

1 **Title:** Cryptic phenology in plants: case studies, implications and recommendations

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3 **Running title:** Cryptic phenology in plants

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5 **Authors:**

6 Loren P. Albert^{1,2}, Natalia Restrepo-Coupe^{1,3}, Marielle N. Smith¹, Jin Wu^{4,5}, Cecilia Chavana-

7 Bryant^{6,7,8}, Neill Prohaska¹, Tyeen C. Taylor¹, Giordane A. Martins⁹, Philippe Ciais¹⁰, Jiafu

8 Mao¹¹, M. Altaf Arain¹², Wei Li¹⁰, Xiaoying Shi¹¹, Daniel M. Ricciuto¹¹, Travis E. Huxman¹³,

9 Sean M. McMahon¹⁴, Scott R. Saleska¹

10

11 **Affiliations:**

12 ¹Department of Ecology and Evolutionary Biology, University of Arizona, Tucson, AZ, 85721,

13 USA

14 ²Institute at Brown for Environment and Society, Brown University, Providence, RI 02912, USA

15 ³Plant Functional Biology and Climate Change Cluster, University of Technology Sydney,

16 Sydney, NSW, Australia

17 ⁴Biological, Environmental & Climate Sciences Department, Brookhaven National Lab, Upton,

18 New York, NY 11973, USA

19 ⁵School of Biological Sciences, The University of Hong Kong, Pokfulam, Hong Kong

20 ⁶Environmental Change Institute, School of Geography and the Environment, University of

21 Oxford, Oxford, UK

22 ⁷Climate & Ecosystem Sciences Division, Lawrence Berkeley National Laboratory, 1

23 Cyclotron Road, Berkeley, CA, 94720, USA

24 ⁸Department of Environmental Science, Policy and Management, UC Berkeley, Berkeley, CA,
25 94720, USA

26 ⁹Ciências de Florestas Tropicais, Instituto Nacional de Pesquisas da Amazônia (INPA), CEP
27 69.067-375, Manaus, AM, Brazil.

28 ¹⁰Laboratoire des Sciences du Climat et de l'Environnement, 91191 Gif sur Yvette, France

29 ¹¹Environmental Sciences Division and Climate Change Science Institute, Oak Ridge National
30 Laboratory, Oak Ridge, TN, 37831-6301, USA

31 ¹²School of Geography and Earth Sciences & McMaster Centre for Climate Change, McMaster
32 University, Hamilton, Ontario, L8S 4K1 Canada

33 ¹³Ecology and Evolutionary Biology & Center for Environmental Biology, University of
34 California, Irvine, CA 92697, USA

35 ¹⁴Smithsonian Institution's Forest Global Earth Observatory & Smithsonian Environmental
36 Research Center, Edgewater, MD, 21037, USA

37

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41 **Corresponding authors:**

42 Loren Albert

43 Address: Brown University; Institute at Brown for Environment and Society; 85 Waterman
44 Street; Providence, RI, 02912

45 Email: lalbert@email.arizona.edu

Phone: (832) 266-3051

Fax: N/A

46 Scott Saleska

47 Address: The University of Arizona; Department of Ecology & Evolutionary Biology; P.O. Box
48 210088; Tucson, AZ, 85721

49 Email: saleska@email.arizona.edu Phone: (520) 626-5838 Fax: N/A

50

51 **Statement of authorship:**

52 LA organized and wrote the paper with advice/input from SMM, TEH and SRS. NRC
53 contributed earth system model runs for the tropical forest case study at K67 and compiled
54 comparison data into figures. MNS reviewed, synthesized and illustrated published LAI and
55 litterfall data. For the temperate forest case study at Harvard Forest, SMM contributed DBH and
56 allometry, and PC, AA, JM, WL, DMR, and XS contributed model runs and feedback on the
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59

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61

62 **ABSTRACT**

63 Plant phenology – the timing of cyclic or recurrent biological events in plants – offers insight
64 into the ecology, evolution, and seasonality of plant-mediated ecosystem processes. Traditionally
65 studied phenologies are readily apparent, such as flowering events, germination timing, and
66 season-initiating budbreak. However, a broad range of phenologies that are fundamental to the
67 ecology and evolution of plants, and to global biogeochemical cycles and climate change
68 predictions, have been neglected because they are “cryptic” – that is, hidden from view (e.g root
69 production) or difficult to distinguish and interpret based on common measurements at typical
70 scales of examination (e.g leaf turnover in evergreen forests). We illustrate how capturing cryptic
71 phenology can advance scientific understanding with two case studies: wood phenology in a
72 deciduous forest of the northeastern USA and leaf phenology in tropical evergreen forests of
73 Amazonia. Drawing on these case studies and other literature, we argue that conceptualizing and
74 characterizing cryptic plant phenology is needed for understanding and accurate prediction at
75 many scales from organisms to ecosystems. We recommend avenues of empirical and modeling
76 research to accelerate discovery of cryptic phenological patterns, to understand their causes and
77 consequences, and to represent these processes in terrestrial biosphere models.

78

79 **1. INTRODUCTION**

80 All organisms have physical limits beyond which they function poorly or perish, and face trade-
81 offs in the allocation of finite resources to different structures and functions (Araújo et al., 2013;
82 Bennett & Lenski, 2007). Evolutionary strategies to establish, survive, grow, and reproduce are
83 shaped by such fundamental constraints and trade-offs (Roff & Fairbairn, 2007; Stearns, 1989).
84 When physical constraints or available resources vary regularly through time, organisms often
85 evolve temporal patterns in their activities to match or complement these variations (Diamond,
86 Frame, Martin, & Buckley, 2011). Temporal rhythms can also arise from time-dependent
87 biological process such as ontogeny and demography (Niinemets, García-Plazaola, & Tosens,
88 2012; Thomas & Winner, 2002). The Earth surface experiences seasonal cycles in temperature,
89 precipitation, and light that influence the availability of resources and the potential to carry out
90 the chemistry underlying biological processes (Schwartz, 2013; A. H. Strahler & Strahler, 2006).
91 Sessile organisms, such as most multicellular plants, are subjected to these seasonal cycles in-
92 place. Plant phenology—the timing of cyclic or recurrent biological events in plants—represents
93 functional strategies to persist within the bounds of natural climate seasonality and biological
94 possibility (Forrest & Miller-Rushing, 2010; Rathcke & Lacey, 1985). The study of phenology
95 has thus long been used as a means for gaining insight into the ecology and evolution of plants
96 and other organisms (Lieth, 1974).

97 The term ‘phenology’ traces to the Greek root *phaino*, meaning ‘to show,’ or ‘to appear’
98 (Schwartz, 2013), and early influential works on phenology promoted observations of
99 phenomena that were ‘sharp,’ ‘visible,’ and easy to detect (Leopold & Jones, 1947). In today’s
100 lexicon, common definitions of phenology broadly encompass the timing of cyclic or recurrent
101 biological events in plants, along with the causes and consequences of that timing (e.g. Lieth

1974, and (*phenology, n.* : *Oxford English Dictionary*, 2005). In contrast with broad contemporary definitions of phenology, studies of phenology often reflect the origin and history of the term by focusing on readily apparent biological events. These are generally aboveground and accompanied by changes that are readily and reproducibly distinguished with human senses such as visible changes in color, position, mass and volume. In plants, these include phenomena such as germination in annual plants, synchronized leaf production (leaf flush) and abscission in deciduous forests (Murali & Sukumar, 1993; Richardson & O’Keefe, 2009), and the onset of anthesis (flower opening) (Schwartz, 2013). Some phenological patterns, such as deciduous forest leaf onset, are also apparent at canopy and larger spatial scales with remote sensing tools ranging from phenocams to satellites (Badeck et al., 2004; Buitenwerf, Rose, & Higgins, 2015). Studying the timing and controls of such apparent biological events has contributed to understanding the evolution of plant traits and strategies in response to cycles in temperature, precipitation, photoperiod, and other physical variables (Chuine, 2010; Z. Huang, Liu, Bradford, Huxman, & Venable, 2016; Pau et al., 2011; van Schaik, Terborgh, & Wright, 1993). Phenological studies have also advanced our understanding of ecology, as many phenological patterns are coupled to biotic interactions such as intra-annual dynamics of predator or mutualist populations (Pau et al., 2011; Schwartz, 2013). More recently, some phenological events, such as date of anthesis or first leaf emergence, have proven useful indicators of biological responses to climate change (e.g. Parmesan & Yohe, 2003), and the relative ease of observing such events has enabled citizen science at regional and continental scales (Schwartz, Betancourt, & Weltzin, 2012).

However, many processes in plants are *not* readily apparent, but are no less cyclic or seasonal than the more easily observed phenomena that humans have historically monitored.

125 These phenologies are what might be called ‘cryptic.’ Some phenological patterns are difficult to
126 detect because they are hidden, including below-ground activities such as allocation to roots, and
127 internal processes such as allocation to carbohydrate reserves or cell differentiation. Other
128 phenological patterns are missed or misinterpreted based on common measurements at typical
129 scales of examination (e.g. changes in mass, area or volume of plant organs or of biomass pools).
130 Cryptic phenologies are not as well understood as apparent phenologies, and they have not been
131 used as indicators of climate change. Yet phenologies, including cryptic phenologies, play
132 critical roles in ecosystems, and mediate large-scale fluxes of carbon, nutrients, water, and
133 energy that are essential to consider as Earth’s climate changes (Abramoff & Finzi, 2015;
134 McCormack, Adams, Smithwick, & Eissenstat, 2014; Noormets, 2009; Richardson, Keenan, et
135 al., 2013b).

136 To address the disparity between the narrow scope of apparent phenology and the much
137 broader scope of cyclic and seasonal plant activities, we first offer a framework with terminology
138 that identifies the underlying challenges to observing, interpreting, and modeling cryptic
139 phenologies. Then, focusing on trees, we review specific case studies in which missing cryptic
140 phenology leads to problems for understanding and modeling seasonal ecosystem processes:
141 wood allocation in a temperate mixed forest, and leaf phenology in tropical evergreen Amazon
142 forests. We emphasize that attention to cryptic phenology is timely because many terrestrial
143 biosphere models (TBMs, the models used to represent vegetation of the land surface in Earth
144 system models and needed for climate change predictions; Fisher, Huntzinger, Schwalm, &
145 Sitch, 2014), assume that cryptic phenologies are strongly correlated with apparent phenologies,
146 and that such assumptions can lead to misattribution of the causes behind observed fluxes of
147 carbon, water, nutrients, and energy. Although we focus on trees, we argue that cryptic

148 phenologies are ubiquitous, and their conceptualization, characterization, and interpretation are
149 essential for accurate prediction at scales from organisms to ecosystems across the globe.

150 Table 1. Categorization of plant phenologies based on our current capacity for successful measurement, observation, and/or interpretation.

Plant phenology category	Specific examples	Frequently measured?	Model representation examples	Examples of current or possible use
<i>Apparent</i>				
Phases and events easily observed by humans.	<ul style="list-style-type: none"> • Bud burst¹ • Leaf abscission² • Anthesis³ • Fruit maturation⁴ 	Yes, and some records extend for decades or longer.	May be prescribed by relying directly on observations to force the model (e.g. remote sensing indices), or may be simulated based on environmental controls (e.g. growing degree days). ⁵	<ul style="list-style-type: none"> • Defining the duration of the growing season in seasonally dormant systems.^{6,7} • Testing capacity of hydrothermal models to predict events such as germination.^{8,9} • Using changes in timing of phenological events as indicators of climate change.^{10,11,12}
<i>Cryptic: hidden</i>				
Phases and events that are internal or obstructed by some barrier and thus difficult to detect.	<ul style="list-style-type: none"> • Below-ground processes such as root production.^{13,14} • Structural changes within cells or tissues such as xylem formation.¹⁵ • Remote sensing in cloudy regions such as wet tropical forests.¹⁶ 	No, but these blind spots are generally acknowledged.	Often assumed to be linked to or dependent on apparent phenology. ^{17,18} This assumption is generally explicit.	<ul style="list-style-type: none"> • Modeling of whole plant carbon and water dynamics.^{19,20} • Estimating intra-annual cycles of biomass gain.¹⁵ • Identifying temporal variation in below-ground interactions and associations.^{21,22}
<i>Cryptic: ambiguous</i>				
Phases and events that are missed or misinterpreted due to summed variables or compensatory processes in the same variable.	<ul style="list-style-type: none"> • Leaf quantity appears constant despite leaf turnover because new leaf production compensates for simultaneous old leaf abscission.²³ • Bole diameter can be affected by both wood formation and water status.²⁴ 	No, and these blind spots are not widely acknowledged. Measurements are needed at fine spatial or temporal scales, or with specialized tools, to capture and/or interpret the phenological pattern.	Often assumed to be represented by apparent phenology and/or assumed constant. These assumptions are generally implicit and often unrecognized.	<ul style="list-style-type: none"> • Decomposing measurements into components that reveal phenological strategies.^{25,26} • Attributing cycles of ecosystem flux to endogenous versus exogenous drivers.^{27,28} • Resolving lagged responses from instantaneous responses and their relationship to periods of stress.^{29,30}

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- ¹ Budburst. (2019). Budburst: An online database of plant observations, a citizen-science project of the Chicago Botanic Garden. Glencoe, Illinois. <https://budburst.org/plant-groups>
- ² (Escudero & Del Arco, 1987)
- ³ (Smith-Ramirez, Armesto, & Figueroa, 1998)
- ⁴ (Spellman & Mulder, 2016)
- ⁵ (Huntzinger et al., 2012)
- ⁶ (Churkina, Schimel, Braswell, & Xiao, 2005)}
- ⁷ (Schwartz, 2013)
- ⁸ (Bauer, Meyer, & Allen, 1998)
- ⁹ (Hardegree, 2006)
- ¹⁰ (Badeck et al., 2004)
- ¹¹ (Schwartz, AHAS, & AASA, 2006)
- ¹² (Cleland, Chuine, Menzel, Mooney, & Schwartz, 2007)
- ¹³ (Steinaker & Wilson, 2008)
- ¹⁴ (Radville, McCormack, Post, & Eissenstat, 2016)
- ¹⁵ (Cuny et al., 2015)
- ¹⁶ (Asner, 2001)
- ¹⁷ (Delpierre, Berveiller, Granda, & Dufrene, 2015)
- ¹⁸ (Abramoff & Finzi)
- ¹⁹ (Hu, Moore, Riveros-Iregui, Burns, & Monson, 2010)
- ²⁰ (Michelot et al., 2012)
- ²¹ (Mullen & Schmidt, 1993)
- ²² (S. W. Simard et al., 2012)
- ²³ (Albert et al., 2018)
- ²⁴ (Chitra-Tarak et al., 2015)
- ²⁵ (Tang & Dubayah, 2017)
- ²⁶ (Smith et al. 2019)
- ²⁷ (Wu *et al.*, 2016)
- ²⁸ (Migliavacca *et al.*, 2015)
- ²⁹ (Ogle *et al.*, 2015)
- ³⁰ (Guo & Ogle, 2018)

152 2. THE CHALLENGE OF CRYPTIC PHENOLOGY: A FRAMEWORK

153 As discussed above, we describe *apparent* phenologies as those that were selected for clear
154 observation by humans, often with minimal technological support. By contrast, *cryptic*
155 phenologies require extensive investigation or validation to capture, and as a consequence have
156 rarely been measured at the temporal or spatial scale necessary to document and understand
157 (Table 1). ‘Cryptic’ is a useful term because it implies concealment and ambiguity—two general
158 challenges to capturing and understanding the full scope of cyclic/recurrent biological events in
159 plants. To highlight these challenges, here we frame cryptic phenology as ‘hidden’ or
160 ‘ambiguous.’

161 Plant phenological patterns are *hidden* when some physical or technological barrier
162 obstructs observation (Table 1). Soil conceals below-ground processes such as cycles of root
163 production and turn-over (Abramoff & Finzi, 2015; Delpierre et al., 2016). Internal plant
164 structures are (by definition) hidden behind layers of cells, making the timing of recurrent
165 processes such as secondary xylem (wood) formation difficult to observe *in vivo* (Chaffey, 1999;
166 Plomion, Leprovost, & Stokes, 2001). Large-scale phenological processes can also be hidden, as
167 cloud cover can consistently obstruct satellite observations of vegetation reflectance over humid
168 regions such as tropical forests (Asner, 2001). In dense forests, the upper canopy leaves partly
169 obstruct remote sensing observations of mid- and understory leaf area patterns (Tang &
170 Dubayah, 2017) and vice versa for ground-based observations, (Smith et al., 2019). When
171 phenological processes are hidden, describing them often requires time-consuming methods,
172 such as minirhizotrons or soil cores (for roots; Abramoff & Finzi, 2015; Gaudinski et al., 2010),
173 fixation of tissue samples from multiple time periods (for wood formation; Arend & Fromm,
174 2007), or ‘ground truth’ observations (for remotely-sensed vegetation greenness indices;

175 Chavana-Bryant et al., 2017; Lopes et al., 2016; Richardson et al., 2018; Wu et al., 2017).
176 Hidden phenological patterns are challenging and/or time-consuming to measure, but the
177 scientific community frequently acknowledges the scarcity of these measurements, and models
178 including hidden phenology explicitly define their representation within allocation schemes (e.g.
179 Abramoff & Finzi, 2015).

180 Phenological patterns in plants are *ambiguous* if phases and events are missed or
181 misinterpreted due to summed variables or compensatory processes in the same variable (Table
182 1). A measured variable (e.g. plant mass, canopy leaf area, or bole volume) may be a function of
183 multiple variables that are not synchronized with each other (Fig. 1a), making it difficult to
184 interpret the temporal changes in the measured variable. For example, determining whether
185 changes in tree stem diameter are caused by long-term carbon gain (such as xylem wall
186 thickening), or reversible changes in plant water status (such as stem expansion or shrinkage), is
187 difficult solely on the basis of stem diameter measurements (Chitra-Tarak et al., 2015; Cuny et
188 al., 2015; Sheil, 1997). Additional examples of ambiguous phenological patterns arise when, for
189 a given system, there are inputs and outputs of the same variable that are compensatory,
190 maintaining the appearance of constancy despite change. For example, compensatory leaf
191 production and abscission could maintain a constant total quantity of leaves in the canopy,
192 belying underlying cycles in leaf production and abscission (Albert et al., 2018; Doughty &
193 Goulden, 2008; Wu et al., 2016). In this example, at least two of the three terms (inputs, outputs,
194 and total) need to be sufficiently constrained by measurements to determine whether the steady
195 state of the total is achieved due to *constant* inputs and outputs (Fig. 1b) versus cyclic, but
196 *compensatory* inputs and outputs (Fig. 1c). Whatever the scale of study, measuring multiple

197 terms over time requires more effort and/or instrumentation, and this difficulty contributes to the
198 challenge of recognizing and resolving ambiguous phenology.

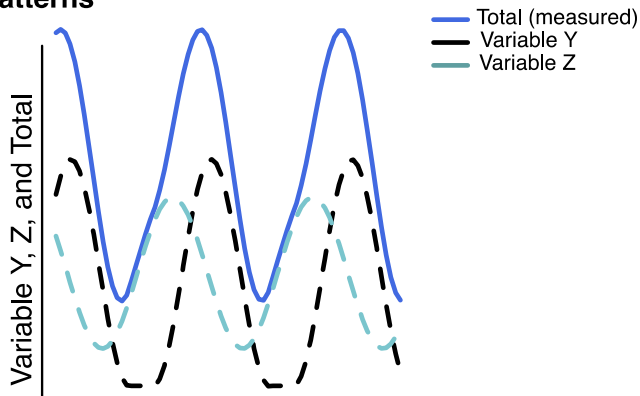
199 Whereas hidden phenological patterns are often acknowledged to exist, but rarely
200 measured, ambiguous phenological patterns are not frequently acknowledged because
201 measurements *are* being made—the challenge lies in interpreting those measurements. For
202 example, we understand that root phenology is hidden, and difficult to measure, because roots
203 are underground. By contrast, we may not even realize that leaf production and loss show
204 seasonal rhythms if the quantity of leaves in a canopy is largely constant (a compensatory
205 scenario). The distinction between hidden and ambiguous categories is not absolute because
206 phenology could be both hidden and ambiguous. For example, the mechanism of biomass gain
207 (xylogenesis) is hidden within stems, and stem diameter represents an integration of cells at
208 different stages in the sequence of xylogenesis: cell expansion, secondary cell wall thickening,
209 lignification and dead cells (Cuny et al., 2015; Plomion et al., 2001). Thus changes in stem
210 diameter emerge from expansion as well as biomass gain (Cuny et al., 2015), resulting in some
211 ambiguity.

212 Cryptic phenologies do not follow fundamentally different rules than their more apparent
213 counterparts. Plant phenologies, in general, are consequences of biology, climatic seasonality,
214 and their interactions. Yet a focus on cryptic phenology challenges us to explicitly consider our
215 current observational blind spots. These blind spots may prevent us from gaining a
216 comprehensive understanding of organismal strategies and limitations in relation to their biology
217 and physical environment, with consequences for our understanding of population, community,
218 and ecosystem ecology. Ultimately, our ability to document, understand, and model the
219 component processes that contribute to large-scale biosphere/atmosphere exchange of CO₂ and

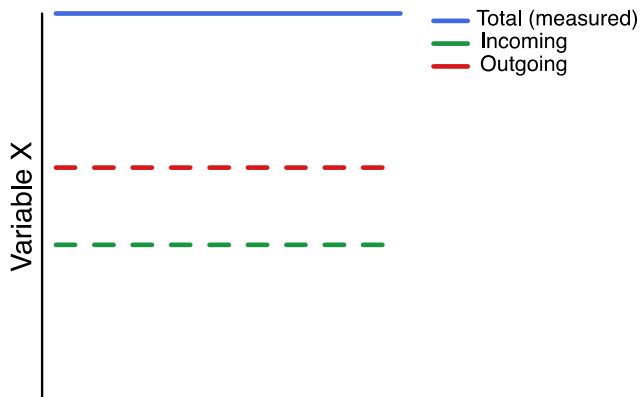
220 water vapor, impacts our ability to predict responses of natural systems to global change (Getz et
221 al., 2017; Noormets, 2009; Richardson et al., 2012).

222

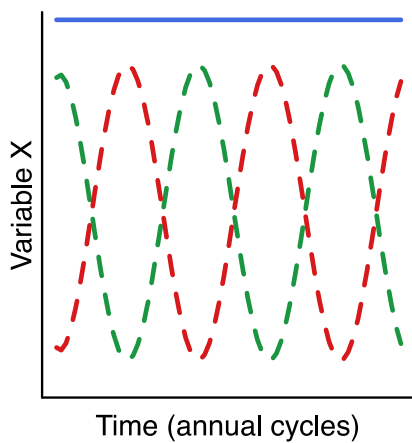
(a) Example 1: Measured variable is the sum of multiple other variables with different temporal patterns



(b) Example 2: Measured variable is the sum of constant incoming and outgoing pools



(c) Example 3: Measured variable is the sum of compensatory incoming and outgoing pools



224 Figure 1: Examples of ambiguous phenological patterns. In example 1, multiple variables (Y and
225 Z) contribute to some total that is measured. Variables Y and Z may have different peak timing,
226 different rates of change, and/or different amplitudes, that become summed for the measured total.
227 Thus the phenological patterns of individual variables Y and Z are ambiguous. In examples 2 and
228 3, the total for some biological variable X is the sum of an incoming (new) pool and outgoing (old)
229 pool. In example 2, the total, the incoming, and the outgoing pools are constant. In example 3, the
230 total is also constant, but the incoming and outgoing pools are dynamic, with inputs compensating
231 for losses. Distinguishing between the scenarios represented by examples 2 and 3 is difficult based
232 solely on measurements of the variable X total, and so phenological patterns of the incoming and
233 outgoing pools remain ambiguous.

234

235 **3. CASE STUDIES IN CRYPTIC PHENOLOGY**

236 In the two case studies below, we draw upon available studies, data, and models to examine the
237 evidence for, and implications of, cryptic phenology in two different plant processes in distinct
238 ecosystems: gross primary productivity in tropical evergreen forests, and allocation to wood in
239 temperate deciduous forests. For each case study we compare observations with simulations
240 from terrestrial biosphere models (TBMs; models that represent land surface vegetation in the
241 Earth system models used to simulate current and future global energy, carbon and water budgets
242 (Fisher et al., 2014; Le Quéré et al., 2015). These model-observation comparisons serve two
243 purposes. First, comparisons of TBMs with observations offer a test of our current ability to
244 reproduce the seasonality of biosphere-atmosphere mass exchanges and represent phenological
245 processes (Richardson et al., 2012) with implications for improving models (Richardson,
246 Keenan, et al., 2013b). Second, the model-observation comparisons, placed in the context of
247 current literature examining multiple scales and using multiple tools, allows us to ask whether
248 cryptic phenology presents obstacles to our ability to test hypotheses about the drivers,
249 consequences, and even the presence of phenology. Together, these case studies represent
250 different plant organs and ecosystems, demonstrating how capturing cryptic phenological

251 processes can be necessary for correct attribution of cause and effect—and ultimately modeling
252 ecosystem processes—in many systems.

253

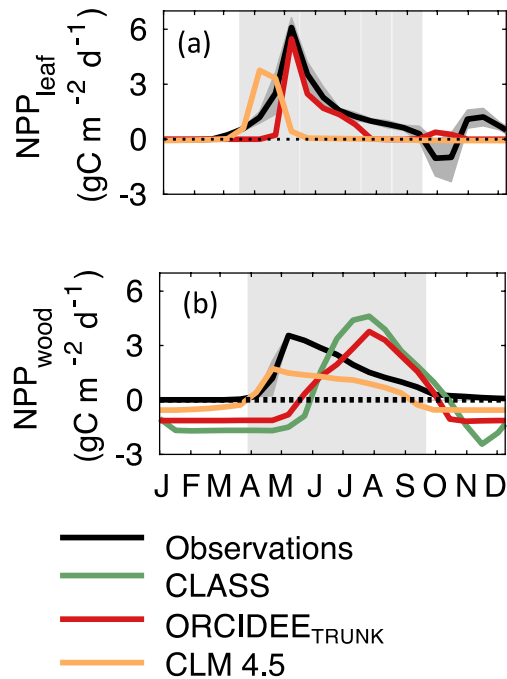
254 **3.1 Cryptic phenology of bole growth in temperate forests: implications for the timing of** 255 **carbon allocation to wood**

256 A TBM model-data comparison of bole growth at Harvard Forest, a temperate mixed forest site,
257 reveals the challenge of estimating and modeling the hidden phenology of wood allocation.

258 Allocation to wood in trees is necessary for estimations of forest carbon stocks (Kellner and
259 Clark 2012), and important for characterizing fast versus slow growth strategies across species
260 (Chave et al., 2009; Reich, 2014). The intra-annual timing of wood allocation may show how
261 carbon gain responds to seasonal climate, and reveal periods of vulnerability or resilience to
262 stress (Babst et al., 2014; Battipaglia et al., 2010). The process of woody biomass gain (from
263 xylogenesis) is hidden within boles (Cuny et al. 2015), and tree or plot scale biomass cannot be
264 directly measured without harvesting trees (Clark & Kellner, 2012). Because of this, the most
265 common approach of estimating woody biomass change is to measure bole diameter growth
266 increment for use with taxa-specific allometric equations (Chave et al., 2014; Chojnacky, Heath,
267 & Jenkins, 2014). This approach is used in both multi-year (e.g. McMahon, Parker, & Miller,
268 2010) and seasonal studies (e.g. Delpierre et al., 2016).

269 We asked whether TBMs captured the phenology of carbon allocation to wood, and the
270 phenology of carbon allocation to leaves, with equal success. We expected that TBMs would be
271 challenged to capture the phenology of carbon allocation to wood because it is hidden, whereas
272 carbon allocation to leaves is more apparent given that this forest contains many deciduous tree
273 species with spring leaf emergence and autumn senescence. To examine the phenological

274 patterns of allocation to wood at Harvard Forest, we estimated the net primary productivity
275 (NPP) allocated to wood (NPP_{wood}) based on allometric regression equations applied to a bi-
276 weekly time series of high accuracy diameter at breast height (DBH) measurements from
277 dendrometer bands (McMahon & Parker, 2014) for three tree species (supporting information
278 appendix S1). We estimated NPP allocated to leaves (NPP_{leaf} , Fig. 2a) at Harvard Forest based
279 on leaf area index and litterfall time series (J. W. Munger, n.d.; W. Munger & Wofsy, 2018;
280 Urbanski et al., 2007; supporting information appendix S2). Resulting NPP_{leaf} and NPP_{wood}
281 reveal that carbon investment in leaves and wood is highest early in the growing season (Fig. 2).
282 The peaks in simulated NPP_{leaf} were within days of the estimated peak NPP_{leaf} (and close to leaf
283 budburst, which typically occurs around May 6 (Keenan and Richardson, 2015). By contrast,
284 simulated phenological patterns in wood-related output variables from three TBMs showed
285 greater variation (Fig. 2, appendix S3). While the Community Land Model version 4.5 (CLM4.5)
286 shows a peak close to that seen in the observations (around the time of budburst, at May 5), the
287 peak for ORCHIDEE_{TRUNK} and CLASS are months later (August 1 and August 11 respectively).



288

289 Figure 2: Seasonality of observed (black \pm gray standard deviation) versus model-simulated
 290 (colors) Net Primary Productivity (NPP) allocated to a) leaf biomass (NPP_{leaf}), and b) woody
 291 biomass (NPP_{wood}) at a temperate deciduous forest. NPP_{leaf} observations were calculated as
 292 $(d\text{LAI}/dt) \cdot \text{LMA} + \text{litterfall}$ where LAI is leaf area index and LMA is leaf mass per area. For
 293 models, NPP_{leaf} is calculated as the change in leaf biomass. NPP_{wood} observations were from
 294 allometry using diameter-at-breast-height (DBH) increment measurements compared with outputs
 295 from three land surface models. For models, NPP_{wood} was calculated as $(X_i - X_{i-1}) / (t_i - t_{i-1})$ where
 296 X is the model output variable most comparable to above-ground woody biomass (in $\text{gC m}^{-2} \text{day}^{-1}$)
 297 for each model (which was vegetation biomass for CLASS, aboveground heartwood plus
 298 sapwood for ORCIDEE_{TRUNK}, and wood biomass for CLM4.5) and t is time in days. Temporal
 299 resolution is 16-day averages. The gray shaded area in all panels indicates the growing season, and
 300 the horizontal dotted line indicates zero. NPP_{leaf} simulations were not available for CLASS. Full
 301 NPP_{wood} estimation and model details are available in online supporting information.

302 In interpreting this model-observation comparison, it is important to remember that using
 303 DBH with allometric scaling equations produces estimates—not direct measurements—of
 304 woody biomass (Clark & Kellner, 2012), and to consider that TBMs differ in how the wood pool
 305 is defined, which is not necessarily identical to above-ground woody biomass (see Table S1 and
 306 S2 for model-specific definitions). In addition, there is some ambiguity in DBH-derived wood
 307 phenology because DBH represents multiple summed variables (Fig. 1a). DBH can be affected

308 by changes in plant water status in addition to changes in biomass, and so seasonal changes in
309 water availability could affect biomass estimations derived from allometry unless a correction is
310 applied (Chitra-Tarak et al., 2015). The actual biomass gain (from xylogenesis) may also lag
311 increases in DBH by weeks (Cuny et al., 2015). Explicit recognition of the distinction between
312 measurable metrics (such as DBH) and the underlying variable we want to characterize or model
313 (such as carbon biomass gain) motivates investigators to quantify uncertainty, and test for
314 scenarios when proxies do not work well.

315 Despite the limitations of the observations and models, the comparison suggests that
316 some models (like CLM 4.5) align moderately well with DBH-derived NPP_{wood} , while others lag
317 DBH-derived NPP_{wood} by months. This divergence in model behaviors highlights the importance
318 of understanding the mechanisms driving both simulated and observed phenologies. The timing
319 of allocation to biomass and wood-related variables in these TBMs is primarily determined by
320 the pattern of NPP across seasons. In ORCHIDEE for example, a fraction of NPP is
321 instantaneously allocated to sapwood, then sapwood biomass is converted into heartwood
322 biomass based on a one-year time constant (Krinner et al., 2005). Ultimately, model NPP is
323 controlled by site-specific climate conditions and representations of forest physiology (e.g. plant
324 functional type), including leaf phenological patterns (e.g. leaf onset/abscission). Thus, in many
325 TBMs, leaf phenology represents wood phenology by proxy because wood allocation is
326 determined by a fraction of NPP, and the cycle of NPP is largely determined by leaf phenology
327 (Delpierre et al., 2016).

328 In contrast with model representations, physiological and tree-ring studies suggest that
329 the mechanisms underlying wood phenology include environmental cues (e.g. temperature;
330 Delpierre *et al.* 2016), ontogeny of wood cells (Cuny et al., 2015; Plomion et al., 2001), and

331 priorities in allocation through time (e.g. allocation of carbon to wood growth versus storage as
332 nonstructural carbohydrates; Richardson, Carbone, et al., 2013a). These factors may play a role
333 in determining when trees are vulnerable or resilient to stress. For example, developing wood
334 cells expand before their cell walls thicken with carbon-rich polysaccharides and lignins, and late
335 wood is more dense than early wood in temperate species (Plomion et al., 2001), so trees at
336 different stages in the sequence of wood development could be more or less sensitive to drought
337 stress. Testing and developing model frameworks for such hypotheses is currently challenging
338 because the timing of carbon allocation to wood is hidden *in vivo*. More direct measurements of
339 wood formation (e.g. Cuny *et al.* 2016), and nonstructural carbohydrates (e.g. Newell, Mulkey,
340 & Wright, 2002), synchronized with frequent measurements of DBH and leaf phenological
341 patterns, would help us to understand and model controls over wood phenological patterns
342 (Delpierre et al., 2016; Guillemot et al., 2017), and how the timing of wood allocation relates to
343 growth strategy, environmental fluctuations, and other plant traits.

344

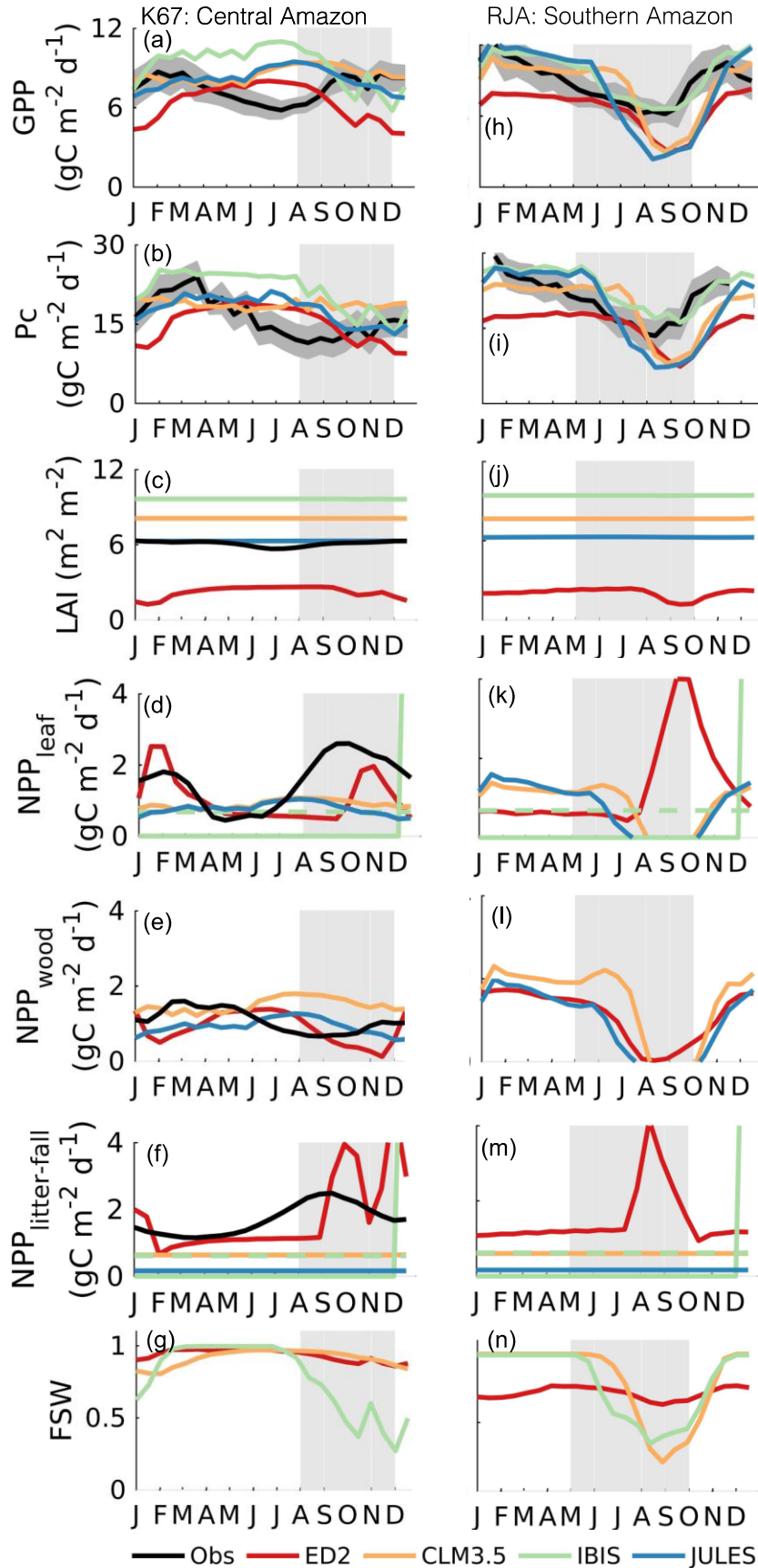
345 **3.2 Amazon evergreen forests: implications of cryptic phenology for seasonality of** 346 **ecosystem carbon fluxes**

347 The challenge of cryptic phenology is not confined to a particular plant organ. Amazon
348 evergreen forests near the equator offer a case study where leaf phenology is ambiguous. Much
349 of the Amazon basin experiences annual wet and dry seasons (Restrepo-Coupe et al., 2013), and
350 this regular seasonal variation in cloud cover and precipitation may select for phenological
351 strategies that match plant activities with resource availability (Doughty et al., 2014; Graham,
352 Mulkey, Kitajima, Phillips, & Wright, 2003; M. O. Jones, Kimball, & Nemani, 2014; van Schaik
353 et al., 1993). Most of the Amazon is remote, making ground-based observations of phenology

354 difficult, especially given the many observations needed to sample the high diversity of tree
355 species (Cardoso et al., 2017) and strategies (Reich, 1995). Observations of canopies from
356 satellites are often obstructed by clouds (Asner, 2001), difficult to interpret (Samanta et al.,
357 2012), and the subject of controversy surrounding technical artifacts and their correction (Huete
358 et al., 2006; Morton et al., 2014; Saleska et al., 2016). Yet many recent studies offer evidence
359 that leaf production, leaf abscission, wood production and root production exhibit annual
360 rhythms in Amazon forests (Doughty et al., 2014; Girardin, Malhi, & Doughty, 2016; Lopes et
361 al., 2016; Wagner et al., 2016).

362 Many TBMs seem to be missing these phenological processes (Restrepo-Coupe et al.,
363 2017). Evidence that TBMs are lacking adequate phenological representation comes from a
364 model inter-comparison for a network of ecosystem flux observations sites (eddy flux towers) in
365 Amazonia (Restrepo-Coupe, et al., 2017). For illustration, we discuss the contrasting cases of
366 equatorial versus southern evergreen forest sites in the Amazon basin of Brazil. At the equatorial
367 site (K67 in the Tapajós National Forest, Brazil), four TBMs showed significant divergence from
368 the estimated interannual pattern of whole-system photosynthetic fluxes (Fig. 3a, gross primary
369 productivity, (GPP; $\text{gC m}^{-2} \text{d}^{-1}$) and a metric of photosynthetic capacity, (Pc; $\text{gC m}^{-2} \text{d}^{-1}$) for K67;
370 (Restrepo-Coupe et al., 2017). The reason for the divergence is that modeled photosynthetic
371 patterns are driven by environmental variability -- measures of soil water stress in this case
372 (model calculated soil water stress index 'FSW' for K67, Fig. 3b-g) -- which suppresses GPP
373 during the long dry season. Yet the observed interannual pattern of photosynthesis in this
374 ecosystem appears to be driven by something beyond instantaneous responses to seasonal
375 climate fluctuations.

376 Since TBMs already include climatic seasonality, their failure to capture GPP seasonality
377 suggests that phenological processes operate at the equatorial site that are separate from the
378 instantaneous physiological responses already represented. Canopy phenological activity could
379 drive the observed GPP via two mechanisms: 1) dry season increases in quantity of canopy
380 leaves (quantified as leaf area index, or LAI) and/or 2) dry season increases in canopy
381 photosynthetic capacity on a per unit area basis (Lopes et al., 2016; Restrepo-Coupe et al., 2017;
382 Wu et al., 2016). Observations of leaf quantity (LAI) from equatorial Amazon sites show that
383 LAI varies little across seasons (e.g. Fig. 3c ‘LAI’ shows low seasonality at K67). Leaf turnover,
384 however, exhibits a dry season pulse (Fig. 3e,f: ‘NPP_{leaf}’ and ‘NPP_{litter-fall}’), suggesting that LAI
385 is maintained because leaf production compensates for near-simultaneous leaf fall during the dry
386 season. As a result, LAI exhibits modest seasonal variation relative to seasonal variation in leaf
387 litterfall and leaf flush (Fig. 4). The seasonality of total LAI also fails to represent within-canopy
388 dynamics, as compensatory leaf area patterns have been identified between the upper and lower
389 canopy levels at K67 (Smith et al., 2019). Since new (recently expanded) leaves have high rates
390 of photosynthesis, replacing old leaves with new leaves can increase photosynthetic capacity of
391 the canopy on a per unit area basis (Albert et al., 2018; Doughty & Goulden, 2008; Niinemets et
392 al., 2012; Pantin, Simonneau, & Muller, 2012; Wu et al., 2016).

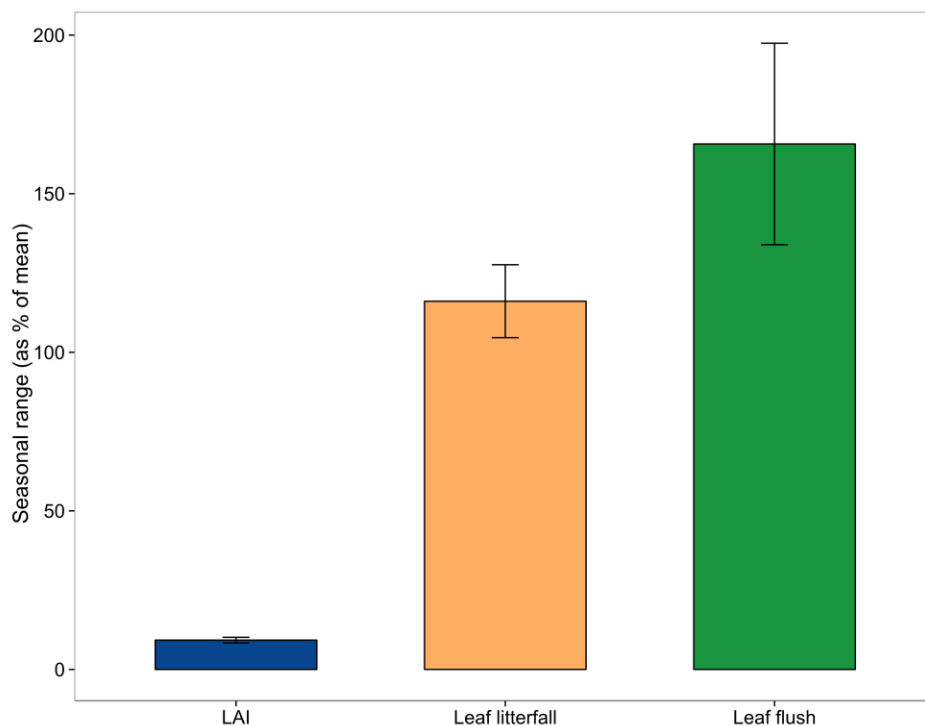


394 Figure 3: Annual cycles of observed (black \pm gray standard deviation) versus model-simulated
395 (colors) forest metrics in two Amazon forests (an equatorial Amazon forest, K67, and a southern
396 Amazon forest, RJA), including (panels from top to bottom): daily average ecosystem-scale
397 photosynthesis (GPP); daily average ecosystem-scale photosynthetic capacity (Pc, GPP at a fixed
398 PAR range ($725\text{-}875\text{umol m}^{-2} \text{s}^{-1}$), vapor pressure deficit, air temperature and light quality
399 measured as cloudiness index (all time mean \pm 1 standard deviation)); leaf area index (LAI); net
400 primary productivity (NPP) allocated to leaves (leaf production; NPP_{leaf}); NPP going to litterfall
401 ($\text{NPP}_{\text{litterfall}}$), and NPP allocated to wood (NPP_{wood}); soil water stress metric (FSW), where 1=no
402 stress (Ju et al., 2006). The light gray shaded box all panels represents the dry season. For K67
403 LAI data, we use data from the control plot of a close-by drought experiment (Juárez et al. 2007;
404 Brando et al. 2010). LAI and NPP observations were not available for the RJA site. Lines are
405 dashed for IBIS NPP to indicate that NPP is allocated only at the end of the year. For further details
406 on model intercomparison, see Restrepo-Coupe et al. (2017).

407 The combination of leaf turnover and leaf age-dependent CO_2 assimilation capacity creates a
408 scenario at K67 in which ecosystem photosynthetic capacity varies more than LAI (Fig. 3b,c).
409 Therefore, the observable canopy total LAI time series does not fully capture phenological
410 patterns of leaf turnover or the resulting shifts in canopy photosynthetic capacity at this site (Fig.
411 5) because leaf phenology is compensatory (Fig. 1c).

412 In contrast to the equatorial Amazon site, at a southern Amazon forest (Reserva Jarú,
413 RJA), observations and models coincide, with GPP and Pc declining during the dry season,
414 consistent with increasing water limitation as the dry season progresses (Fig. 3h,i,n). We lack an
415 observational time series of LAI and litterfall for the southern site, but remote sensing (GLAS
416 satellite lidar) suggests that in the southern Amazon, LAI increases early in the dry season and
417 decreases late in the dry season (Tang & Dubayah, 2017). Thus the equatorial (K67) and
418 southern (RJA) Amazon sites appear to include trees with different phenological strategies
419 (Restrepo-Coupe et al., 2013). We hypothesize that many trees in high water availability
420 equatorial sites may be adapted to optimize light use over time, synchronizing leaf production
421 with the sunny dry season as a strategy for increasing annual carbon gain (Restrepo-Coupe et al.,
422 2017). The tree communities at southern sites like RJA may experience a weaker peak in dry

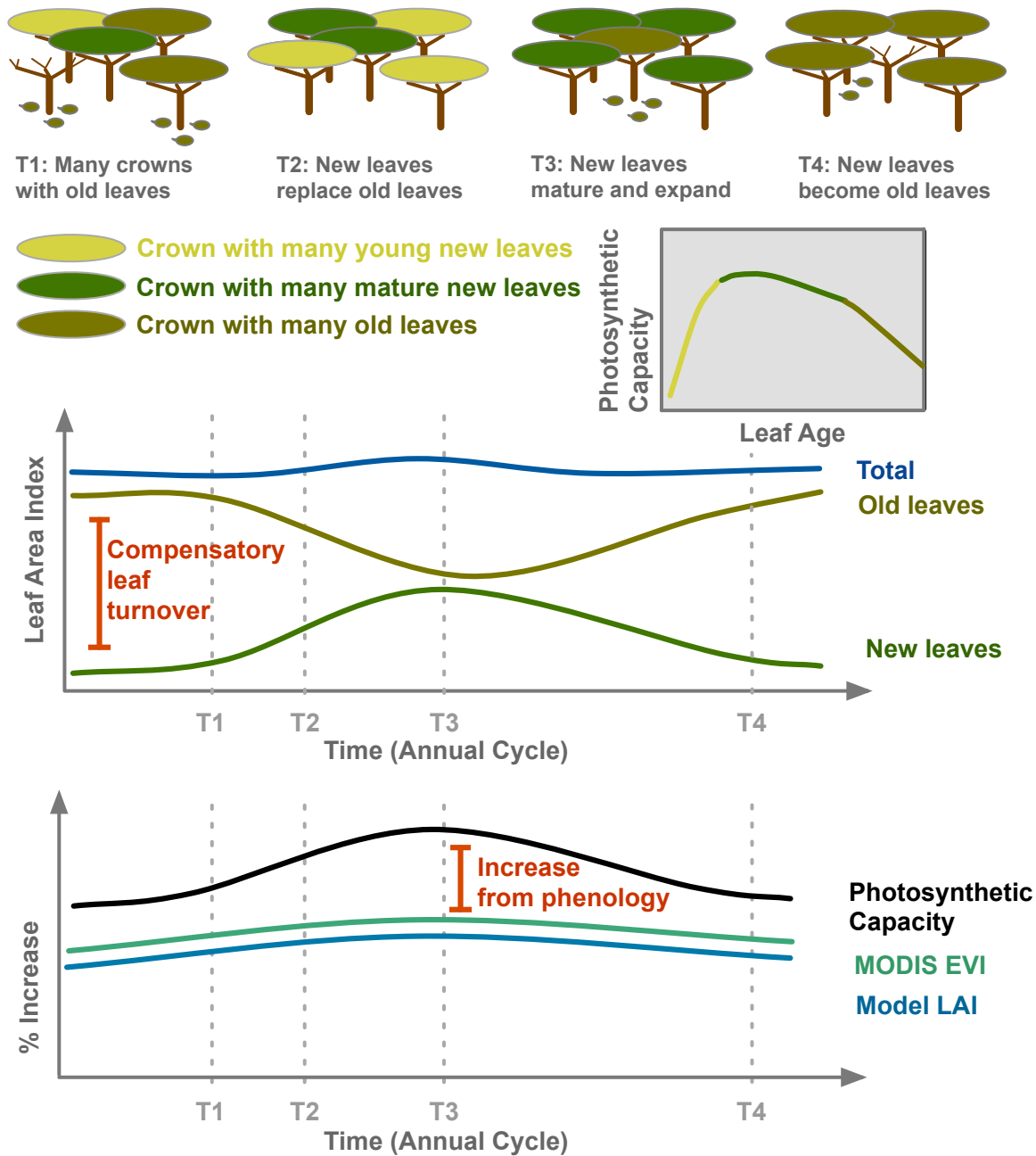
423 season sunlight (Restrepo-Coupe et al., 2013), and may shed leaves during dry seasons to protect
424 plant water status.



425
426 Figure 4: Seasonal canopy dynamics of Leaf Area Index (LAI), leaf litterfall, and leaf production
427 averaged across five Amazonian sites, showing that large leaf turnover is concealed by near-
428 constant LAI. The pulses of litterfall and leaf production support compensatory leaf phenology
429 (Fig 1c) rather than constant leaf phenology (Fig 1b). Bars show mean values of annual amplitude
430 scaled for studies (n=5 sites) of lowland evergreen tropical forests where both LAI and litterfall
431 have been measured. Seasonal range is the annual amplitude scaled by mean value and is calculated
432 as the difference between the maximum dry season value and the minimum wet season value,
433 divided by the mean annual value (%). Error bars show standard deviation of the mean. Studies
434 included in this figure: Tambopata-Candamo Reserve, south-eastern Peru (Girardin et al. 2016);
435 Caxiuana, Floresta Nacional de Caxiuana, Pará, Brazil (Girardin et al. 2015); K83 (Doughty &
436 Goulden 2008) and K67 (Brando et al. 2010 and Malhado et al. 2009) are located in the Tapajós
437 National Forest, Pará, Brazil. Sites experience a range of mean annual precipitation values (1900
438 – 2572 mm).

439 This interpretation is consistent with studies asserting that tropical evergreen forests produce new
440 leaves during periods of high light if they are not strongly water-limited (Doughty & Goulden,
441 2008; Graham et al., 2003; Guan et al., 2015; M. O. Jones et al., 2014; Reich & Borchert, 1984;
442 Restrepo-Coupe et al., 2013; van Schaik et al., 1993; Wu et al., 2016). This continuum between

443 precipitation-driven and light-driven tropical evergreen forest phenological strategies is not
 444 included in most TBMs, and therefore might account for some of the divergence in their GPP
 445 projections (Restrepo-Coupe et al., 2017).



446

447 Figure 5: Illustration of how cryptic leaf turnover creates a phenological pattern in canopy
 448 photosynthetic capacity. Top: Individual crowns drop old leaves and produce new leaves with
 449 some degree of synchronization. Middle: the proportion of leaf area index belonging to previous

450 year's growth (old leaves) and new leaf growth (new leaves) changes through the dry season. Here
451 leaf phenology is difficult to detect because of compensatory inputs and outputs (see Fig 1c). Inset:
452 leaf photosynthetic capacity depends upon leaf age. Lower panel: the combination of leaf turnover
453 and leaf ontogeny increase the canopy photosynthetic capacity, but neither total LAI nor satellite-
454 based proxies for LAI and greenness show this same increase.

455 Equatorial Amazon sites such as K67 provide an example where resolving ambiguous
456 phenology by testing whether leaf phenology is compensatory versus constant, and
457 acknowledging the age-dependent physiology of leaves, is important for understanding and
458 modeling a process, such as forest photosynthesis, at a large scale. Some plant functional types
459 (PFTs) within TBMs allow for photosynthesis to vary with leaf age, but with a focus on
460 temperate deciduous plants. For example, the Joint UK Land Environment Simulator (JULES)
461 accounts for damage and senescence accumulation by reducing photosynthesis during the
462 growing season (Clark et al., 2011), and the Ecosystem Demography model (ED2) decreases the
463 maximum carboxylation rate of Rubisco (V_{cmax}) in the autumn as a function of Julian day
464 utilizing historical MODIS data (Medvigy, Wofsy, Munger, Hollinger, & Moorcroft, 2009). In
465 these cases, time of year or 'season' serves as a proxy for leaf age, which may work well for
466 some PFTs, but not for tropical evergreen broadleaf forests where the 'evergreen' canopy belies
467 cyclic leaf turnover that the PFT ruleset does not include. This case study suggests that
468 accounting for cryptic phenology is necessary for correctly detecting, attributing, and modeling
469 the carbon exchange dynamics of tropical forests (De Weirdt et al., 2012; Y. Kim et al., 2012;
470 Manoli, Ivanov, & Fatichi, 2018; Restrepo-Coupe et al., 2017).

471

472 **4. IMPLICATIONS OF CRYPTIC PHENOLOGY FOR PREDICTION ACROSS** 473 **SCALES**

474 Fine-scale processes, integrated over space and time, create large-scale exchanges of mass and
475 energy between the biosphere and the atmosphere (Monson & Baldocchi, 2014). Here we

476 consider some of the fine scale processes associated with cryptic phenology that, scaled up, have
477 implications for our ability to understand, model, and predict biosphere-atmosphere interactions
478 under climate change.

479

480 *Organ scale*

481 Plant traits can show very high within-species variation due to phenology (Chavana-
482 Bryant et al., 2017), and this variation can surpass interspecific variation for some traits (Fajardo
483 & Siefert, 2016). Specifically, leaf development and aging is associated with changes in internal
484 leaf structure (Lim, Kim, & Gil Nam, 2007; Niinemets et al., 2012), concentrations of secondary
485 metabolites (Z. Liu et al., 1998; Virjamo & Julkunen-Tiitto, 2014), emissions of volatile organic
486 compounds (Alves, Harley, Goncalves, da Silva Moura, & Jardine, 2014; Niinemets et al., 2010),
487 and metabolic rates (Albert et al., 2018; Niinemets et al., 2012; Pantin et al., 2012). For the goal
488 of scaling fluxes from leaves to canopies, these many physiological changes associated with leaf
489 age suggest that distinguishing between constant leaf phenology and compensatory leaf
490 phenology is important not only for tropical forests (as we describe in the first case study above),
491 but for evergreen forests in general.

492 Similarly, root production is accompanied by physiological changes. There are species-
493 specific relationships between root age and root physiology such as respiration rates and nutrient
494 uptake capacity (Bouma et al., 2001; Fukuzawa, Dannoura, & Shibata, 2011). Existing studies
495 that have characterized the hidden phenology of roots have shown evidence of interspecific
496 differences in cycles of fine root production—single flushes, multiple flushes, or constant
497 growth—that could represent strategies for responding to seasonal changes in climate or resource
498 availability (Fukuzawa et al., 2011; McCormack et al., 2014).

499

500 *Organismal scale*

501 Natural selection would be expected to favor coordination in the timing of resource acquisition
502 with resource storage and allocation (Sala, Woodruff, & Meinzer, 2012). Since selection acts at
503 the level of individuals, if we want to understand the adaptive value of phenological changes, we
504 need to understand how all plant organs function together, as a unit, through time. It is very
505 difficult to study ‘whole’ plants in the wild, especially woody plants. Few (if any) studies have
506 quantified the phenologies of all plant organs in wild woody plants to gain an integrated
507 organism-level perspective on phenology (but see Hu et al., (2010) for whole-tree carbon
508 assimilation during the growing season; see Würth, Peláez-Riedl, Wright, & Korner (2005) for
509 seasonal variation in non-structural carbohydrate pools by plant organ; and see Doughty et al.,
510 (2014) for an example plot-scale study of wood, fine root, and canopy phenology). Studies
511 examining phenologies of two organs suggest that phenology is often asynchronous across
512 organs (Abramoff & Finzi, 2015; Wagner, Rossi, Stahl, Bonal, & Hérault, 2013). Comparing
513 phenological patterns of roots and shoots frequently reveals offsets between maximum root
514 growth and shoot growth, and these offsets vary across biomes (Abramoff & Finzi, 2015). In
515 tropical forests, leaf and wood production is often asynchronous (Wagner et al., 2013). The onset
516 and/or termination of growth may also vary; roots in temperate deciduous white oak, for
517 example, continue to elongate in winter after senescence of leaves (Teskey & Hinckley, 1981).
518 Nonstructural carbohydrate reserves also show phenological patterns that are species-dependent
519 (Würth et al., 2005) and affected by phenological patterns of leaves (Palacio, Maestro, &
520 Montserratmarti, 2007). Rates of carbon use regulate carbon uptake in plants (sink-driven
521 photosynthesis; (Fatichi, Leuzinger, & Korner, 2014), so phenological changes in carbon

522 demand should impact the timing of photosynthetic activity.

523 These findings show that capturing the patterns and drivers of hidden and ambiguous
524 phenologies will be needed for a comprehensive understanding of how plants prioritize amongst
525 competing uses of resources and maintain carbon balance, with implications for modeling plant
526 resource use. In most TBMs, the temporal patterns of leaf activity (the size of the leaf pool and
527 the rate of photosynthesis) drive temporal patterns of carbon allocation because carbon allocation
528 to other plant organs is often modeled as a constant proportion of carbon uptake (Abramoff &
529 Finzi, 2015; Delpierre et al., 2016; Guillemot et al., 2017). However, if different plant organs
530 respond to different environmental drivers (Wagner et al., 2016), then models that use leaf
531 activity to generate interannual patterns of activity in hidden organs may fail to simulate
532 observed patterns of root or bole activity at seasonal timescales.

533 How plants prioritize their allocation, through time, to various plant organs or to storage
534 may have consequences for plant resilience or vulnerability to extreme events, and several
535 studies already show that plant vulnerability and/or resilience to extreme events varies due to
536 phenological status and/or season (Craine et al., 2012; M. Huang, Wang, Keenan, & Piao, 2018).
537 We suggest that the timing of extreme events in relation to plant phenological status may be
538 necessary for predicting plant community responses to future climate. For example, plant
539 tolerance to drought or cold could depend on nonstructural carbohydrates (Dietze et al., 2014;
540 Sala et al., 2012), and nonstructural carbohydrates follow seasonal cycles that could indicate
541 internal phenology (Richardson, Carbone, et al., 2013a). Tests of such hypotheses are timely,
542 given that the frequency of extreme climate events is increasing under global climate change
543 (Bellprat & Doblas-Reyes, 2016; Ummenhofer & Meehl, 2017).

544

545 *Community scale*

546 As climate changes, many studies have demonstrated that phenological patterns shift,
547 impacting species interactions (CaraDonna, Iler, & Inouye, 2014; Memmott, Craze, Waser, &
548 Price, 2007; Miller-Rushing, Hoyer, Inouye, & Post, 2010; Polgar & Primack, 2011; Rafferty,
549 CaraDonna, & Bronstein, 2014; Yang & Rudolf, 2010). Fewer studies have probed how hidden
550 phenologies shape species interactions, or how those interactions may be changing. To do so
551 could reveal that phenology mediates impacts of species interactions on plant mortality,
552 reproduction, and metabolism. For example, the timing of insect outbreaks in relation to
553 nonstructural carbohydrate reserves (which are affected by the timing of leaf renewal) may
554 explain interspecific differences in tolerance to defoliation (Chen, Wang, Dai, Wan, & Liu,
555 2017). Further investigation into how species interactions affect hidden phenologies would help
556 gain a more complete understanding of the interplay between climate change, whole plant
557 physiology, and species interactions.

558

559 *Ecosystem to global scale*

560 Projections of Earth's future climate are particularly sensitive to uncertainties in the land
561 carbon cycle (Friedlingstein et al., 2014). Improving representation of the land carbon cycle in
562 TBMs requires understanding the drivers of phenology, and the role of phenology in mediating
563 biosphere-atmosphere exchanges (Richardson, Keenan, et al., 2013b). Recognizing phenological
564 rhythms at scales from plant organs to communities is prerequisite to identifying their role in
565 large scale (ecosystem to global) cycling of carbon. For example, investigating the distribution of
566 root ages at different times of the year could elucidate larger scale autotrophic respiration or soil
567 resource acquisition processes (because root age affects root respiration and nutrient uptake

568 capacity; Bouma et al., 2001). TBMs which are calibrated to match current observations, but that
569 include inaccurate relationships between drivers and vegetation responses, risk making biased
570 predictions of forest response to future climate changes because they do not incorporate
571 underlying biological mechanisms (De Weirdt et al., 2012; Restrepo-Coupe et al., 2017).

572

573 **5. RECOMMENDATIONS FOR MEASURING AND MODELING CRYPTIC**

574 **PHENOLOGY**

575 **5.1 Recommendations for empirical research**

576 To reveal cryptic phenological patterns empirically, we need to consider the target, frequency,
577 and methods of measurements. We recommend complementing existing studies and
578 measurements of above-ground, clearly visible phenological changes with measurements of
579 hidden phenological changes (Table 1). Specifically, we need more time series of development
580 and growth of roots (e.g. Abramoff and Finzi 2015, McCormack et al 2014), and internal
581 structures (e.g. Cuny et al. 2015), to learn when leaf phenology directly fuels the phenological
582 patterns of other plant organs (and thus can represent them by proxy), and when it does not.
583 Building upon studies examining synchrony in phenology of multiple plant organs (Bazié et al.,
584 2017; Delpierre et al., 2016; Michelot, Simard, Rathgeber, Dufrene, & Damesin, 2012; Omondi,
585 Odee, Ongamo, Kanya, & Khasa, 2016; Perrin, Rossi, & Isabel, 2017; Wagner et al., 2013),
586 whole-plant phenology studies in which all plant organs and their associated processes
587 (acquisition and allocation of carbon, water, and nutrients) are continuously monitored in the
588 same individual plants across seasons could elucidate the relationship between the phenology of
589 plant organs with each other, and with climate, and test the representation of phenology for
590 various PFTs.

591 Revealing cryptic phenological patterns will require more studies explicitly testing
592 whether compensatory processes (Fig. 1c) mistaken for constancy (Fig. 1b) mislead our
593 interpretation of mass, area, or volume time series. To this end, sampling schemes need to go
594 beyond measuring mass, area or volume of plant organs or “pools” (in aggregate) to also
595 measure rates of inputs and outputs to and from organs/pools across time. (Since mass-balance
596 equations have three terms—inputs, outputs, and the accumulated pool—at least two must be
597 measured to obtain a single solution). For example, litterfall time series should be collected to
598 correspond with total leaf area time series. To examine the metabolic consequences of constant
599 versus compensating phenology, we need more (1) measurements of plant organ activity as
600 organs develop and age, and (2) experiments manipulating phenological status to test the
601 interaction between phenology and physiology (including photosynthesis and respiration) under
602 various treatments (e.g. drought, temperature, and herbivory). When a measured variable (e.g.
603 mass or volume) is the sum of multiple component variables (Fig 1a) then those components
604 should be characterized (if possible) in tests for scenarios when the time series of the measured
605 variable is not aligned with that of the component variable of interest. Fourier analysis is a
606 promising tool for decomposing phenological cycles (Bush et al., 2016), and should be explored
607 for revealing phenology that is otherwise ambiguous.

608 Phenological events can happen quickly and vary across landscapes. Fine temporal and
609 spatial resolution will capture patterns that might otherwise be missed (e.g. Smith et al. 2019).
610 How we measure phenological patterns has moved beyond plant-level measurements to
611 landscape measurements as technology has evolved, and we need to continue expanding our
612 capacity for detecting plant phenological activity at multiple spatial scales (e.g. leaves to
613 canopies to landscapes). Remote sensing technologies offer valuable tools for gathering

614 phenological data on large spatial scales. Chlorophyll fluorescence remote sensing products
615 promise to test the physiological interpretation of ‘greenness’ from the more traditional MODIS
616 products (Guan et al., 2015; Lee et al., 2013; Porcar-Castell et al., 2014). Continuous or frequent
617 high resolution near-surface remote sensing instrumentation such as phenocams (Klosterman et
618 al., 2014; Lopes et al., 2016; Wu et al., 2016) and lidar (Calders et al., 2015) offer finer spatial
619 resolution data to complement and potentially validate satellite-based phenology-related
620 products.

621 Although satellite-derived products are valuable tool for phenology (e.g. Guan et al
622 2015), some phenological patterns remain cryptic when relying on remote sensing tools.
623 Reflectance-based indices from satellites reveal more about the phenological status of upper
624 canopy leaves and shoots than about the hidden phenological activity of roots, boles, and internal
625 plant processes. Further development of remote sensing tools may help reveal hidden
626 phenologies; for example, lidar can be used to estimate the LAI of understory plants, helping
627 infer leaf phenological patterns for canopy layers that are hidden from other sensors (Tang &
628 Dubayah, 2017). The development of high-throughput methods for evaluating gene expression
629 (Kris et al., 2007), together with the growing databases of annotated genomes, offer the
630 opportunity to complement above-ground measurements with information about regulation of
631 internal or below-ground activities. We urge more tests to evaluate when remote sensing signals
632 do, and do not, link to phenology, including time series of comparisons between remote sensing
633 signals and plant-level measurements (e.g. changes in leaf production or woody biomass).

634

635 **5.2 Recommendations for model development**

636 In TBMs, plant structures (e.g. leaves) are produced or shed, and processes are switched ‘on’ or
637 ‘off’, based on rule sets about temperature, moisture, and photoperiod, or (in about a third of
638 TBMs), are prescribed based on remotely sensed indices and other derived products instead of
639 being simulated internally (Fisher et al., 2014; Huntzinger et al., 2012). In either case, the TBM
640 representation of phenological processes relies heavily on observations that are readily collected
641 at large scales, such as climate data and satellite-based remote sensing products. We need to
642 determine when this reliance on apparent phenology limits our ability to make robust long-term
643 predictions of terrestrial carbon, water, and energy budgets or future boundary shifts of biomes.

644 A process or parameter in a model is important, in terms of our predictive ability, if it
645 causes large changes in a response that we want to predict (high sensitivity), and/or if it is highly
646 uncertain (Dietze, 2017). For TBMs, we need more sensitivity analyses that evaluate the impact
647 of including or excluding potential phenological schemes, and uncertainty assessments that
648 quantify sources of uncertainty (e.g. Migliavacca et al., 2012). Specific phenological dynamics
649 ripe for possible implementation in TBMs include asynchronous allocation to various plant
650 organs (e.g. through prioritization schemes or time lags), environmental controls over carbon
651 allocation (Guillemot et al., 2017), and plant organ age-dependency of metabolic capacity (e.g.
652 photosynthetic capacity as a function of leaf age and root respiration as a function of root age;
653 (Albert et al., 2018; De Weirdt et al., 2012; Fukuzawa et al., 2011). By examining the sensitivity
654 of modelled ecosystem-scale fluxes to such processes, modelers can strike a balance between
655 over-parametrizing versus excluding important processes in TBM models. Knowledge of which
656 phenological states, processes, and parameters within models show high sensitivity or
657 uncertainty can also help guide empirical research priorities.

658 Evaluations of uncertainty and sensitivity require first having model formulations of
659 phenology. As we have argued, study efforts are not uniform, and phenological patterns may be
660 cryptic such that they can only be resolved with multiple measurements (e.g. inputs and outputs
661 or multiple variables). In these cases, it may be difficult to find enough information to develop
662 phenology schemes. Model-data comparisons, with observational data coming from multiple
663 independent sources (and multiple organs) at multiple scales (e.g. eddy covariance time series,
664 and measurements of allocation in individuals) should help determine if an important
665 phenological process could be wholly missing from models. Joint model and empirical efforts
666 can then identify, characterize, model, and evaluate the importance of the excluded phenological
667 processes.

668 Finally, we emphasize the value of drawing upon empirical and theoretical ecology,
669 evolution, and physiology for the development and refinement of phenological models. In
670 systems where the temporal dynamics of plant acquisition and allocation have been shown to be
671 under selection to increase fitness within climatic and biological constraints, optimization models
672 may be useful (e.g. Caldararu, Purves, & Palmer, 2014; Kikuzawa, 1991; 1996), but they should
673 be expanded to include multiple resources (e.g. moisture and nutrient optimization in addition to
674 carbon), and trade-offs between multiple purposes, such as growth and reproduction (Iwasa,
675 2000). However, it is also important to recognize that life history imposes temporal structure
676 relevant to modelling at the seasonal time scale, such as timelines for recruitment, maturation,
677 and mortality in annual plants, or timelines for development of the photosynthetic apparatus in
678 new leaves with different lifespans. Thus, a valuable challenge will be to formalize demographic
679 and physiological timelines in models and test their impact on model sensitivity and uncertainty.
680

681 **6. CONCLUSION**

682 A growing body of research shows that capturing cryptic phenologies is required for a complete
683 picture of seasonal resource allocation and acquisition strategies, constraints, and consequences
684 across many scales. Understanding the full scope of cyclic and recurrent biological events in
685 plants is critical for advancing our understanding of plant ecology and evolution, and for
686 predicting responses and feedbacks to climate change. We call for further recognition and
687 exploration of cryptic phenologies—including compensatory processes, non-structural
688 carbohydrates dynamics, wood formation, and root production —through new technologies,
689 TBM development, and time series of intensive plant-scale measurements.

690

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