tower site is used here for demonstration purposes. Light grey shading indicates the dry season of the k67 site.

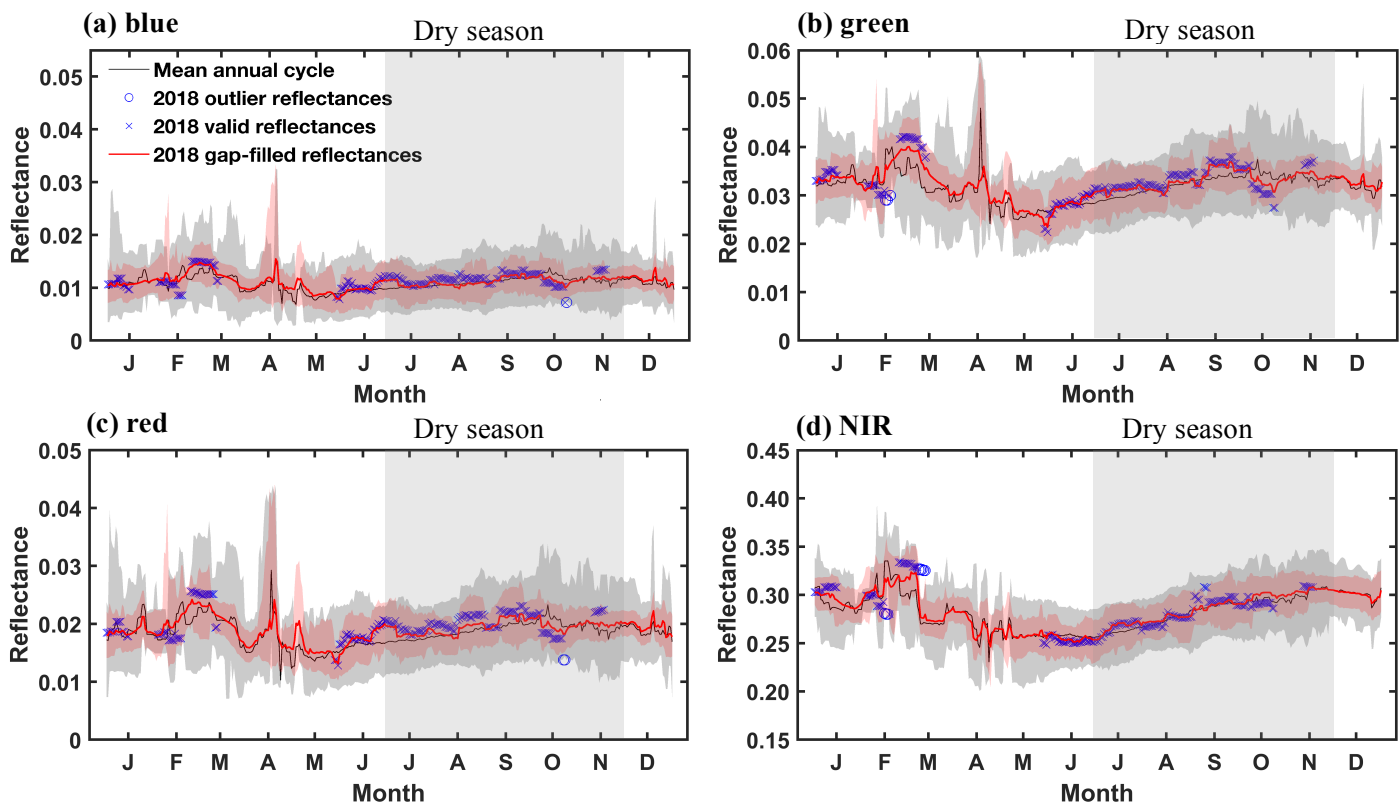


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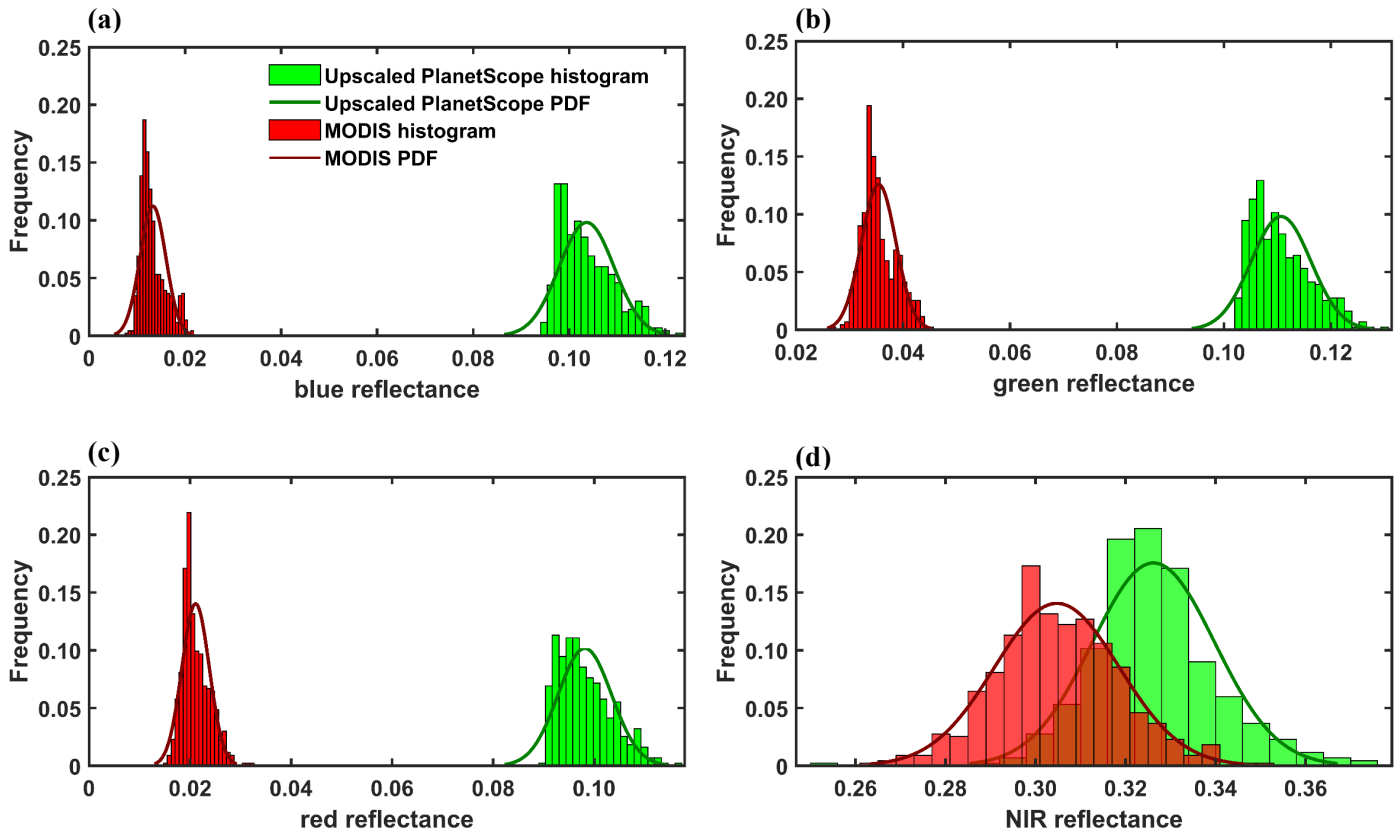


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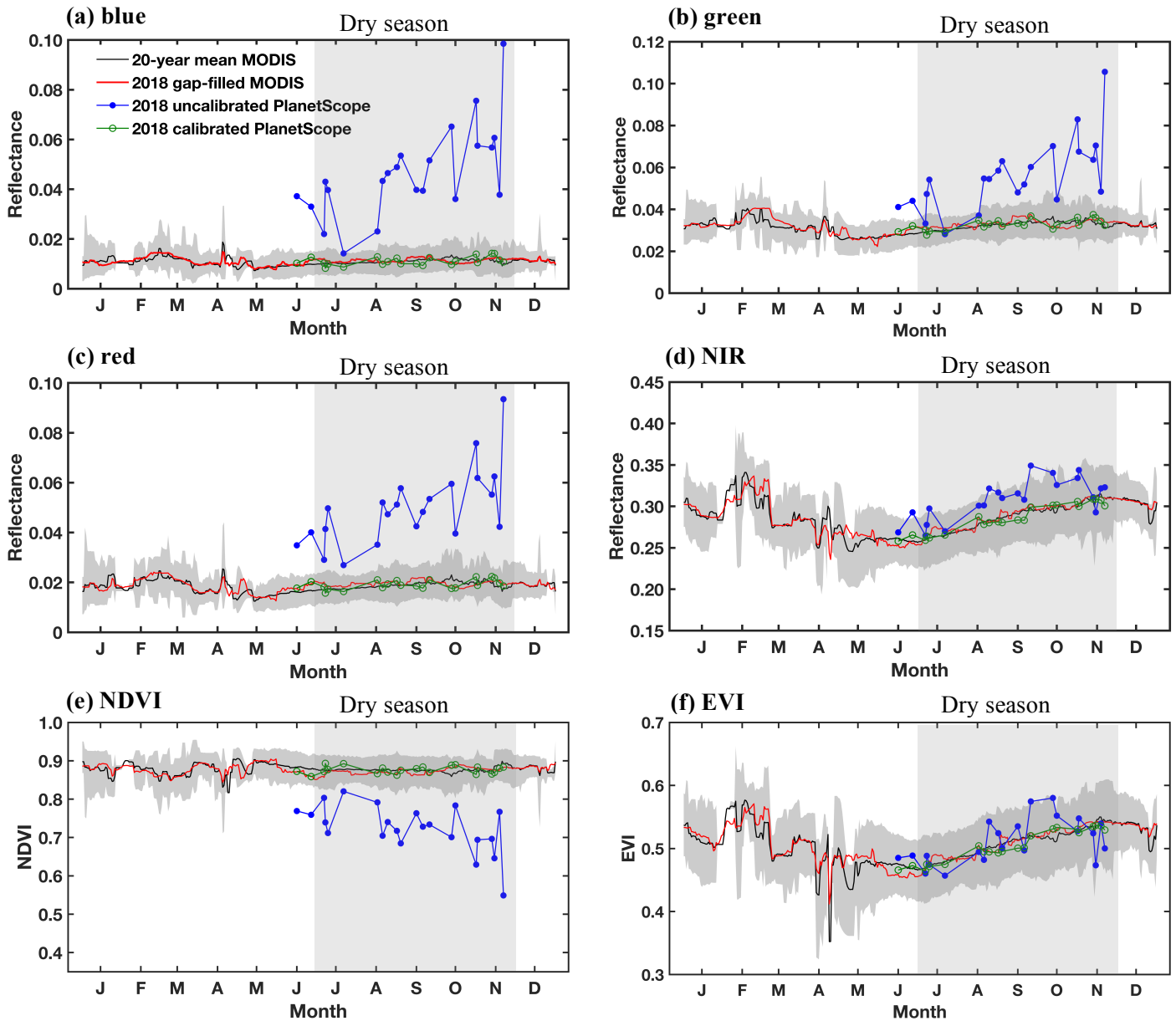


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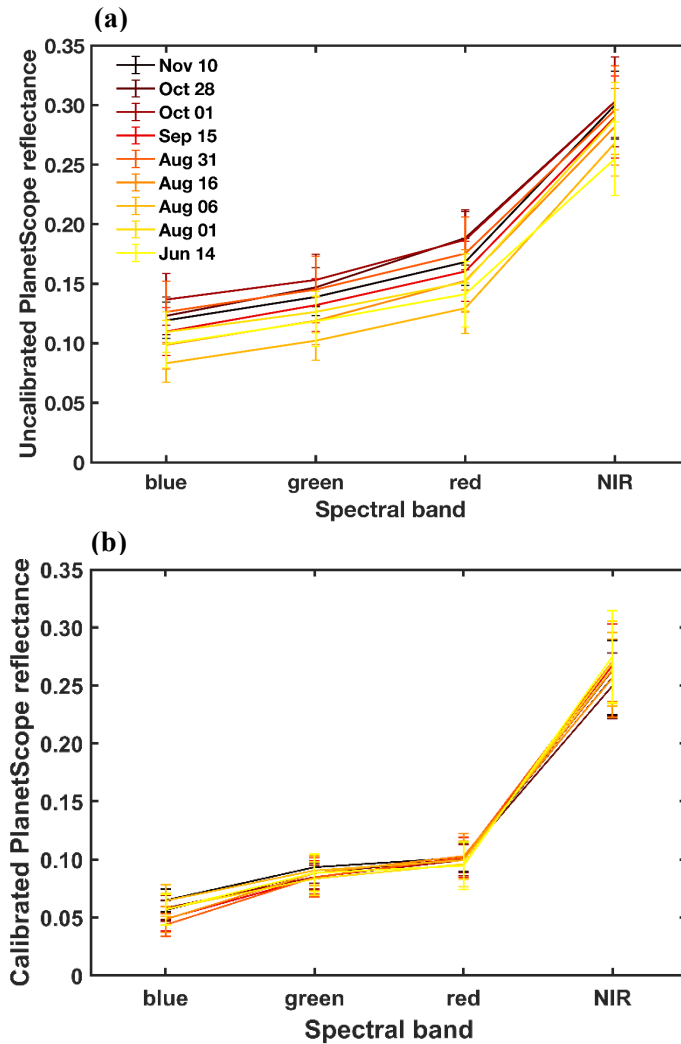


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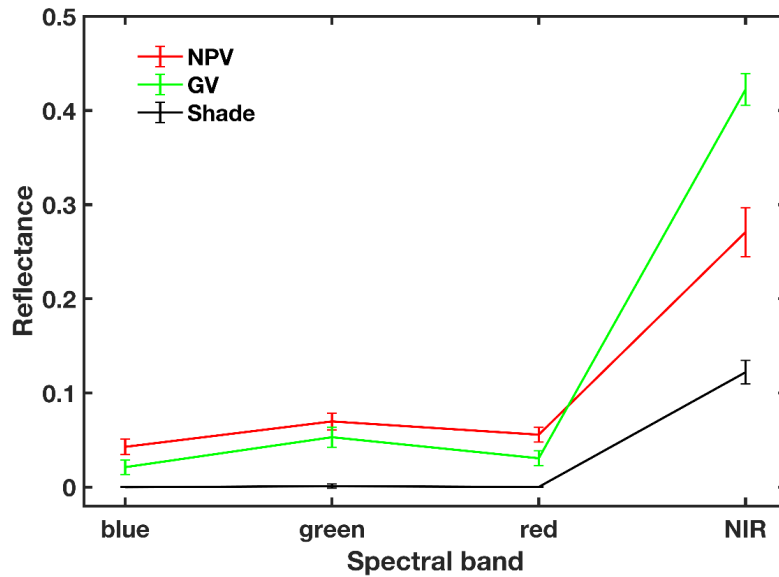


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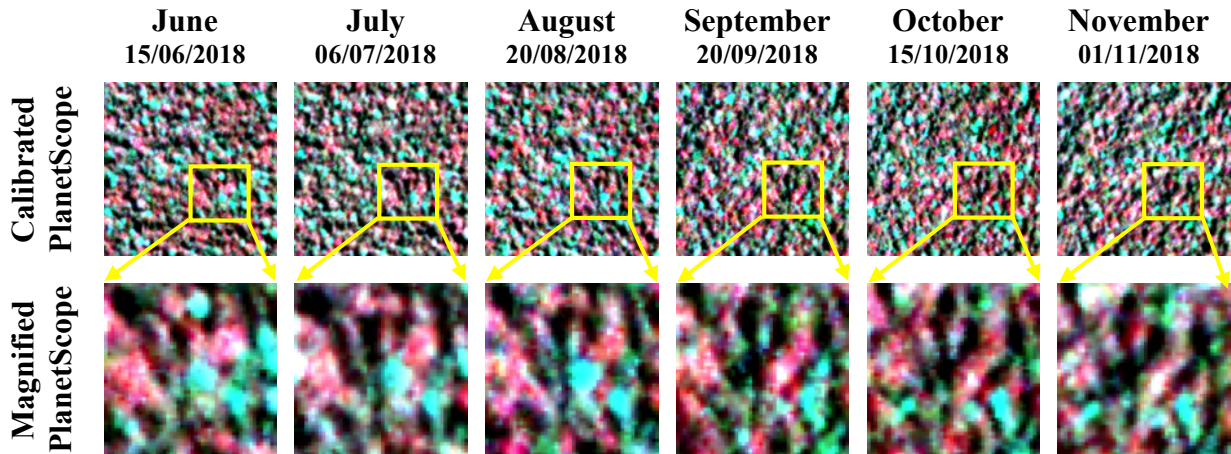


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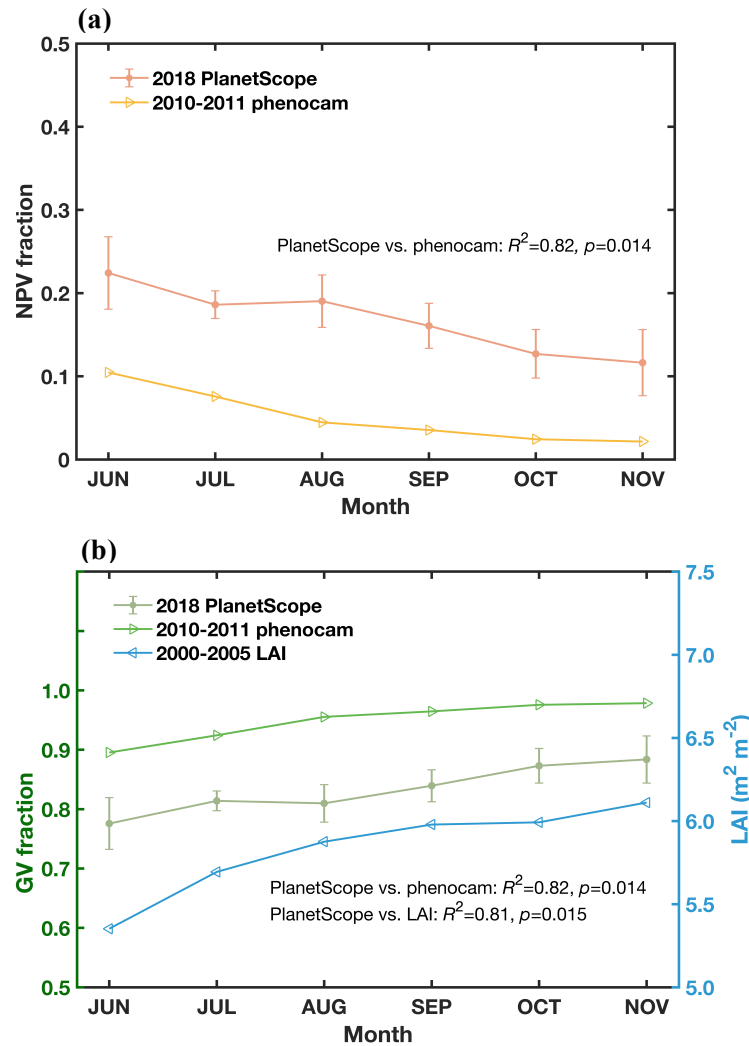


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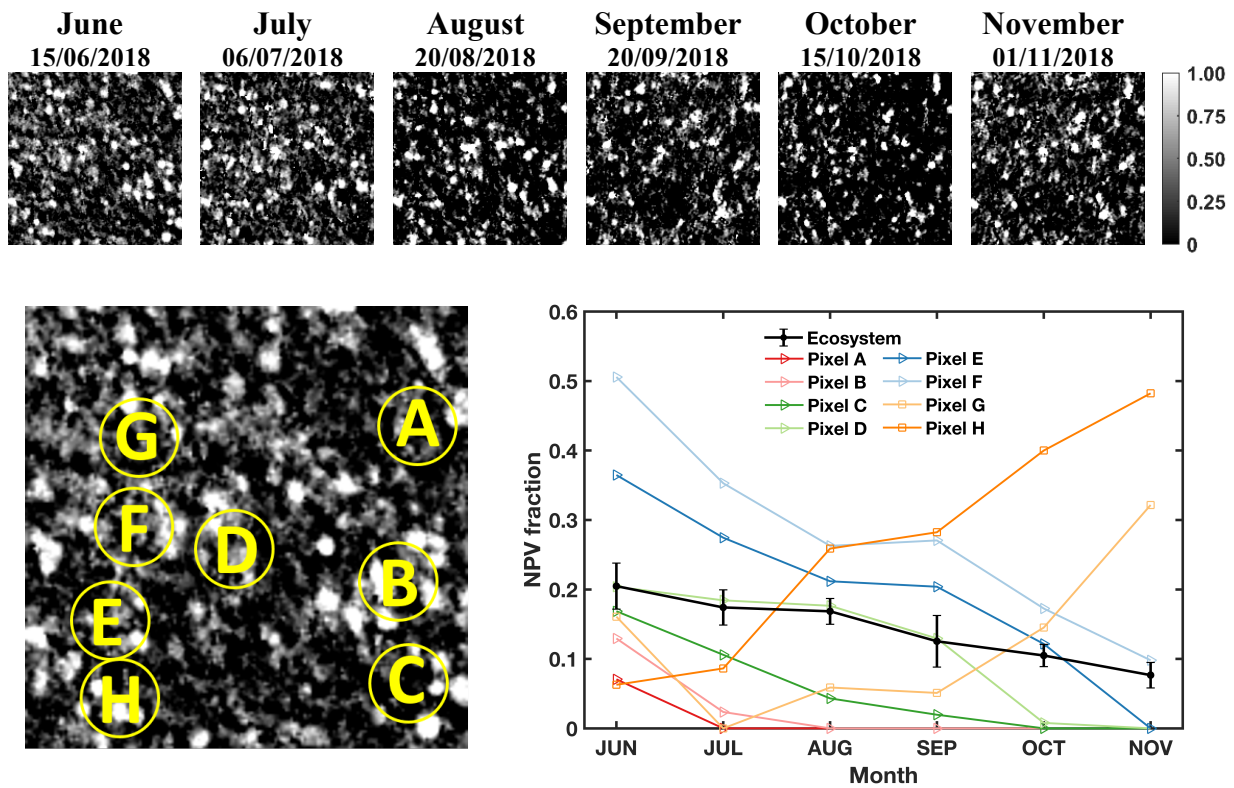


Figure 11. Assessing the dry-season change rate in NPV fraction derived from the calibrated PlanetScope data in 2018, including (a) a map of dry-season change rate in NPV fraction for an area of 500m×500m centered on the k67 tower site (the same as Fig. 10), and (b) statistical summary on frequency distribution of dry-season change rate for all pixels shown in panel (a). The dry-season change rate in NPV fraction is assessed through a linear regression analysis between dry-season change in NPV fraction and day of year in the dry season. The dry-season change rate in NPV fraction has a range from -0.005 to 0.005, with a negative value indicating dry-season decrease in NPV fraction (or green-up) and a positive value rate indicating dry-season increase in NPV fraction (or brown-down). In this area, there are 84.6% of pixels showing a dry-season green-up trend and 15.4% showing a brown-down trend.

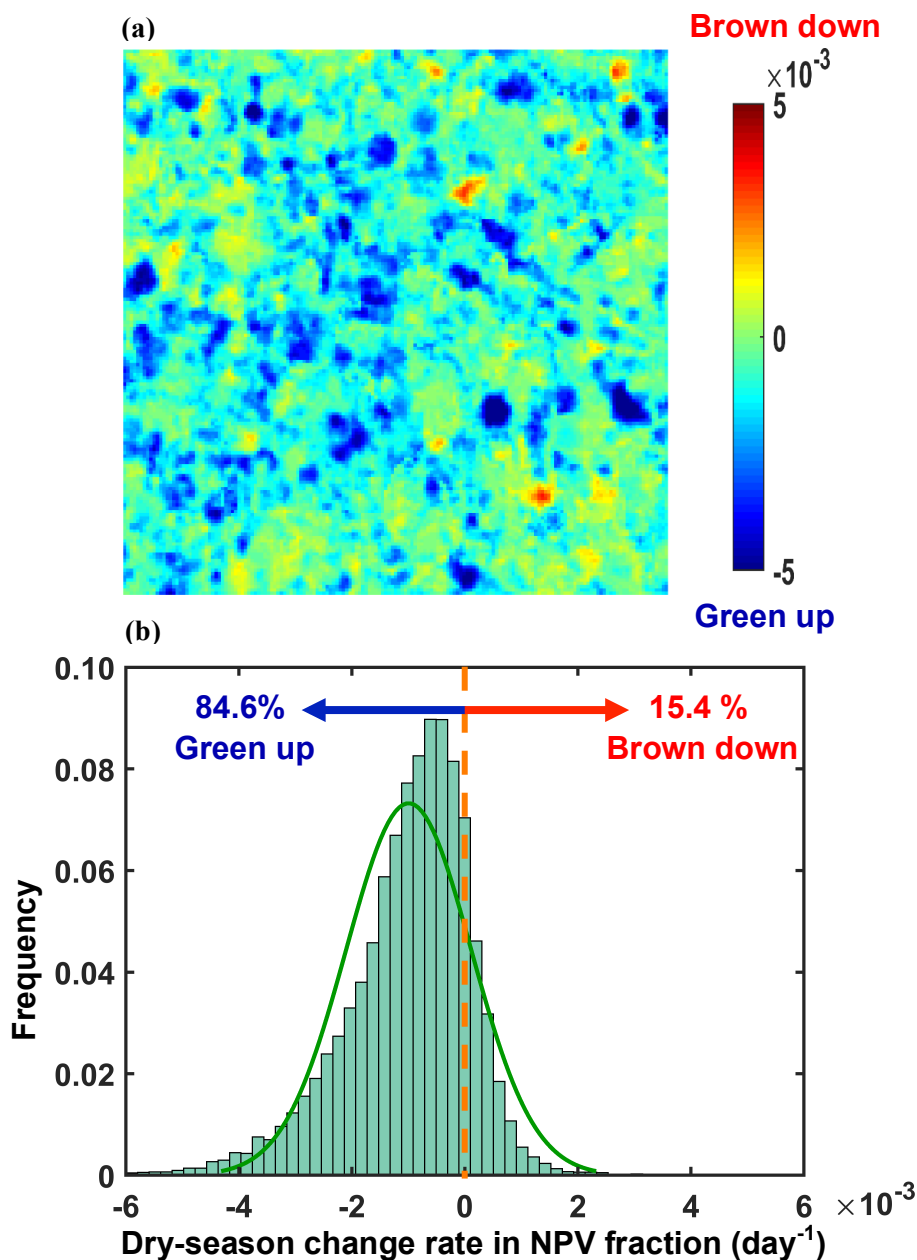


Figure S1. A true-colour RGB composited WorldView-2 image acquired at the k67 site (i.e. orange pentagram) on 28 July 2011, near-nadir view (2° off nadir). Three key endmembers, non-photosynthetic vegetation (NPV), green vegetation (GV) and shade, can be clearly identified in the image, with a few examples shown in red, light blue, and yellow, respectively. The figure is adopted from Fig. S1 in Wu et al. (2018), and for demonstration purposes, the WorldView-2 image of an area of about $1\text{km}\times 0.6\text{km}$ surrounding the k67 tower is displayed here as background.

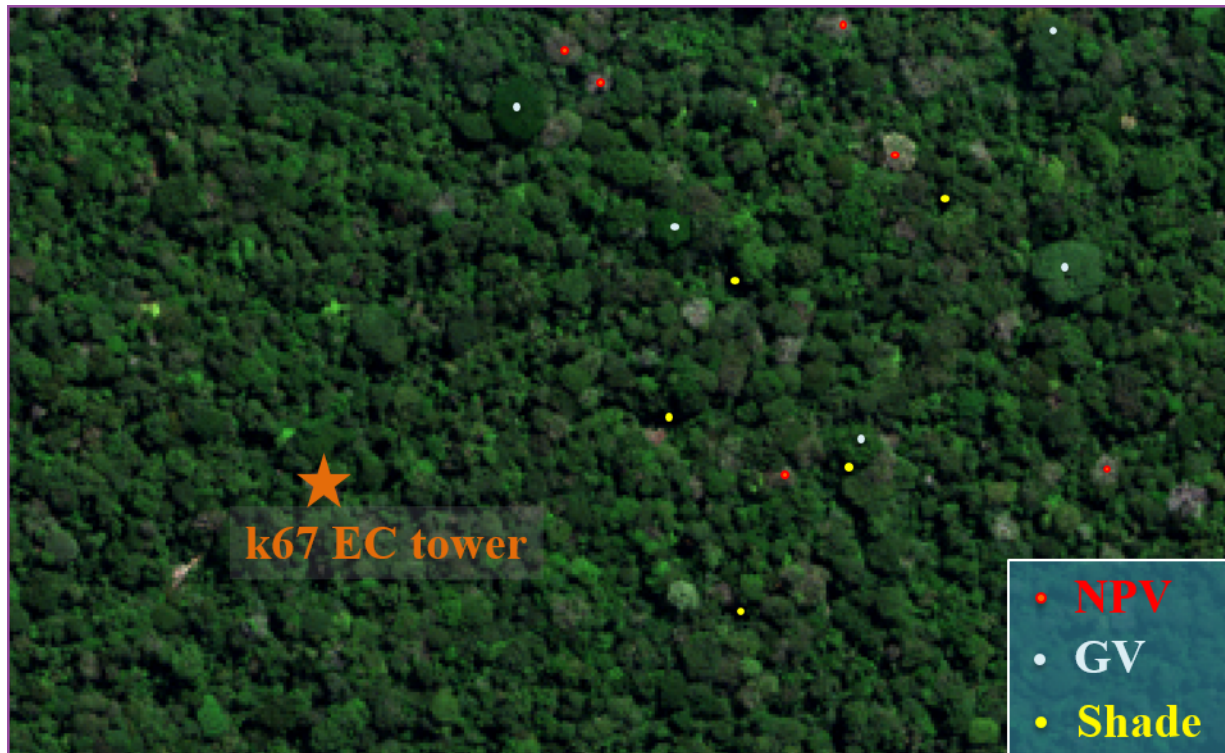


Figure S2. Color maps of the three gap-filling scenarios for each reflectance band (i.e. namely, blue, green, red and NIR) of BRDF-adjusted MODIS data at the k67 site in 2018 and 2019 and the Alter do Chão site in 2018. The color index for the following three scenarios are: for each band of a target MODIS pixel in the current year (2018/2019), it has (i) ≥ 200 valid daily measurements (blue color), (ii) ≥ 50 and <200 valid daily measurements (green color), and (iii) < 50 valid daily measurements (red color). The spatial extent of the k67 site is a $10\text{km}\times 10\text{km}$ area; the spatial extent of the Alter do Chão site is a $8\text{km}\times 8\text{km}$ area; the spatial resolution of MODIS data used is 500m. Also see more details in Section 3.2.2.

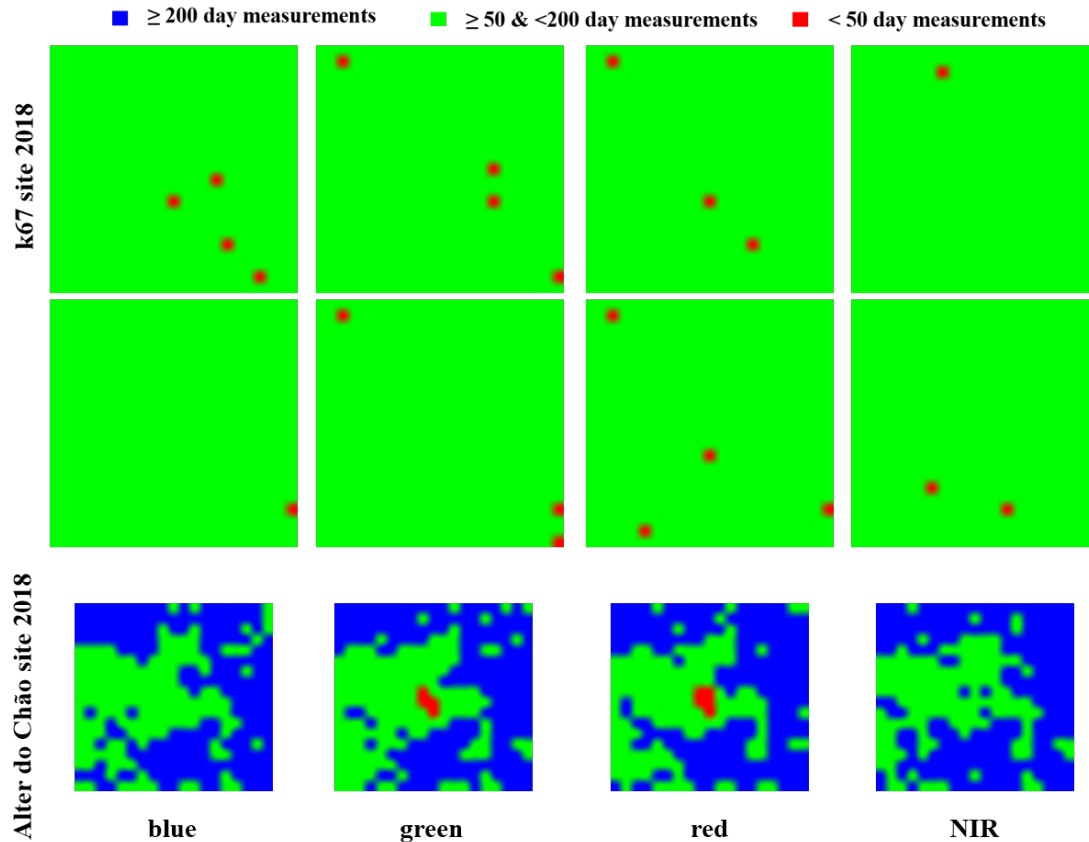


Figure S3. Example demonstration of band-specific outlier detection and gap-filling for BRDF-adjusted MODIS seasonality in 2019, including reflectance bands of (a) blue, (b) green, (c) red, and (d) NIR. The 20-year (i.e. 2000-2019) mean MODIS seasonality is displayed with the mean values in black lines, and grey shading for 95% confidence interval (i.e. 2.5 percentile for bottom and 97.5 percentile for top); the 2019 seasonality after quality control is displayed with valid data points (blue crosses, within the 95% confidence interval of 20-year mean seasonality) and outlier data points (blue circles, outside the 95% confidence interval); the gap-filling results in 2019 are displayed as red lines for daily means and red shadings for 95% confidence interval. A 5km×5km area centered on the k67 tower site is used here for demonstration purposes. Light grey shading indicates the dry season of the k67 site.

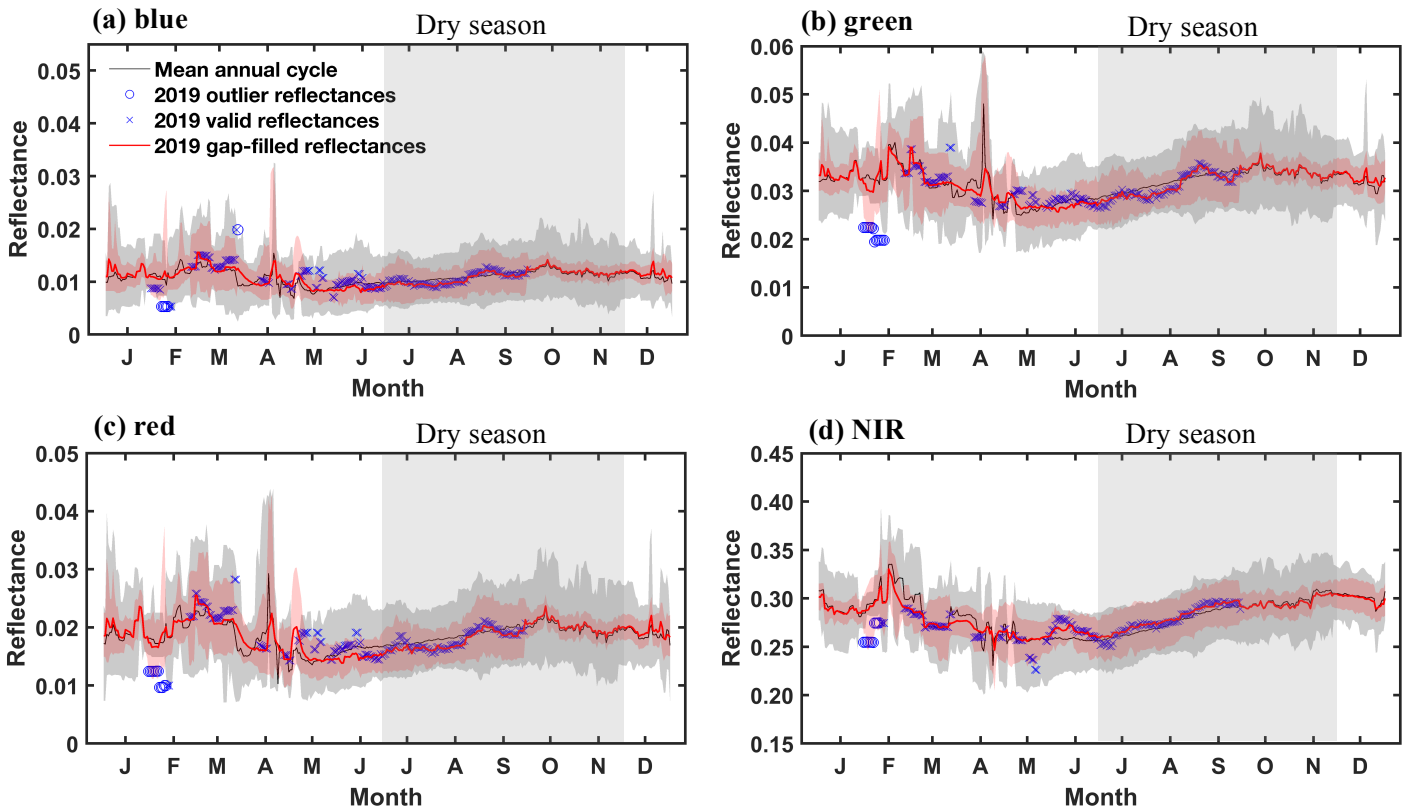


Figure S4. Evaluating the effect of gap-filling thresholds on the derived gap-filling results. Ecosystem-scale average daily MODIS reflectance of a 5km×5km area centered on the k67 tower site in 2018 is displayed here for demonstration, including reflectance bands of (a) blue, (b) green, (c) red, and (d) NIR. The original, non-gap-filled daily MODIS measurements are displayed in blue, with blue cross for valid data and blue circles for outlier data; color lines indicate five pairs of thresholds (i.e. 150 & 50; 180 & 50; 200 & 50; 200 & 80; 200 & 100) examined here, with each pair (two values) determining the cut-off of three different gap-filling scenarios. The results suggest that the seasonal trends of gap-filled MODIS in 2018 very well track that of original, non-gap-filled MODIS in 2018, and the gap-filling results are overall very stable, regardless of some variation in the thresholds used. Light grey shading indicates the dry season of the k67 site.

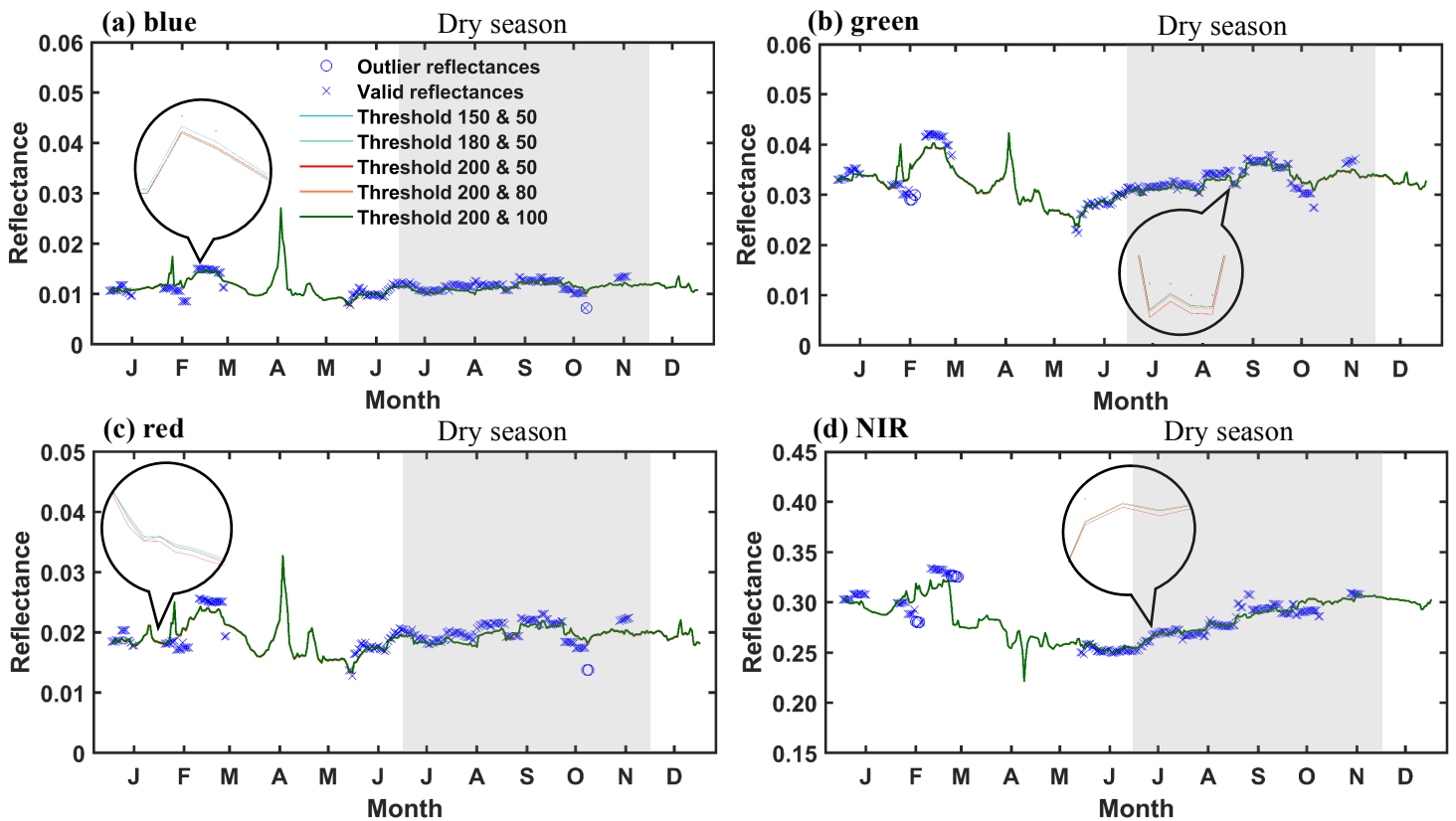


Figure S5. Comparisons of BRDF-adjusted MODIS time series in 2018 before (top panel) and after (bottom panel) gap-filling. The spatial extent shown here includes an area of 10km×10km centered on the k67 site; and the missing values (black color) represent those pixels that did not pass through the pixel quality control or suffered from the cloud/aerosol contaminations. For demonstration purposes, 6 daily MODIS images (composited by RGB=NIR-red-green bands) among a total of 183 images in the dry season of 2018 are selected, including one image per month from June to November (i.e. June 22, July 05, August 17, September 24, October 09, and November 14 from left to right).

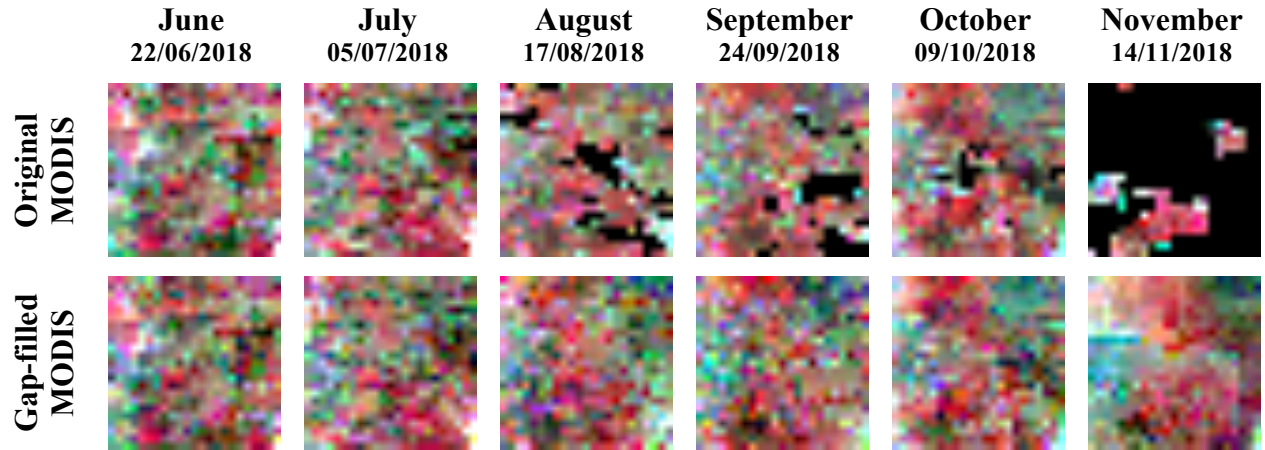
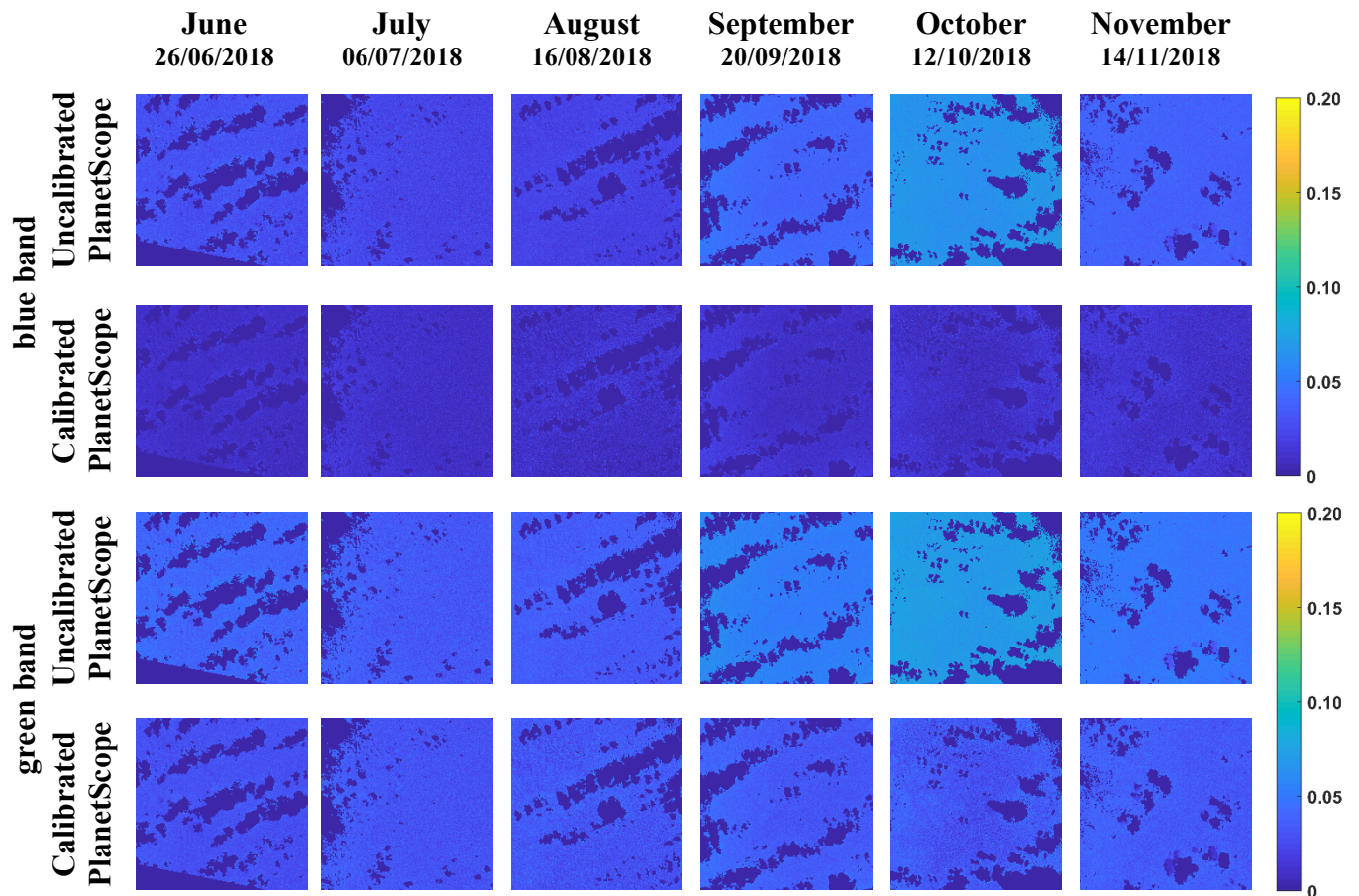


Figure S6. Comparisons of band-specific PlanetScope time series in 2018 before and after cross-calibration. Four reflectance bands (i.e. blue, green, red, and NIR) are shown here, with band-specific cross comparison indicating the difference in seasonal trend and magnitude of seasonal fluctuation between prior and post cross-calibration. The spatial extent shown here includes an area of 10km×10km centered on the k67 site; and the missing values represent those pixels that did not pass through the pixel quality control or suffered from the cloud/aerosol contaminations. For demonstration purposes, 6 dates of PlanetScope measurements among a total of 22 dates in the dry season of 2018 are selected, including one date per month from June to November (i.e. June 26, July 06, August 16, September 20, October 12, and November 14 from left to right). Color bars represent an increasing reflectance value from 0 to 0.2 (for blue, green and red bands) and 0 to 0.55 (for NIR band).



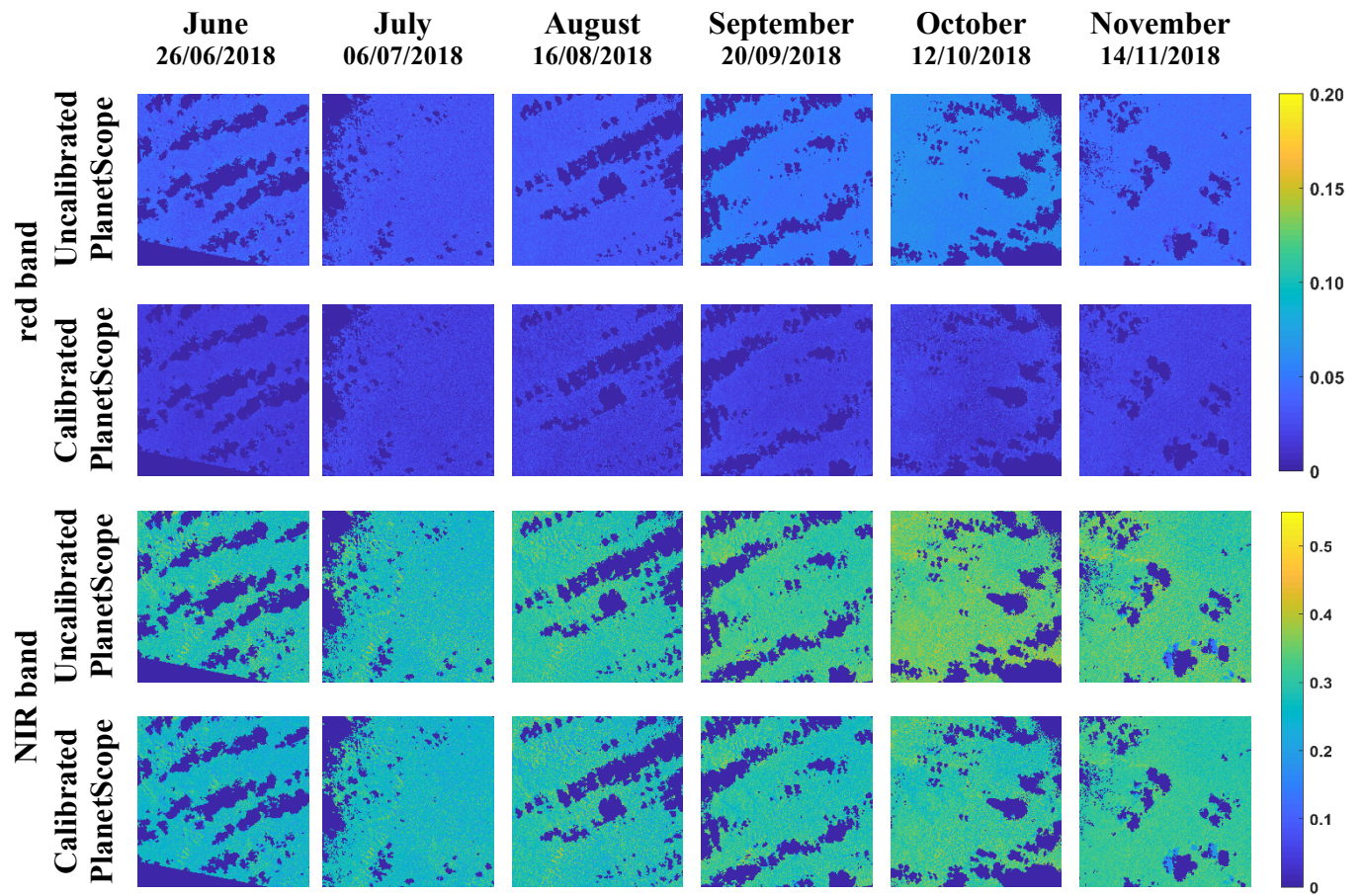


Figure S7. Example demonstration of endmember-specific reflectance spectra extraction, using the principle component analysis (PCA). A single PCA transformation is performed on all the calibrated PlanetScope images in 2018 and 2019 at the k67 site, and a randomly selected 1% of all image pixels of the two-dimensional feature space (i.e. the 2-D scatter plot of PC1 and PC2) is displayed in blue dots as background. The visually identified pixels of the three pure endmembers (i.e. red dots for NPV, green dots for GV, and black dots for Shade) are then overlaid on the 2-D scatter plot to aid the identification of these three key endmembers across all calibrated PlanetScope images. Further, we delimited the rectangle for each endmember, based on the mean and two standard deviations of each PCA axis derived from those manually identified pixels. Last, the average reflectance per band for all image pixels found within the three rectangles is calculated to derive endmember-specific reflectance spectra (see the results shown in Fig. 7).

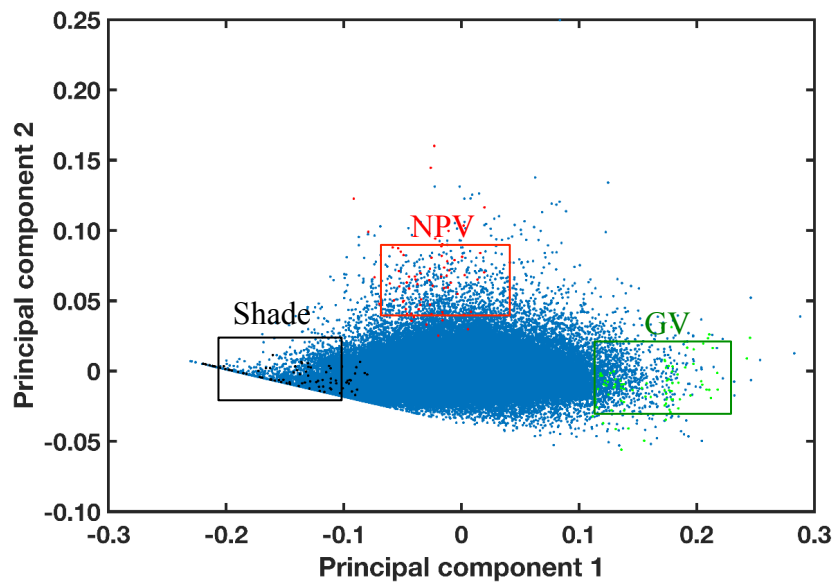


Figure S8. The mean (color lines) and one standard deviation (error bars) of reflectance spectra for three key endmembers derived from the visual assessment (i.e. averaging all the manually identified pure pixels extracted from calibrated PlanetScope images over the dry season of 2018 and 2019 at the k67 site). These three endmembers are pixels of completely leafless tree-crowns for pure non-photosynthetic vegetation (NPV), leafy tree-crowns for pure green vegetation (GV), and deep shade/shadow portions of the canopy (shade).

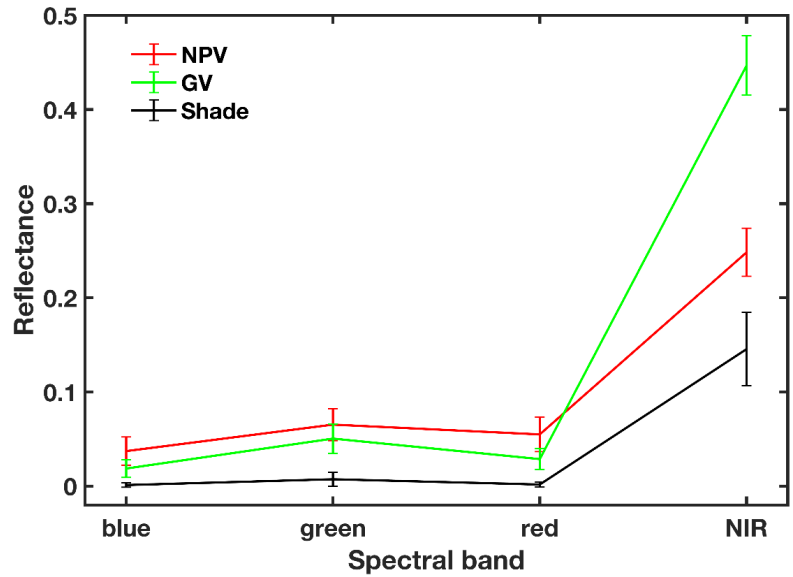


Figure S9. Evaluating the effects of temporal offset and spatial extent on the derived ecosystem-scale seasonal pattern, including the top panel for the temporal analysis and the bottom panel for the spatial analysis. In the top panel, (a) NIR reflectance and (b) EVI of BRDF-adjusted MODIS data is used, and the five time periods (i.e. 2019, 2018, 2010-2011, 2000-2005 and 2000-2019) in a 10km×10km centered on the k67 site are displayed. It is important to note that 2010-2011 and 2000-2005 present the same time periods as that of phenocam and LAI measurements, respectively. Black lines and error bars indicate the mean and one standard deviation of MODIS 20-years mean seasonality. In the bottom panel, (c) NIR reflectance and (d) EVI of the calibrated PlanetScope covering the full dry-season (from June to November) of 2018 is used, and seven spatial extents (i.e. 100m×100m, 200m×300m, 500m×500m, 1km×1km, 3km×3km, 5km×5km, and 10km×10km) centered on the k67 site are displayed. It is important to note that 100m×100m and 200m×300m present the same spatial extents as that of LAI and phenocam measurements, respectively. Black lines and error bars indicate the monthly mean and one standard deviation of the calibrated PlanetScope in the whole 10km×10km area. Light grey shading indicates the dry season of the k67 site.

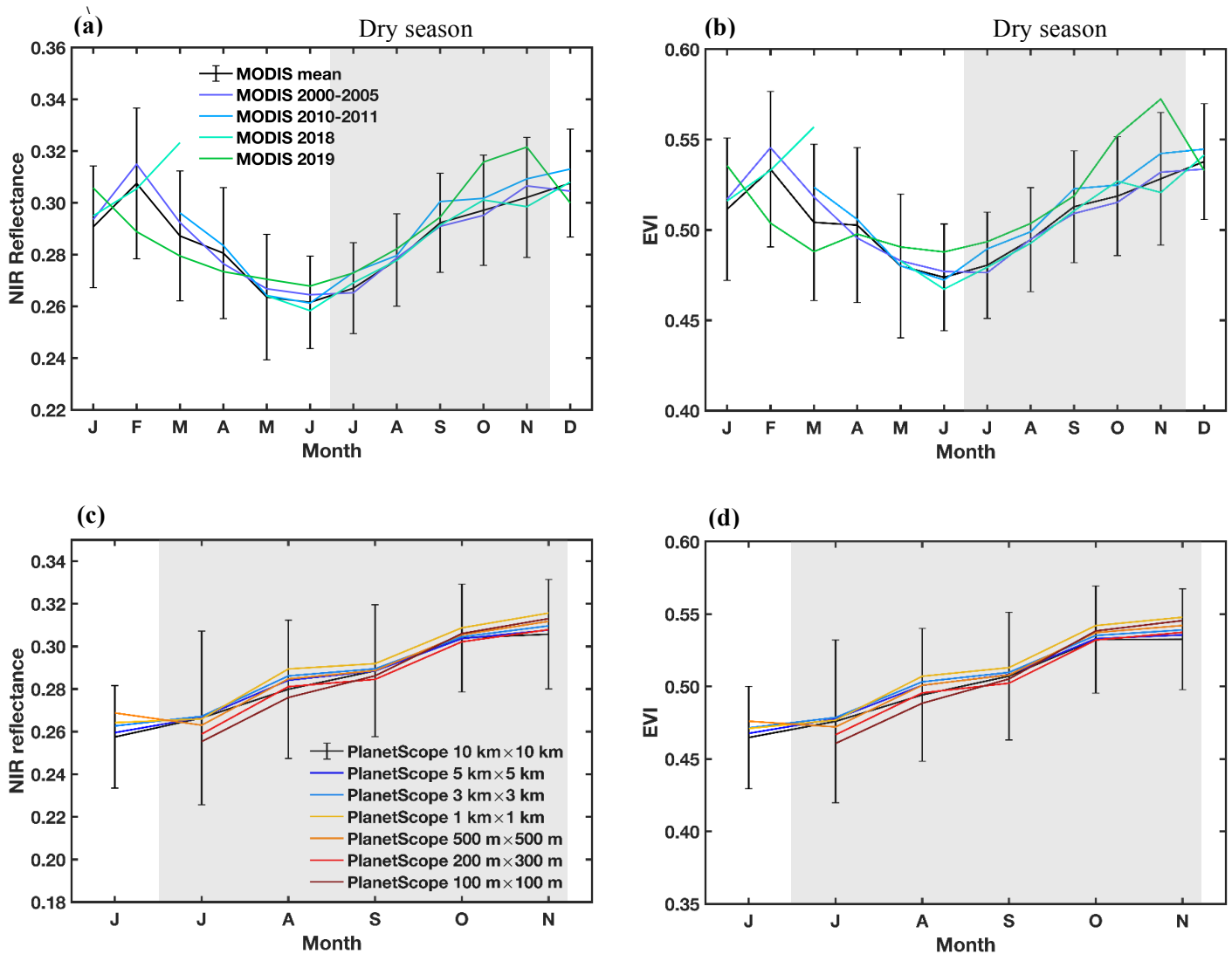


Figure S10. Ecosystem-scale seasonality of BRDF-adjusted MODIS (20-year mean in black and its 95% confidence interval in grey shading, and 2019 gap-filled in red) and 2019 PlanetScope (uncalibrated in blue and calibrated in green), including reflectance bands of (a) blue, (b) green, (c) red, and (d) NIR, and vegetation indices of (e) Normalized Difference Vegetation Index (NDVI), and (f) Enhanced Vegetation Index (EVI). BRDF-adjusted MODIS including both 20-year mean and 2019 gap-filled are displayed as background information; the uncalibrated/calibrated PlanetScope is based on the histogram matching analysis as shown in Fig. 4; and a 3km×3km area centered on the k67 tower site is used to calculate ecosystem-scale seasonality. Light grey shading indicates the dry season of the k67 site.

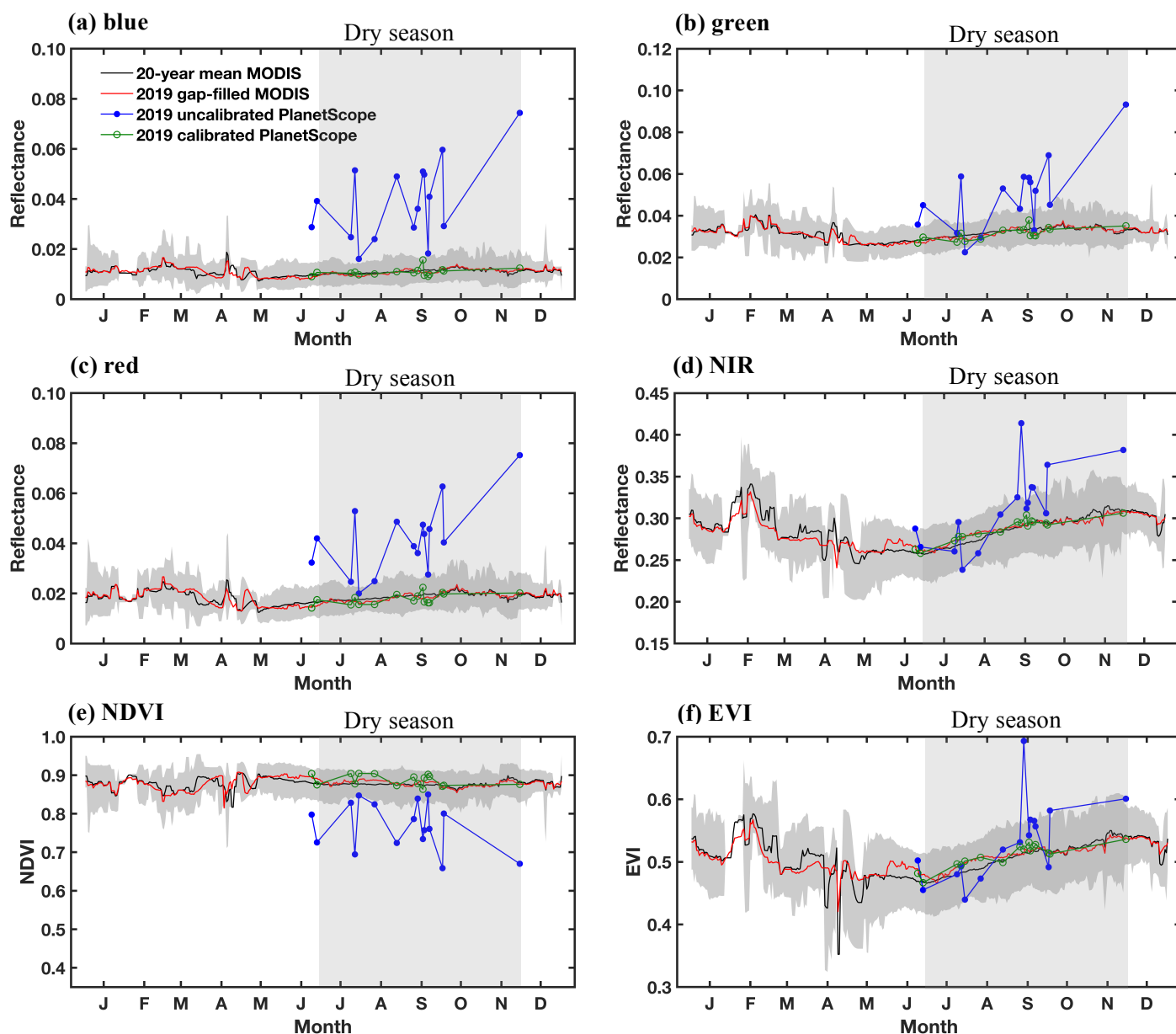


Figure S11. Comparisons of PlanetScope-derived phenology metrics and ground-based phenology measurements, including a) PlanetScope-derived and tower-phenocam measurements of NPV fraction, and b) PlanetScope-derived and tower-phenocam measurements of GV fraction and field LAI measurements. The calibrated PlanetScope data in 2019 are used here, and the PlanetScope-derived phenology metrics represent an average of a 3km×3km area centered on the k67 site; error bars indicate one standard deviation. Tower-phenocam measurements in 2010-2011 (conducted in an about 200m×300m area centered on the k67 site) and field LAI measurements in 2000-2005 (a 100m×100m plot, ~5 km apart from the k67 site) are based on the literature values (Brando et al., 2010; Wu et al., 2016; see methods for details).

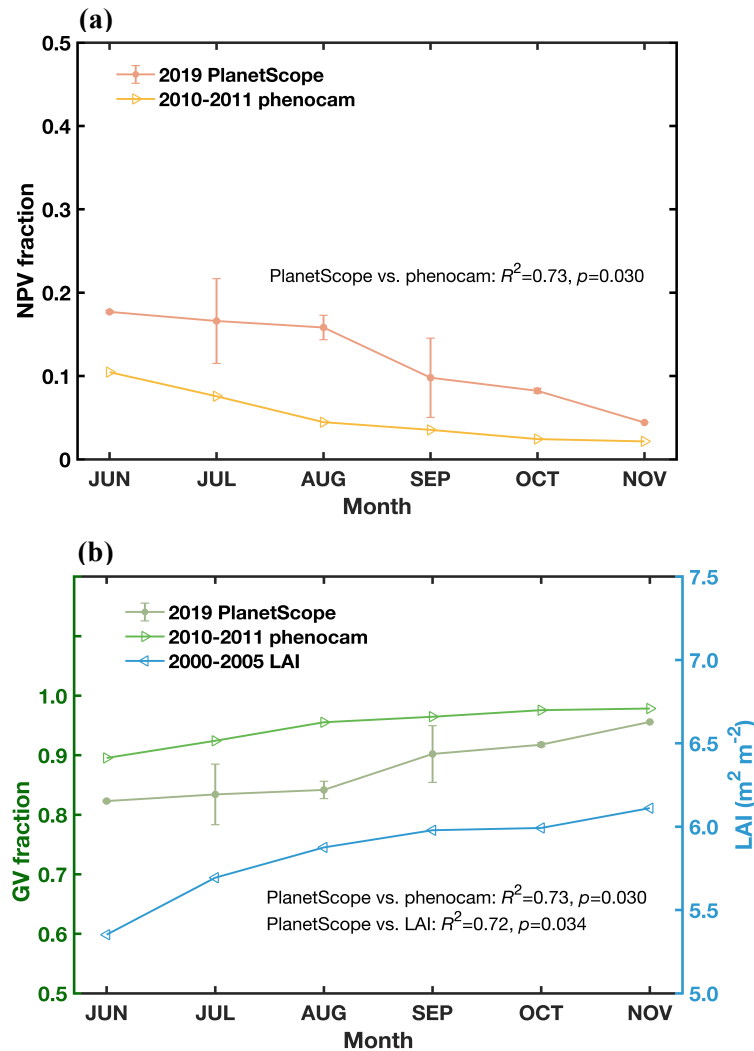


Figure S12. Dry-season trends in NPV fractions derived from two satellites (i.e. PlanetScope and MODIS), plotted against phenocam-derived leafless tree-crown fraction (here also called NPV fraction). PlanetScope of 2018 and BRDF-adjusted MODIS of 2018 and 2010-2011 are all from a 3km×3km area centered on the k67 tower site; NPV fractions derived from the two satellites use the same linear spectral unmixing (including the same endmember-specific reflectance spectra derived from calibrated PlanetScope); error bars indicate one standard deviation. Tower-phenocam-derived leafless tree-crown fraction is available only for 2010-2011 and is adopted from a previous study at the same site (Wu et al., 2016).

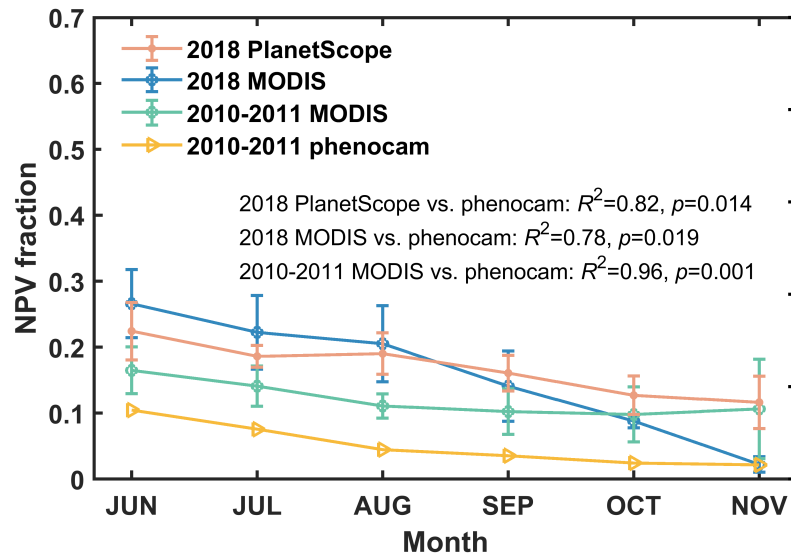


Figure S13. Assessing the dry-season change rate in NPV fraction in a 10km×10km area centered on the k67 tower site, including maps of dry-season change rate in NPV fraction in 2018 (a) and 2019 (b); and statistical summary on frequency distribution of dry-season change rate in NPV fraction in 2018 (c) and 2019 (d). All dates of calibrated PlanetScope in the dry season of 2018/2019 were used to assess the seasonal trend in NPV fraction based on a linear regression analysis. Color bar has a range from -0.005 to 0.005, with a negative value for dry-season decrease in NPV fraction (or green-up) and a positive value rate for dry-season increase in NPV (or brown-down) fraction. The missing values (grey color) indicate those pixels without enough (i.e. $n=4$) valid day measurements across the full dry-season after applying pixel quality control and cloud/aerosol contaminations removal. The missing values are pronounced in 2019, which might be related with the large-scale Amazon fire in 2019 (Lizundia-Loiola et al., 2020). Overall, our results suggest that the patterns shown in 2018 are comparable with that in 2019.

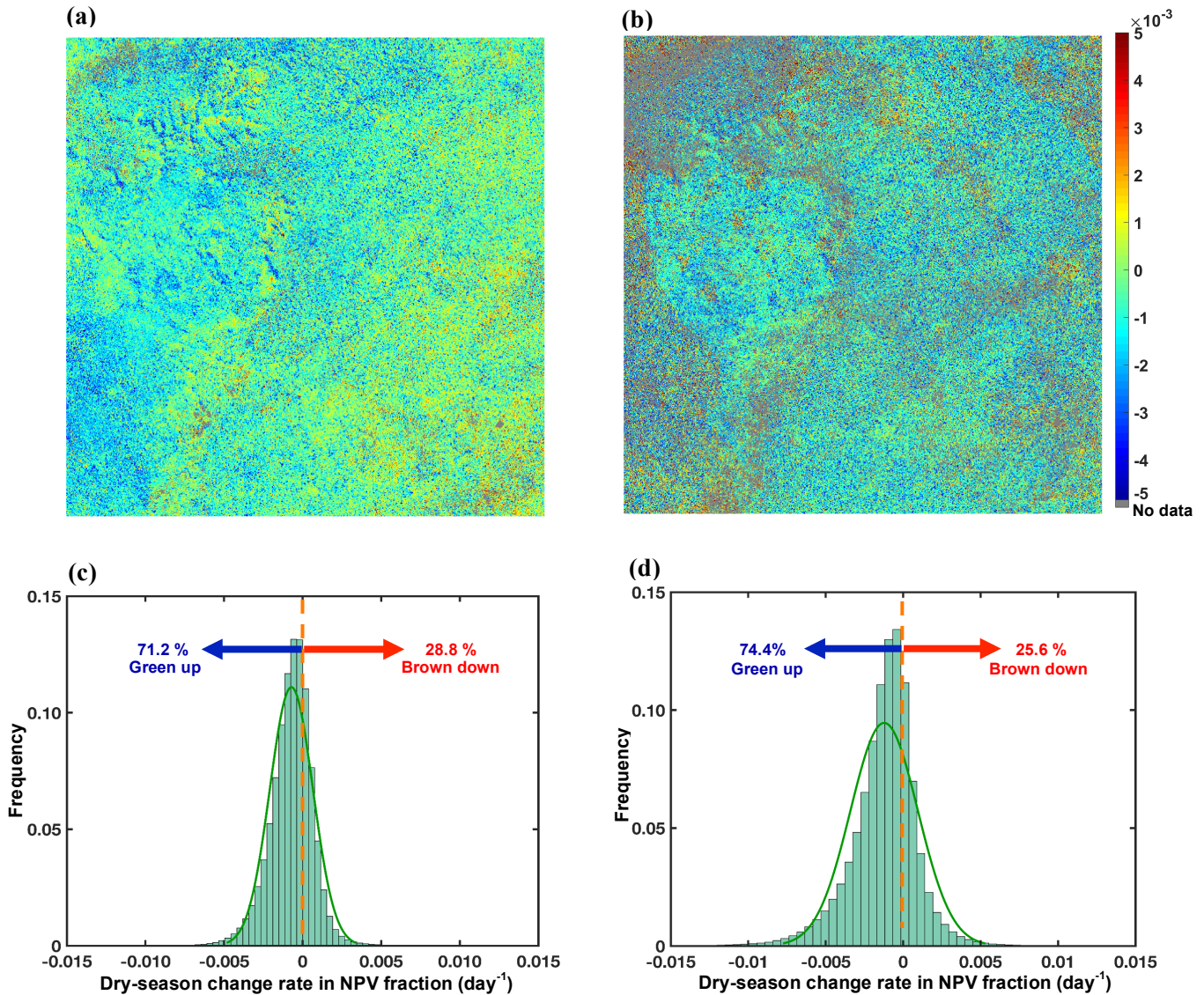


Table 1. Spatial resolutions, accessed data time ranges, and spectral bands and band-specific wavelength ranges of PlanetScope and MODIS data used at the k67 site.

Satellite	Spatial resolution (m)	Accessed data time range	Spectral band and wavelength range (nm)			
			Blue	Green	Red	NIR
PlanetScope	3	06/2018-11/2018 06/2019-11/2019	455-515	500-590	590-670	780-860
MODIS	500	02/2000-12/2019	459-479	545-565	620-670	841-876

References:

- Brando, P.M., Goetz, S.J., Baccini, A., Nepstad, D.C., Beck, P.S.A., Christman, M.C., 2010. Seasonal and interannual variability of climate and vegetation indices across the Amazon. *Proceedings of the National Academy of Sciences of the United States of America*, 107, 14685-14690.
- Lizundia-Loiola, J., Pettinari, M. L., & Chuvieco, E., 2020. Temporal Anomalies in Burned Area Trends: Satellite Estimations of the Amazonian 2019 Fire Crisis. *Remote Sensing*, 12, 151. <https://doi.org/10.3390/rs12010151>.
- Wu, J., Albert, L.P., Lopes, A.P., Restrepo-Coupe, N., Hayek, M., Wiedemann, K.T., Guan, K.Y., Stark, S.C., Christoffersen, B., Prohaska, N., Tavares, J.V., Marostica, S., Kobayashi, H., Ferreira, M.L., Campos, K.S., da Silva, R., Brando, P.M., Dye, D.G., Huxman, T.E., Huete, A.R., Nelson, B.W., Saleska, S.R., 2016. Leaf development and demography explain photosynthetic seasonality in Amazon evergreen forests. *Science*, 351, 972-976.
- Wu, J., Kobayashi, H., Stark, S.C., Meng, R., Guan, K.Y., Tran, N.N., Gao, S.C., Yang, W., Restrepo-Coupe, N., Miura, T., Oliviera, R.C., Rogers, A., Dye, D.G., Nelson, B.W., Serbin, S.P., Huete, A.R., Saleska, S.R., 2018. Biological processes dominate seasonality of remotely sensed canopy greenness in an Amazon evergreen forest. *New Phytologist*, 217, 1507-1520.