

1 **Short-term effects of fine particulate matter on acute myocardial infraction**
2 **mortality and years of life lost: a time series study in Hong Kong**

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13 **Abstract**

14 Previous studies have applied years of life lost (YLL) as a complementary indicator to assess the
15 short-term effect of the air pollution on the health burden from all-cause mortality, but sparsely focused
16 on individual diseases such as acute myocardial infraction (AMI). In this study, we aimed to conduct a
17 time-series analysis to evaluate short-term effects of fine particulate matter (PM_{2.5}) on mortality and YLL
18 from AMI in Hong Kong from 2011 to 2015, and explore the potential effect modifiers including sex and
19 age by subgroup analysis. We applied generalized additive Poisson and Gaussian regression model for
20 daily death count and YLL, respectively. We found that per 10 µg/m³ increment in concentration of PM_{2.5}
21 lasting for two days (lag₀₁) was associated with a 2.35% (95% CI 0.38% to 4.36%) increase in daily
22 mortality count and a 1.69 (95%CI 0.01 to 3.37) years increase in YLL from AMI. The association
23 between PM_{2.5} and AMI mortality count was stronger among women and older people than men and
24 young people, respectively. We concluded that acute exposure to PM_{2.5} may increases the risk of mortality
25 and YLL from AMI in Hong Kong and this effect can be modified by age and gender. These findings add to the
26 evidence base for public health policy formulation and resource allocation.

27 **Key Words**

28 PM_{2.5}; AMI; mortality; years of life lost; time-series study; Hong Kong

29

30 **1. Introduction**

31 Cardiovascular disease (CVD) as the third leading cause of deaths in Hong Kong, accounted for 13.2% of
32 all deaths in 2015; ischemic heart disease (IHD) as a major category was responsible for 66.6% of CVD
33 deaths (Centre for Health Protection, 2017). Acute myocardial infraction (AMI), as an important
34 manifestation of IHD (Wichmann et al., 2014), is one of major public health concerns in Hong Kong and
35 it is urgent to assess the burden of AMI and related risk factors.

36 Ambient air pollution is a large threat to public health in the world (WHO, 2016). Hong Kong is
37 experiencing deteriorating air quality and the health impacts of air pollution might be even higher than
38 those in the developing countries in South Asia (Wong et al., 2008). Fine particulate matter (PM_{2.5}),
39 defined as atmospheric particulate matter with aerodynamic diameter ≤ 2.5µm, is one of the principal air
40 pollutants in Hong Kong. PM_{2.5} is a mixture of various compounds including chemical and biological
41 ingredients rather than a self-contained pollutant and is associated with a wide range of adverse health
42 effects mainly including respiratory and cardiovascular diseases (Kim et al., 2015).

43 The short-term effects of PM_{2.5} morbidity and mortality risk of AMI have been demonstrated in numerous

44 epidemiologic research studies in the world (Lanki et al., 2006; Nuvolone et al., 2011; von Klot et al.,
45 2011; Wang et al., 2015; Wang et al., 2016; Wichmann et al., 2014). However, the short-term effect may
46 differ because of the varying exposure level, components and the characteristic of population in different
47 geographic locations (HEI, 2010). No studies have been conducted in Hong Kong to examine the
48 association between acute exposure to PM_{2.5} and AMI mortality. Moreover, mortality count alone depicts
49 only a partial story of disease burden. Years of life lost (YLL), taking premature deaths and the life
50 expectancy at death into consideration, would be an important complementary index to reflect the health
51 burden due to air pollution, which is significant for public policy making and health service planning
52 (Guo et al., 2013). Yet, studies applying YLL to quantify the disease burden have been sparse so far (Guo
53 et al., 2013; He et al., 2016; Lu et al., 2015; Zhu et al., 2017) and to the best of our knowledge the
54 short-term effect of PM_{2.5} on YLL from AMI and the potential effect modification by demographic
55 factors such as sex and age have not been investigated.

56 We performed a time-series study to evaluate the short-term effect of fine particulate matter on mortality
57 and YLL from AMI, from 2011 to 2015 in Hong Kong, and explored the potential effect modification by
58 sex and age.

59

60 **2. Materials and methods**

61 **2.1 Data collection**

62 **2.1.1 Mortality data**

63 Daily data on mortality due to AMI in Hong Kong from January 1, 2011 to December 31, 2015 were
64 obtained from Hong Kong Census and Statistics Department (CSD). The anonymous records provided
65 information such as sex, age, date of death, and underlying death cause which was coded according to the
66 International Classification of Diseases, Tenth Revision (ICD-10). In this study, the daily mortality count
67 from AMI (ICD-10: I21) was abstracted and stratified by sex and age group (≤ 65 and >65 years old).
68 Since we only used aggregated data rather than individualised data in this study, ethics approval and
69 consent from individual subjects were not required by our institute.

70 **2.1.2 YLL data**

71 Life tables for Hong Kong population from 2011 to 2014 were obtained from Hong Kong CSD (Census
72 and Statistics Department, 2015), which provided the life expectancy at every exact age for males and
73 females respectively. Life table for the year 2015 was unavailable, so we used life expectancies in 2014 as
74 a substitute to compute the YLL for 2015. YLL values were calculated by matching sex and age to the life

75 tables and daily total YLL were calculated as the sum of YLL of all deaths due to AMI on the same day
76 (Guo et al., 2013). The daily YLL data were also stratified by sex and age.

77 **2.1.3 Air pollution and meteorology data**

78 Hourly monitoring data for PM_{2.5}, particulate matter with aerodynamic diameters less than 10µm (PM₁₀),
79 sulphur dioxide (SO₂), nitrogen dioxide (NO₂) and ozone(O₃) in 14 monitoring stations were collected by
80 the Hong Kong Environmental Protection Department from 2011 to 2015. Excluding three roadside
81 stations and one general station on a remote island, we used the data of the remaining 10 general stations
82 to represent the general population exposure on a regular basis. We calculated the twenty-four hour mean
83 concentrations of PM_{2.5}, PM₁₀, SO₂, NO₂ and O₃ for each station first and then averaged over ten stations
84 to represent the daily exposure levels of the whole population. The daily mean temperature and relative
85 humidity during 2011-2015 were obtained from the Hong Kong Observatory.

86

87 **2.2 Statistical analysis**

88 **2.2.1 Spearman correlation**

89 The correlation between air pollutants and meteorological conditions was evaluated by Spearman's rank
90 correlation test.

91 **2.2.2 Association between PM_{2.5} and daily mortality count for AMI**

92 We applied generalized additive Poisson regression model to estimate the association between daily
93 mortality count for AMI and daily concentration of PM_{2.5}. We applied smoothing spline functions to
94 control for secular trend and seasonality in daily mortality count, daily mean temperature, and relative
95 humidity (*Humidity*₀). To adjust for the immediate and delayed effects of temperature, daily mean
96 temperature of the same day (*Tmean*₀) the moving average of lag 1-3 days (*Tmean*₁₋₃) were included in the
97 multiple regression model. The day of the week (*DOW*) and public holidays (*Holiday*) as dummy
98 variables were also included in the model. Following the methods in previous studies (Bell et al., 2008;
99 Peng et al., 2008; Qiu et al., 2012), we applied degrees of freedom (*df*) of 7/year for the time trend, 6 for
100 *Tmean*₀ and *Tmean*₁₋₃, and 3 for relative humidity. The basic model was:

$$101 \quad \log(E(Y)) = \alpha + s(\text{time}, df=7/\text{year} \times 5 \text{ years}) + s(Tmean_0, df=6) + s(Tmean_{1-3}, df=6) \\ 102 \quad + s(Humidity_0, df=3) + \beta_1 \times DOW + \beta_2 \times Holiday$$

103 where $E(Y)$ represents the expected daily mortality count for AMI and $s(.)$ the smoothing spline function
104 for nonlinear variables. Residual plot and partial autocorrelation function (PACF) plot demonstrated the

105 successful control for secular trend and seasonality. The association of daily mortality count for AMI with
106 PM_{2.5} over two days, which was the moving average concentration over the same day and the previous
107 day (lag₀₋₁) was included as the main analysis; association of AMI with PM_{2.5} over the same day (lag₀)
108 and three days before (lag₁ to lag₃) was also examined. We first fitted single pollutant models for PM_{2.5}
109 and the other three air pollutants (NO₂, SO₂ and O₃) and then included pollutants in multiple regression
110 models. To find out whether there is effect modification by sex and age, we also examined the pollution
111 and disease association in the subgroups and calculated the 95% confidence interval (CI) for
112 difference: $(\beta_1 - \beta_2) \pm 1.96\sqrt{SE_1^2 + SE_2^2}$, where β_1 and β_2 are the estimates for two subgroups and
113 SE₁ and SE₂ are their standard errors respectively (Schenker and Gentleman, 2001).

114 **2.2.3 Association between PM_{2.5} and daily YLL for AMI**

115 We applied generalized additive Gaussian models to examine the association of PM_{2.5} with YLL for AMI
116 because daily YLL for AMI followed a normal distribution according to the previous studies (Guo et al.,
117 2013; Zhu et al., 2017). In the current study, the distribution of YLL from AMI and the plot of model
118 residuals showed that the normality was not violated (Figure S1).

119 All statistical analysis was performed with the *mgcv* package in R software, version 3.4.0. The results
120 were presented in percent excess risk of daily mortality count (ER %) or the increment in YLL for AMI
121 per 10µg/m³ increase of PM_{2.5}.

122

123 **3. Results**

124 **3.1 Descriptive statistics**

125 During the 1,826 days from 1 January 2011 to 31 December 2015, a total of 9,252 deaths due to AMI
126 were recorded. The means of daily deaths and YLL due to AMI were 5.1 cases and 68.3 person years,
127 respectively. Both daily death counts and YLL were higher for men and older people (age>65 years old)
128 than women and younger people (age≤65 years old), respectively (**Table 1**).

129 The daily mean concentration of PM_{2.5} was 29.1µg/m³, with a SD of 17.3µg/m³. For the other air
130 pollutants, the daily mean concentrations of PM₁₀, NO₂, SO₂ and O₃ were 43.7µg/m³, 54.5µg/m³,
131 11.6µg/m³ and 40µg/m³ respectively (**Table 1**). PM_{2.5} was strongly correlated with PM₁₀ (Spearman
132 correlation coefficient r=0.98), NO₂ (r=0.74), and moderately correlated with O₃ (r=0.58) and SO₂
133 (r=0.41). Mean temperature and relative humidity were negatively correlated with PM_{2.5} (**Table 2**).

134 **Figure S2** shows the time trend and daily variation of air pollution as well as the mortality count and

135 YLL from AMI.

136

137 **Table 1| Levels of daily PM_{2.5}, mean temperature, relative humidity, YLL and daily death counts**
138 **for AMI in Hong Kong, China, 2011-15 (Number of days=1826)**

	Minimum	25% quartile	Median	75% quartile	Maximum	Mean	Standard deviation
Pollution concentration ($\mu\text{g}/\text{m}^3$)							
PM _{2.5}	4.9	14.6	25.9	39.5	115.6	29.1	17.3
PM ₁₀	7.6	23.9	38.8	57.9	157.4	43.7	23.8
NO ₂	12.9	41	51.8	64.3	162.3	54.5	18.1
SO ₂	3.3	7.5	10.4	14.3	46.9	11.6	5.8
O ₃	4.7	20.6	34.1	55.3	134.4	40.0	23.6
Meteorology measures							
Mean temperature(°C)	8.4	19	24.8	28.2	32.4	23.5	5.3
Relative humidity (%)	29	74	79	85	99	78.3	10.3
YLL (years)							
Total	0	39.5	63.2	91.3	255.3	68.3	39.5
Women	0	7.9	18.4	33.9	109.7	23.2	20.1
Men	0	19.6	39.3	64.6	202.5	45.1	32.8
Age \leq 65 years	0	0	24.8	48.0	175.0	30.2	31.5
Age >65 years	0	21.8	35.3	53.2	136.2	38.4	22.4
Daily death counts (No of deaths)							
Total	0	3	5	7	16	5.1	2.5
Women	0	1	2	3	9	2	1.5
Men	0	2	3	4	12	3	1.9
Age \leq 65 years	0	0	1	2	6	1	1.1
Age >65 years	0	2	4	5	13	4	2.2

139 Abbreviations: YLL, year of life lost; PM_{2.5}, particles with an aerodynamic diameter less than 2.5 μm ;
140 PM₁₀, particles with an aerodynamic diameter less than 10 μm ; NO₂, nitrogen dioxide; SO₂, sulphur
141 dioxide; O₃, ozone.

142

143
 144 **Table 2| Spearman correlation between air pollutants and weather conditions in Hong Kong, China,**
 145 **during 2011–15**

	PM ₁₀	NO ₂	SO ₂	O ₃	Mean temperature	Relative humidity
PM _{2.5}	0.98*	0.74*	0.41*	0.58*	-0.53*	-0.48*
PM ₁₀	—	0.72*	0.40*	0.62*	-0.52*	-0.53*
NO ₂	—	—	0.53*	0.29*	-0.43*	-0.32*
SO ₂	—	—	—	-0.05	0.04	-0.48*
O ₃	—	—	—	—	-0.16*	-0.47*
Mean temperature	—	—	—	—	—	0.11*

146 *P<0.01

147 Abbreviations: PM_{2.5}, particles with an aerodynamic diameter less than 2.5 µm; PM₁₀, particles with an
 148 aerodynamic diameter less than 10 µm; NO₂, nitrogen dioxide; SO₂, sulphur dioxide; O₃, ozone.

149

150 3.2 Modelling results

151 **Table 3** presents the excess risk estimates in AMI mortality and YLL per 10µg/m³ increase in PM_{2.5} on
 152 different lag days in single pollutant models. PM_{2.5} was associated with daily mortality on lag₀ and lag₁
 153 days; the largest risk estimates were found with lag₀₋₁. The association between PM_{2.5} and YLL was
 154 statistically significant at lag₀₋₁.

155 **Table 4** summarizes the relationship between PM_{2.5} at lag₀₋₁ and AMI in single and co-pollutant models.
 156 An increment of PM_{2.5} by 10µg/m³ at lag₀₋₁ was associated with 2.35% (95% CI: 0.38% to 4.36%)
 157 increase in daily mortality and 1.69 (95%CI: 0.01 to 3.37) years increase in daily YLL from AMI. SO₂
 158 was also associated with increased risk of daily death and YLL from AMI, while no associations were
 159 found with NO₂ or O₃. In the co-pollutant models, the risk estimates for PM_{2.5} changed slightly.

160 The exposure–response relationship was approximately linear, as seen in **Figure S3**.

161 **Table 5** shows the excess risk estimates in AMI mortality and YLL per 10µg/m³ increase in PM_{2.5} at lag₀₋₁
 162 in single pollutant models for different sex and age groups respectively. The AMI mortality risk estimates
 163 were higher for women and elders than for men and young people, respectively. The association of PM_{2.5}
 164 with YLL from AMI did not vary by sex or age group.

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166

167

168 **Table 3| Association between 10 µg/m³ increase in PM_{2.5} and YLL and increase in deaths for AMI**
 169 **by lags using single pollutant models during 2011-15***

Lag days	Increase in YLL(years)	ER of daily mortality count (%)
Lag0	1.47 (-0.05, 3.00)	1.82 (0.04, 3.62)
Lag1	1.20 (-0.26, 2.66)	1.87 (0.16, 3.60)
Lag2	0.59 (-0.83, 2.01)	1.13 (-0.53, 2.81)
Lag3	0.30 (-1.09, 1.69)	1.40 (-0.23, 3.05)
Lag01†	1.69 (0.01, 3.37)	2.35 (0.38, 4.36)

170 * Generalized additive Poisson model for mortality count and Gaussian model for YLL was applied and
 171 controlled for long-term trend, seasonality, weather factors, calendar effect. Statistically significant effect
 172 estimates are in bold.

173 †Overall cumulative effects of PM_{2.5} lasting for 0–1 days were estimated

174 Statistically significant effect estimates are in bold.

175 ER, excess risk; PM_{2.5}, particles with an aerodynamic diameter less than 2.5 µm.

176

177 **Table 4| Association between 10µg/m³ increase in PM_{2.5} (lag 0-1 day) and YLL and increase in**
 178 **deaths for AMI using single, two, and three pollutant models during 2011-15***

	Increase in YLL(years)	ER of daily mortality count (%)
Single pollutant model		
PM_{2.5}	1.69 (0.01, 3.37)	2.35 (0.38, 4.36)
NO₂	0.62 (-0.92, 2.17)	1.50 (-0.36, 3.39)
SO₂	4.97 (0.28, 9.66)	6.64 (0.90, 12.7)
O₃	-0.15 (-1.28, 0.99)	-0.19 (-1.56, 1.2)
Co-pollutant model		
+NO₂	1.66 (-0.19, 3.51)	1.48 (-0.63, 3.63)
+SO₂	0.87 (-0.91, 2.65)	1.01 (-1.06, 3.12)
+O₃	1.77 (0.14, 3.41)	2.14 (0.25, 4.06)
+NO₂+SO₂	1.29 (-0.60, 3.18)	1.09 (-1.08, 3.30)
+NO₂+O₃	2.07 (0.07, 4.07)	1.86 (-0.40, 4.17)
+SO₂+O₃	1.17 (-0.85, 3.19)	1.30 (-1.02, 3.68)

179 * Generalized additive Poisson model for mortality count and Gaussian model for YLL was applied and
 180 controlled for long-term trend, seasonality, weather factors, calendar effect. Statistically significant effect
 181 estimates are in bold.

182 ER, excess risk; PM_{2.5}, particles with an aerodynamic diameter less than 2.5 µm.

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Table 5| Association between 10µg/m³ increase in PM_{2.5} (lag 0-1 day) and YLL and increase in deaths for AMI using the single pollutant model during 2011-15, according to sex and age*

	Increase in YLL(years)	ER of daily mortality count (%)
Gender		
Women	0.68 (-0.19, 1.55)	4.05 (1.00, 7.18)
Men	1.01 (-0.39, 2.41)	1.22 (-1.32, 3.82)
Age group		
Age ≤65 years	0.63 (-0.74, 2.01)	1.76 (-2.51, 6.21)
Age >65 years	0.93 (-0.01, 1.87)	2.39 (0.23, 4.61)

187 * Generalized additive Poisson model for mortality count and Gaussian model for YLL was applied and
188 controlled for long-term trend, seasonality, weather factors, calendar effect. Statistically significant effect
189 estimates are in bold.
190 Differences of the effect estimates between different sex and age group were not statistically significant.
191 ER, excess risk; PM_{2.5}, particles with an aerodynamic diameter less than 2.5 µm.

192
193

194 **4. Discussion**

195 This was the first study to examine the association of short-term PM_{2.5} exposure with mortality and YLL
196 from AMI. With the time-series data of 2011-2015 in Hong Kong, we found PM_{2.5} was associated with an
197 increase in AMI mortality and YLL due to AMI. The magnitude of PM_{2.5} effect on AMI mortality
198 appeared to be larger among women and older people with age > 65 years old.

199 The short-term association of PM_{2.5} with AMI mortality was consistent with previous epidemiological
200 studies (Mate et al., 2010; Sharovsky et al., 2004; Ueda et al., 2009; Zanobetti and Schwartz, 2009),
201 although the mechanisms of the PM_{2.5} effect remain not entirely clear. One of the major mechanisms may
202 be oxidative stress and inflammation (Brook et al., 2010). PM_{2.5} induces not only pulmonary oxidative
203 stress and inflammation but also a systemic inflammatory response. After inhalation of PM_{2.5}, a local
204 inflammatory response is developed and several proinflammatory mediators such as IL-6 and TNF-α are
205 increased as well, which induces the increment in the concentrations of blood fibrinogen and C-reactive
206 protein (CRP), which are important risk factors for AMI. Numerous studies have demonstrated that
207 exposure to particulate matters is associated with increase of fibrinogen and CRP, resulting in an
208 increased risk of AMI (Ghio et al., 2000; Pope et al., 2004; Ruckerl et al., 2007; Tornqvist et al., 2007).
209 With the inflammation after exposure to PM_{2.5}, the haemostatic system can be activated abnormally

210 (Seaton et al., 1995) and the blood viscosity can be increased (Peters et al., 1997), which promotes acute
211 thrombosis formation and atherosclerotic plaque.

212 PM_{2.5} can also disturb autonomic nervous system (ANS), which is another potential mechanism of its
213 association with AMI (Brook et al., 2004). Through decreased parasympathetic input to the heart, PM_{2.5} is
214 negatively associated with heart rate variability (HRV) (Devlin et al., 2003; Gong et al., 2004), which is a
215 significant and independent predictor of mortality after an AMI (Electrophysiology, 1996). Another
216 potential pathway is the direct translocation of PM_{2.5} into circulatory system and cause an acute
217 cardiovascular response. Reactive oxygen species (ROS) production and regulation of calcium levels are
218 two major pathways of the direct cardiovascular effect (Fiordelisi et al., 2017). Endothelial cells are
219 damaged especially by the specific metal components of PM_{2.5} (Niu et al., 2013). Recently, increasing
220 evidence (Bai et al., 2001; Okayama et al., 2006; Zuo et al., 2011) showed that ROS such as superoxide
221 and hydrogen peroxide play a role in various situations including pulmonary and systemic inflammatory
222 responses, vascular cytotoxicity and cardio myocyte dysfunction.

223 In the current study we estimated the YLL from AMI related to ambient PM_{2.5} pollution, which was probably
224 first such report. YLL, accounting for premature deaths and life expectancy, could provide more information on
225 the scale of the loss of life and would be a more informative indicator to assess the health burden compared with
226 mortality (Rabl, 2003). It would be an important complementary index for public policy making and health
227 service planning (Guo et al., 2013). But it was argued that mortality should be a typical indicator to demonstrate
228 the health outcome while YLL could be a supplementary indicator to reflect the disease burden because YLL
229 was less sensitive than mortality (Zhu et al., 2017). Furthermore, the normal distribution assumption for YLL
230 can be violated, especially for the YLL from some specific disease categories and the YLL in the subgroup
231 analyses, which may distort the true association between air pollution and YLL.

232 We could not directly compare the PM_{2.5} associated YLL estimates for AMI in the current study with those from
233 the previous studies in which all-cause mortality was the major health outcome (Guo et al., 2013; He et al.,
234 2016; Zhu et al., 2017). Meanwhile, the magnitude of the association between PM_{2.5} and YLL may also be
235 related to the population size, which makes the results from different settings uncomparable. He T et.al (He et
236 al., 2016) reported that a 10 $\mu\text{g}/\text{m}^3$ increase of PM_{2.5} was associated with 2.97 (95%CI -2.01 to 7.95) years
237 of YLL for all-cause mortality in Ningbo during 2009 to 2013. Yang J et.al (Yang et al., 2016) and Li G
238 et.al (Li et al., 2016) explored the short-term effect of air pollution on YLL from CVD and IHD
239 respectively, but PM_{2.5} was not included in their analyses.

240 For the metric of AMI mortality, females and older people were found to be more susceptible to PM_{2.5},
241 which was consistent with previous observations on AMI morbidity and mortality (D'Ippoliti et al., 2003;
242 Nuvolone et al., 2011). In 2010, the American Heart Association (AHA) statement concluded that women

243 may be at higher risk for cardiovascular mortality related to the particulate matter exposure (Brook et al.,
244 2010). Regional deposition of inhaled particles could be enhanced in women (Kim and Hu, 1998).
245 Women have fewer red blood cells (RBCs) compared with men, thus they would be more sensitive to the
246 toxicological effects of airborne pollution (Sorensen et al., 2003). On the other hand, the elderly is also a
247 high-risk group compared with the youth, which has been supported by AHA statement (Brook et al.,
248 2004). The immune system might be weaker among older people and other chronic diseases might occur,
249 which are the potential reasons for greater susceptibility among the elderly.

250 With co-pollutant adjustment in multi-pollutant models, the association of PM_{2.5} with mortality and YLL
251 from AMI decreased slightly, which may be due to the high correlation and the co-linearity between air
252 pollutants. We observed approximately linear concentration-response relationship of ambient PM_{2.5}
253 exposure with both AMI mortality and YLL, which was consistent with a previous study by Samoli E
254 (Samoli et al., 2005), in which the association between ambient particles and all-cause mortality in
255 Europe was linear without a threshold.

256 The current study was based in one single city of Hong Kong and thus the issue of generalizability should
257 be considered along with emerging evidence elsewhere. Outdoor monitoring data were used to
258 approximate the total population exposure to air pollution, which would cause measurement error of the
259 exposure level of PM_{2.5}, especially for the elderly and other subgroup populations who stay indoors longer.
260 Therefore, the exposure to PM_{2.5} might have been underestimated (Zeger et al., 2000) although Schwartz
261 J et al.(Schwartz et al., 2007) found that the ambient pollutant concentration of PM_{2.5} was reasonable
262 surrogates for personal exposures.

263 Despite these limitations, there were two strengths in the present study. First, as the first study applying
264 YLL as a complementary indicator to explore the short-term effect of PM_{2.5} on death from AMI, the
265 disease burden due to AMI could be assessed more precisely because the life expectancy at the age of
266 death was incorporated in the calculation of YLL. Second, unlike the previous studies in which the life
267 table applied to calculate YLL was not in accordance with the study period and the life expectancy was
268 only provided for age groups of 5 years interval (Guo et al., 2013; He et al., 2016; Yang et al., 2016), the
269 calculation of YLL in the current study was based on the corresponding life table of each year and life
270 expectancy of each exact age, which improved precision of the YLL estimates.

271 In conclusion, acute exposure to PM_{2.5} increased the mortality risk and YLL from AMI in Hong Kong.
272 Moreover, age and gender modified the effect of PM_{2.5} on AMI mortality. With the estimates of YLL, this
273 study adds to the evidence base for public health policy formulation and resource allocation.

274

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