

1 **Particulate matter from re-suspended mineral dust and emergency cause-**
2 **specific respiratory hospitalizations in Hong Kong**

3
4 Vivian C. PUN: Jockey Club School of Public Health and Primary Medicine, Chinese University
5 of Hong Kong, Hong Kong, Prince of Wales Hospital, Shatin, New Territory, Hong Kong;
6 Department of Health Sciences, Northeastern University, Boston Massachusetts, 02115 USA.

7 c.pun@neu.edu

8 Linwei TIAN: School of Public Health, University of Hong Kong, Hong Kong. linweit@hku.edu

9 Kin-fai HO: Jockey Club School of Public Health and Primary Medicine, Chinese University of
10 Hong Kong, Hong Kong, Prince of Wales Hospital, Shatin, New Territory, Hong Kong.

11 kfho@cuhk.edu.hk

12
13
14 *Corresponding author

15
16 Vivian C. Pun

17 Northeastern University

18 Department of Health Sciences

19 Boston, MA 02115

20 Phone: 617-373-7756

21 Email: c.pun@neu.edu

24 **ABSTRACT**

25 While contribution from non-exhaust particulate matter (PM) emissions towards traffic-related emissions is
26 increasing, few epidemiologic evidence of their health impact is available. We examined the association of short-
27 term exposure to PM₁₀ apportioned to re-suspended mineral dust with emergency hospitalizations for three major
28 respiratory causes in Hong Kong between 2000 and 2008. Time-series regression model was constructed to examine
29 association of PM₁₀ from re-suspended mineral dust with emergency hospitalizations for upper respiratory infection
30 (URI), chronic obstructive pulmonary disease (COPD) and asthma at exposure lag 0 to 5 days, adjusting for time
31 trends, seasonality, temperature and relative humidity. An interquartile range (6.8 µg/m³) increment in re-suspended
32 mineral dust on previous day was associated with 0.66% (95% CI: 0.12, 0.98) increase in total respiratory
33 hospitalizations, and 1.01% (95% CI: 0.14, 1.88) increase in URI hospitalizations. A significant 0.66%–0.80%
34 increases in risk of COPD hospitalizations were found after exposure to re-suspended mineral dust at lag 3 or later.
35 Exposure to mineral dust at lag 4 was linked to 1.71% increase (95% CI: 0.14, 2.22) in asthma hospitalizations.
36 Associations from single-pollutant models remained significant in multi-pollutant models, which additionally
37 adjusted for PM₁₀ contributing from vehicle exhaust, regional combustion, residual oil, re-suspended mineral dust,
38 fresh sea salt, aged sea salt, secondary nitrate and secondary sulfate, or gaseous pollutants (i.e., nitrogen dioxide,
39 sulfur dioxide, or ozone), respectively. Our findings provide insight into the biological mechanism by which non-
40 exhaust pollution may be associated with risk of adverse respiratory outcomes, and also stress the needs for
41 strategies to reduce emission and re-suspension of mineral dust. More research is warranted to assess the health
42 effects of different non-exhaust PM emissions under various roadway conditions and vehicle fleets.

43
44 **Keywords:** re-suspended mineral dust, respiratory hospitalization, upper respiratory, COPD, asthma, time-series
45 analysis

46 1. INTRODUCTION

47 Road traffic emissions are a major source of urban air pollution (Charron et al., 2007). They contribute to particulate
48 matter (PM) pollution via two ways: primary exhaust emissions from the tailpipe of motor vehicles, and non-exhaust
49 emissions from brake/tyre wear, road surface abrasion (road wear), or re-suspension of deposited material (e.g.,
50 mineral dust), which may be generated from a wide variety of sources (e.g., construction) near the roadside, due to
51 traffic-induced turbulence (Grigoratos and Martini, 2014; HEI, 2010; Pant and Harrison, 2013; Thorpe and Harrison,
52 2008). Non-exhaust emission of PM becomes increasingly important, particularly in developed countries in recent
53 decades. It is because contribution from primary exhaust emissions to total traffic PM pollution have reduced
54 considerably, due to the introduction of stringent standards for exhaust emission from diesel and gasoline vehicles,
55 technological upgrades of engine and exhaust systems, and cleaner fuels (HEI, 2010; Pant and Harrison, 2013;
56 Thorpe and Harrison, 2008). Thus, it is estimated that 80–90% of the total traffic PM emissions will come from non-
57 exhaust emission by the end of the decade (Rexeis and Hausberger, 2009).

58
59 As the contribution from non-exhaust PM emissions increases, so should our understanding of the potential toxicity
60 and health impact of non-exhaust-related particles. A few toxicological studies reported possible link between brake
61 wear particles and pro-inflammatory responses in lung cells *in vitro* (Gasser et al., 2009; Mazzarella et al., 2007),
62 and that studded tyre wear particles may have associated with considerable pro-inflammatory potential and profound
63 effects on macrophages *in vitro* and animal models (Gustafsson et al., 2008; Karlsson et al., 2011; Lindbom et al.,
64 2006). Yet, no epidemiologic studies have directly linked brake/tyre wear PM with adverse health effects on humans.
65 In contrast, several epidemiologic studies have investigated the potential association between fine particles (PM_{2.5})
66 from crustal soil or re-suspended mineral dust and human health using source apportionment modeling (Andersen et
67 al., 2007; Bell et al., 2014; Halonen et al., 2009; Lall et al., 2011; Sarnat et al., 2008; Schreuder et al., 2006). Though
68 these studies show suggestive association of re-suspended mineral dust with adverse respiratory outcomes,
69 heterogeneity in the findings remains. This may be explained by the application of PM_{2.5} rather than coarse particles
70 (PM_{2.5-10}) that are predominantly released by much of the non-exhaust processes (e.g., mechanical abrasion and
71 corrosion). Other possible reasons include variation in the chemical composition depending on factors including
72 brake and tire manufacturers, vehicle type and movement, street maintenance, season and meteorological parameters
73 (Pant and Harrison, 2013; Stanek et al., 2011), as well as difference in the underlying study populations.

74

75 In a recent study conducted by our research group, we found positive association between risks of emergency
76 hospitalizations for total respiratory causes and several PM₁₀ apportioned factors in Hong Kong (Pun et al., 2015).
77 Among the observations, the statistically significant association with PM₁₀ factor whose chemical profile was most
78 reflective of that of re-suspended mineral dust was intriguing, given that this was the first Asian study to report such
79 association, and the only other study that examined such relationship reported no significant association in
80 Copenhagen (Andersen et al., 2007). The current study built upon this observation by assessing the association of
81 PM₁₀ apportioned to re-suspended mineral dust with emergency hospitalization for three major respiratory causes
82 respectively. Parallel epidemiologic evaluation was also performed for aluminum, calcium and iron, the chemical
83 tracers that were largely apportioned to the re-suspended mineral dust factor, as well as evaluation for other PM₁₀
84 sources for comparison. Findings from this study may provide a useful perspective on and further understanding of
85 non-exhaust particles and their associated health impact, especially in Asia where studies of such kind is scarce.

86

87 **2. METHODS**

88 **2.1. Hospitalization Data**

89 We have previously compiled daily count data of emergency hospitalization into twenty-six publicly funded,
90 regulated by the Hong Kong Hospital Authority, hospitals across the territory for the period of 2001 to 2008 (Figure
91 S1; Pun et al., 2015, 2014a). In this study, we used the 9th revision of the International Classification of Diseases
92 (ICD-9) code to extract hospitalizations for three major respiratory causes: chronic obstructive pulmonary disease
93 (COPD; ICD-9: 491,492 and 496), asthma (ICD-9: 493). ICD-9 classification for acute upper respiratory infection
94 (URI, ICD-9: 464–466) was only available from 2001 to 2007; thus analysis for URI was restricted to 2001–2007,
95 whereas analyses for all other respiratory causes focused on 2001–2008. Since prior work has shown influenza
96 outbreaks to be a confounder of the association between PM and respiratory hospital admissions (Ren et al., 2006),
97 influenza hospitalizations (ICD-9: 487) were excluded from the total count of respiratory hospitalizations, and were
98 subsequently adjusted in the regression analysis.

99

100 **2.2. Exposure Assessment**

101 We employed the positive matrix factorization (PMF) 5.0 with the inclusion of nineteen PM₁₀ chemical species,
102 characterized from filter samples collected from six monitoring stations between 1 January 2001 and 31 December
103 2008, to identify physically realistic sources. Detailed analytical methods and profiles for source-apportioned PM₁₀
104 have been published elsewhere (Pun et al., 2014b, 2015; Z. B. Yuan et al., 2013), and a summary can be found in the
105 eAppendix (Supplementary). Briefly, eight apportioned factors, reflective of chemical profiles for vehicle exhaust,
106 regional combustion, residual oil, re-suspended mineral dust, fresh sea salt, aged sea salt, secondary nitrate and
107 secondary sulfate, were obtained (Figure S2). Some factors have distinctive chemical profile corresponding to a
108 specific emission source, such as vehicle exhaust emission that explained 80% of the variation in elemental carbon.
109 Two factors are identified as composite of two or more sources that cannot be reasonably separated in the PMF
110 model due to similarity in the source profile. Regional combustion emission was associated with large variation for a
111 mixture of combustible constituents, arising from wood/biomass burning (e.g., potassium ion) and coal combustion
112 (e.g., arsenic, cadmium, lead and zinc) in power plants and industrial facilities in the adjacent Pearl River Delta
113 region (Z. Yuan et al., 2013). In this study, we primarily focused on PM₁₀ from re-suspended mineral dust, a
114 composite characterized by high loadings of aluminum, calcium and iron. This factor refers to material that is
115 derived from a combination of exposed soil and dust from traffic, unpaved roads and construction activities from
116 within Hong Kong and nearby Pearl River Delta region (Yuan et al., 2006). It might also take into account, to some
117 degree, the particles emitted directly from brake/tyre/road wear and abrasion and then re-suspended, as evident by
118 the present of aluminum and iron that appear to be ubiquitous in brake linings, tyres and car paint. However, further
119 separation of brake/tire wear emission, road surface wear emission and re-suspension emissions was not attainable
120 in the PMF model.

121
122 Daily mean temperature and relative humidity were also obtained from the Hong Kong Observatory for the same
123 study period. Daily mean concentrations of nitrogen dioxide (NO₂) and sulfur dioxide (SO₂), as well as 8-hour mean
124 (10:00 AM–6:00 PM) concentration of ozone (O₃) were also calculated from hourly air pollutant data.

126 **2.3. Statistical analysis**

127 Given fairly high temporal correlations of PM₁₀ constituents and sources across the six monitoring stations ($r >$
128 0.70), we used *a priori* method of centering and averaging to remove the station-specific influence on the resulting

129 PMF-resolved PM₁₀ source contributions (Pun et al., 2014a, 2014b; Tian et al., 2013). The resultant time series
130 contained nonmissing territory-wide mean source-apportioned PM₁₀ for five consecutive days a week over the entire
131 study period. No notable difference was observed between days with nonmissing sampling data and those with
132 missing data in terms of cause-specific respiratory hospitalizations, apportioned sources and meteorological factors;
133 thus missing data were not imputed. Previous analyses have shown that risk estimates of PM₁₀ constituents and
134 sources were insensitive to either regression models in which no data imputation was used or models in which
135 imputation of missing sampling days was applied (Pun et al., 2014a, 2014b).

136
137 We used generalized additive model with log link, Poisson-distributed errors and autoregressive terms to estimate
138 the associations of PM₁₀ from re-suspended mineral dust and hospital admissions for total respiratory causes, COPD,
139 URI and asthma, respectively (Hastie and Tibshirani, 1990). Each model adjusted for time-varying confounders,
140 with smoothing splines with 8 degrees of freedom (df) per year for time trends and seasonality, 6 df for current day
141 temperature and previous 3-day moving average, and 3 df for current day relative humidity and previous 3-day
142 moving average selected *a priori* to minimize problems associated with multiple testing and core model selection
143 strategies (Bell et al., 2009). Dummy variables for day of week, public holidays and influenza epidemics were also
144 controlled for. We estimated hospitalization risks associated with levels of pollution on the same day (lag₀) and up to
145 5 days (lag₅) prior to hospitalization, while controlling for time trend, seasonality, meteorological conditions,
146 calendar effects and influenza epidemic. Parallel epidemiological evaluation for aluminum, calcium and iron, the
147 chemical tracers of re-suspended mineral dust, was also conducted. We further tested the robustness of the
148 association in multi-pollutant models, where we adjusted for all eight PM₁₀ sources simultaneously. Although not
149 the primary focus of the study, the associations of cause-specific respiratory hospitalizations with the remaining
150 PM₁₀ sources were also explored in the multi-pollutant models to provide comparison with that of re-suspended
151 mineral dust. In addition, we constructed two-pollutant models adjusting for NO₂, SO₂ or O₃ respectively. In an
152 exploratory analysis, the average PM_{2.5}-to-PM₁₀ ratios for calcium and iron were calculated using the available daily
153 constituent concentrations in PM_{2.5} and PM₁₀ measured in Tsuen Wan air monitoring station, which is located close
154 to the geographic center of Hong Kong and is likely to be more representative of Hong Kong's air quality in general,
155 between November 2004 and November 2005. Ratio for aluminum was not computed due to unstable PM_{2.5}
156 aluminum measurements. All estimates were reported as the percent increase [(relative risk-1)×100%] in daily

157 cause-specific emergency hospital admissions for an interquartile range (IQR) increment in pollutant concentrations.
158 The analyses were performed in the statistical environment R Software, version 3.1.2 (R Development Core Team,
159 Vienna).

160

161 3. RESULTS

162 Between 2001 and 2008, there were 659,963 emergency respiratory hospitalizations (~226 admissions per day) in
163 Hong Kong (Table 1). COPD hospitalizations accounted for 23.9% of total respiratory causes, followed by URI
164 (18.6%) and asthma (7.5%). The daily mean temperature and relative humidity were 23.5°C and 78.1% respectively.
165 The average daily PM₁₀ concentration was 55.9 µg/m³, of which 13% was from re-suspended mineral dust that was
166 almost comparable to the contribution from vehicle exhaust (15%; see Table S1 for descriptive statistics of all
167 apportioned PM₁₀ sources). The annual average contribution of mineral dust was relatively constant (p-trend =
168 0.053), compared to the rapid decreasing contribution of vehicle exhaust over the study period (p-trend = 0.015;
169 Figure S3 in the Supplementary). Refer to Table S1 in the Supplementary for Pearson's correlations between source-
170 apportioned PM₁₀ and meteorological factors.

171

172 PM₁₀ from re-suspended mineral dust and its chemical tracers was positively associated with hospitalizations for
173 total respiratory, COPD, URI and asthma at various lag exposures, respectively (Figure 1). An IQR of 6.8 µg/m³
174 increment in PM₁₀ contribution from mineral dust on previous day (lag 1) was statistically significantly associated
175 with 0.66% (95% CI: 0.12, 0.98) increase in total respiratory hospitalizations, and 1.01% (95% CI: 0.14, 1.88)
176 increase in URI hospitalizations. Re-suspended mineral dust was also positively linked to increased risk of COPD
177 hospitalizations at all lags examined, and the associations were statistically significant at 3 or more days prior to
178 hospitalization (lag 3), corresponding to 0.66%–0.80% increases in risk. Risks of asthma hospitalizations were
179 significantly increased (1.71%; 95% CI: 0.14, 2.22) with exposure to mineral dust at lag 4.

180

181 Results from multi-pollutant models that further adjusted for potential confounding of other PM₁₀ sources were
182 similar to those from single-pollutant models (Table 2). The increased hospitalization risks of total respiratory
183 causes (0.63%; 95% CI: 0.22, 1.04), URI (0.93%; 95% CI: 0.01, 1.85) and COPD (0.77%; 95% CI: 0.17, 1.36) were
184 slightly attenuated relative to single-pollutant model estimates; the risks of asthma hospitalizations was strengthened

185 at lag 4 based on the multipollutant model (1.44%; 95% CI: 0.34, 2.55) compared with the single-pollutant model. In
186 contrast to re-suspended mineral dust that exhibited differential temporal lag pattern of associations by cause-
187 specific respiratory causes, increment in PM₁₀ apportioned to vehicle exhaust, secondary nitrate, secondary sulfate at
188 earlier days prior to hospitalization (e.g., lag 3-5) were generally and positively associated with respiratory
189 hospitalizations regardless of causes, after adjusting for all other sources (Figure S4). Level of PM₁₀ from aged sea
190 salt was consistently linked to cause-specific respiratory hospitalizations at shorter lag (e.g., lag 0). However, no
191 discernible lag pattern was found for PM₁₀ apportioned to regional combustion, residual oil and fresh sea salt.
192 Additional sensitivity analysis showed that the patterns of associations were similar in two-pollutant models
193 adjusting for O₃, NO₂, or SO₂ (Figure S5 in the Supplementary). Figure 2 shows that tracer species of mineral dust
194 were most abundantly found in the coarser mode of PM₁₀, with average PM_{2.5}-to-PM₁₀ ratio ≤ 0.38 .

195

196 4. DISCUSSION

197 Particles arising from re-suspension of mineral dust are primarily coarse particles generated by mechanical grinding
198 of tyre and brakes, and agricultural activities, in contrast to their combustion counterparts that are mostly fine
199 particles (Peng et al., 2008). Currently, no study has examined the adverse health associated with PM_{2.5-10} from non-
200 exhaust traffic sources, because quantitative measurements of composition or sources of coarse PM are lacking.
201 Instead, most epidemiologic studies used PM_{2.5} composition to identify non-exhaust traffic PM. While studies from
202 Connecticut and Massachusetts USA, Finland and Chile have reported that PM_{2.5} from soil and/or road dust are
203 predictors of respiratory emergency department visits/hospitalizations (Bell et al., 2014; Cakmak et al., 2009a;
204 Halonen et al., 2009), research from New York City, Atlanta, and Washington State USA and Korea found no
205 evidence of association (Heo et al., 2014; Lall et al., 2011; Sarnat et al., 2008; Schreuder et al., 2006). Such
206 conflicting findings might attest to the limitation that PM_{2.5} does not capture majority of the particles emitted by
207 non-exhaust traffic sources, and thus may not capture their true health impacts.

208

209 This study is one of the two existing research that apportioned PM₁₀ composition to re-suspended mineral dust,
210 along with other sources, and is the first to examine the dust association with cause-specific respiratory
211 hospitalizations. The resultant exposure indicator, with an average PM_{2.5}-to-PM₁₀ ratio ≤ 0.38 , enabled us to capture
212 a wider spectrum of health effects associated with particles in either fine or coarse fraction of PM₁₀, compared to the

213 use of $PM_{2.5}$ composition alone. We associated an IQR increment in PM_{10} from re-suspended mineral dust on the
214 previous day with an increased risk of total respiratory hospitalization, after adjusting for covariates; this increase in
215 risk was mainly seen for URI hospitalizations. Statistically significantly positive associations were also observed for
216 COPD and asthma hospitalizations, though with pollutant exposure at 3 or more days prior to hospitalizations. The
217 patterns of associations were similar in models considering tracer species, multi-pollutant models adjusting for all
218 PM_{10} sources, and two-pollutant models adjusting for gaseous pollutants. Our observed positive association between
219 respiratory hospitalization and PM_{10} mineral dust is in contrast to findings reported by Andersen et al. (2007), which
220 showed no significant association of PM_{10} crustal materials (comprised of road concrete, tire wear, igneous rock and
221 limestone) with respiratory and asthma hospital admissions among elderly and children in Copenhagen. The
222 heterogeneous findings may due to the difference in PM_{10} composition and source profiles between cities, which
223 vary considerably depending on factors including brake and tire manufacturers, vehicle movement, street
224 maintenance, season and meteorological parameters. Other factors, such as longer study period, larger sample size in
225 our study, as well as population susceptibility between cities, may also explain the difference.

226
227 Precise biological mechanisms of how PM influence pathogenesis of respiratory illnesses remain unknown; however,
228 several key factors (e.g., size fraction, chemical composition, surface area) must be taken into account in order to
229 better understand the toxicity and potential adverse health effects of PM. In this study, we found PM_{10} mineral dust
230 or its tracer species, which were generally $PM_{2.5-10}$, to be associated with URI hospitalizations at shorter day
231 exposure lag, whereas association of hospitalizations for lower respiratory tract illnesses (i.e., COPD and asthma)
232 was at longer lag (e.g., lag 3). Such observed differential temporal lag patterns of associations may be explained by
233 the different triggering mechanisms in the lung by the fine and coarse fraction of PM_{10} soil/road dust (Centers for
234 Disease Control and Prevention et al., 2010). The differential temporal lag pattern of associations of respiratory
235 hospitalization by PM_{10} from vehicle exhaust and those from age sea salt may be also explained by particle size,
236 which has been shown to be important factor affecting particles deposition in the respiratory tract (Samet and
237 Dominici, 2000). Coarse particles, carrying more biological agents (e.g., endotoxins) than finer particles, deposit
238 primarily in the upper airway by impaction, thereby causing irritation and epithelial disruptions, inducing and/or
239 exacerbating upper respiratory illnesses with shorter delay (Ferguson et al., 2013; Schulz et al., 2000). On the other
240 hand, finer particles (e.g., $PM_{2.5}$) tend to travel deeper in airways and alveoli of the lung by impaction, sedimentation

241 or diffusion, causing inflammation in the lower respiratory tract and lung tissues in a more delayed observable
242 fashion (Donaldson and Stone, 2003; Kreyling et al., 2006). Animal models have shown coarse particles, on a mass
243 basis, to have higher hydroxyl radical generating capacity, greater cytokine production of macrophages and bacterial
244 endotoxin content, and produce significantly more pulmonary inflammatory responses compared to fine particles
245 (Shi et al., 2003; Tong et al., 2010); whereas smaller particles may be more potent if taking into consideration of
246 particle number or surface area. Our findings of PM₁₀ mineral dust provide support for the importance of PM size-
247 mediated impacts on upper and lower respiratory tract illnesses.

248
249 In addition, the heterogeneous associations across cause-specific respiratory illnesses or PM₁₀ sources reported in
250 this study may reflect effects of different generation processes, and of the combination of chemical constituents
251 emitted from these sources (Kim et al., 2012). Toxicological studies have suggested that carbonaceous constituents,
252 tracer for vehicle exhaust, may elevate levels of biomarkers for systemic inflammation and platelet activation
253 (Delfino et al., 2009), and exposure to Pb, present in coal combustion as well as construction material, enhanced the
254 production of oxidative stress, leading to inflammatory reactions (Saxena and Flora, 2004). Aluminum and iron,
255 tracers for re-suspended mineral dust, may contribute to the formation of free radicals, induce oxidative processes
256 and up-regulate pro-inflammatory mediators *in vitro*, thereby resulting in pulmonary inflammation and respiratory
257 diseases (Garçon et al., 2000; Risom et al., 2005). There is also suggestive evidence that patterns of temporal lag
258 structure of association varied by PM chemical constituents, as Kim et al. (2012) reported that delayed patterns of
259 increased relative risks for asthma hospitalizations began at later lags for sulfate and nitrate than for elemental and
260 organic carbon. Further work is needed to examine the biological mechanism in which PM constituents may affect
261 human health.

262
263 As the relative non-exhaust contribution towards any traffic related emissions is becoming greatly significant and
264 posing challenges for authority, findings from the current study stress the needs for strategies and measures for
265 abatement of particle emission and re-suspension. While road/street sweeper alone may not significantly reduce
266 PM₁₀ concentration (Amato et al., 2010; Keuken et al., 2010), other measures such as cleaning road and pavement
267 with water, or applying dust suppressants/dust binders (e.g., aqueous solutions of magnesium/calcium chloride and
268 calcium magnesium acetate) to maintain wet road surface have shown to be effective in certain conditions (Aldrin et

269 al., 2008; Norman and Johansson, 2006). Amato et al. (Amato et al., 2014) concluded in a recent review that a
270 combination of strategies aiming at minimizing the emission sources (e.g., improve wear properties of materials and
271 traffic, better maintenance of road) and minimizing re-suspension (e.g., road cleaning, reduce traffic) would likely
272 provide optimal abatement of PM non-exhaust emissions from road traffic.

273
274 Findings from our study should be interpreted with caution, as we cannot eliminate exposure measurement error.
275 Particles from re-suspended mineral dust tend to be more spatially heterogeneous, which might be subject to more
276 error and bias the estimates towards the null (Kioumourtzoglou et al., 2014b). Quantitative measurement of
277 composition or sources of coarse PM, or measurements of both PM_{2.5} and PM₁₀ composition for the calculation of
278 PM_{2.5-10} composition for source apportionment are needed to determine clear differences in source contribution by
279 size fraction and validate our findings. Finally, we did not account for the uncertainty of the estimated source
280 contributions directly in the health model. Any underestimation of the resulting inferences, however, is expected to
281 be small given the consistency in source identification across monitors and in effects obtained, and the agreement of
282 our findings with previous studies (Kioumourtzoglou et al., 2014a).

283 284 **5. CONCLUSION**

285 In summary, we found evidence of significantly positive association of total respiratory hospitalizations with
286 exposure PM₁₀ from re-suspended mineral dust, adjusting for covariates (e.g., vehicle exhaust). Exposure at early
287 lag day was linked to URI hospitalization, whereas exposure at later lags was associated with COPD and asthma
288 hospitalizations. These findings help prioritize research on the biologic mechanisms of PM effect on specific
289 pulmonary conditions, and stress the needs for strategies to reduce emission and re-suspension of mineral dust.

290 291 **6. Acknowledgement**

292
293 The authors thank the Hong Kong Environmental Protection Department for providing air pollution data, the Hong
294 Kong Observatory for providing meteorological data, and the Hospital Authority for providing hospital admission
295 data. This work was funded by the Health Medical Research Fund (Grant No. 11120311) and the Collaborative

296 Research Award of the Faculty of Medicine, the Chinese University of Hong Kong (Grant No. CRA14-15/07). The
297 authors declare that they have no competing interests.

AUTHORS' COPY

298 **7. REFERENCE**

- 299 Aldrin, M., Hobæk Haff, I., Rosland, P., 2008. The effect of salting with magnesium chloride on the concentration
300 of particular matter in a road tunnel. *Atmos. Environ.* 42, 1762–1776. doi:10.1016/j.atmosenv.2007.11.024
- 301 Amato, F., Cassee, F.R., Denier van der Gon, H.A.C., Gehrig, R., Gustafsson, M., Hafner, W., Harrison, R.M.,
302 Jozwicka, M., Kelly, F.J., Moreno, T., Prevot, A.S.H., Schaap, M., Sunyer, J., Querol, X., 2014. Urban air
303 quality: the challenge of traffic non-exhaust emissions. *J. Hazard. Mater.* 275, 31–6.
304 doi:10.1016/j.jhazmat.2014.04.053
- 305 Amato, F., Querol, X., Johansson, C., Nagl, C., Alastuey, A., 2010. A review on the effectiveness of street sweeping,
306 washing and dust suppressants as urban PM control methods. *Sci. Total Environ.* 408, 3070–84.
307 doi:10.1016/j.scitotenv.2010.04.025
- 308 Andersen, Z.J., Wahlin, P., Raaschou-nielsen, O.L.E., Scheike, T., Loft, S., 2007. Ambient particle source
309 apportionment and daily hospital admissions among children and elderly in Copenhagen. *J Expo Sci Env.*
310 *Epidemiol* 17, 625–636. doi:10.1038/sj.jes.7500546
- 311 Bell, M.L., Ebisu, K., Leaderer, B.P., Gent, J.F., Lee, H.J., Koutrakis, P., Wang, Y., Dominici, F., Peng, R.D., 2014.
312 Associations of PM_{2.5} Constituents and Sources with Hospital Admissions: Analysis of Four Counties in
313 Connecticut and Massachusetts (USA) for Persons ≥ 65 Years of Age. *Environ. Health Perspect.* 122, 138–44.
314 doi:10.1289/ehp.1306656
- 315 Bell, M.L., Ebisu, K., Peng, R.D., Samet, J.M., Dominici, F., 2009. Hospital admissions and chemical composition
316 of fine particle air pollution. *Am. J. Respir. Crit. Care Med.* 179, 1115–1120. doi:10.1164/rccm.200808-
317 1240OC
- 318 Cakmak, S., Dales, R.E., Vida, C.B., 2009a. Components of particulate air pollution and mortality in Chile. *Int J*
319 *Occup Env. Heal.* 15, 152–158.
- 320 Cakmak, S., Dales, R.E.R., Gultekin, T., Vidal, C.B., Farnendaz, M., Rubio, M.A., Oyola, P., Santiago, M.M. De,
321 Mario, C., Santiago, M.M. De, Cakmak, S., Dales, R.E.R., Gultekin, T., Vidal, C.B., Farnendaz, M., Rubio,
322 M.A., Oyola, P., 2009b. Components of Particulate Air Pollution and Emergency Department Visits in Chile.
323 *Arch Env. Occup Heal.* 64, 37–41.

324 Cao, J., Xu, H., Xu, Q., Chen, B., Kan, H., 2012. Fine particulate matter constituents and cardiopulmonary mortality
325 in a heavily polluted Chinese city. *Environ. Health Perspect.* 120, 373–8. doi:10.1289/ehp.1103671

326 Centers for Disease Control and Prevention, National Center for Chronic Disease Prevention and Health, Office on
327 Smoking and Health, 2010. Pulmonary Diseases, in: *How Tobacco Smoke Causes Disease: The Biology and*
328 *Behavioral Basis for Smoking-Attributable Disease: A Report of the Surgeon General.* Centers for Disease
329 Control and Prevention (US).

330 Charron, A., Harrison, R.M., Quincey, P., 2007. What are the sources and conditions responsible for exceedences of
331 the 24h PM₁₀ limit value (50µg^m-³) at a heavily trafficked London site? *Atmos. Environ.* 41, 1960–1975.
332 doi:10.1016/j.atmosenv.2006.10.041

333 Delfino, R.J., Staimer, N., Tjoa, T., Gillen, D.L., Polidori, A., Arhami, M., Kleinman, M.T., Vaziri, N.D., Longhurst,
334 J., Sioutas, C., 2009. Air pollution exposures and circulating biomarkers of effect in a susceptible population:
335 clues to potential causal component mixtures and mechanisms. *Environ. Health Perspect.* 117, 1232–8.
336 doi:10.1289/ehp.0800194

337 Donaldson, K., Stone, V., 2003. Current hypotheses on the mechanisms of toxicity of ultrafine particles. *Ann. Ist.*
338 *Super. Sanita* 39, 405–10.

339 Ferguson, M.D., Migliaccio, C., Ward, T., 2013. Comparison of how ambient PM_c and PM_{2.5} influence the
340 inflammatory potential. *Inhal. Toxicol.* 25, 766–773. doi:10.3109/08958378.2013.847993. Comparison

341 Garçon, G., Shirali, P., Garry, S., Fontaine, M., Zerimech, F., Martin, a, Hanothiaux, M.H., 2000. Polycyclic
342 aromatic hydrocarbon coated onto Fe₂O₃ particles: assessment of cellular membrane damage and
343 antioxidant system disruption in human epithelial lung cells (L132) in culture. *Toxicol. Lett.* 117, 25–35.

344 Gasser, M., Riediker, M., Mueller, L., Perrenoud, A., Blank, F., Gehr, P., Rothen-Rutishauser, B., 2009. Toxic
345 effects of brake wear particles on epithelial lung cells in vitro. *Part. Fibre Toxicol.* 6, 30. doi:10.1186/1743-
346 8977-6-30

347 Grigoratos, T., Martini, G., 2014. Non-exhaust traffic related emissions. Brake and tyre wear PM.
348 doi:10.2790/21481

349 Gustafsson, M., Blomqvist, G., Gudmundsson, A., Dahl, A., Swietlicki, E., Bohgard, M., Lindbom, J., Ljungman, A.,

350 2008. Properties and toxicological effects of particles from the interaction between tyres, road pavement and
351 winter traction material. *Sci. Total Environ.* 393, 226–40. doi:10.1016/j.scitotenv.2007.12.030

352 Halonen, J.I., Lanki, T., Yli-Tuomi, T., Tiittanen, P., Kulmala, M., Pekkanen, J., 2009. Particulate air pollution and
353 acute cardiorespiratory hospital admissions and mortality among the elderly. *Epidemiology* 20, 143–153.
354 doi:10.1097/EDE.0b013e31818c7237

355 Hastie, T.J., Tibshirani, R.J., 1990. *Generalized Additive Models*. Chapman & Hall/CRC, Boca Raton, Florida.

356 HEI, 2010. *Traffic-Related Air Pollution: A Critical Review of the Literature on Emissions, Exposure, and Health*
357 *Effects*. SPECIAL REPORT 17. Heal. Eff. Inst. 386.

358 Heo, J., Schauer, J.J., Yi, O., Paek, D., Kim, H., Yi, S.-M., 2014. Fine particle air pollution and mortality:
359 importance of specific sources and chemical species. *Epidemiology* 25, 379–88.
360 doi:10.1097/EDE.0000000000000044

361 Karlsson, H., Lindbom, J., Ghafouri, B., Lindahl, M., Tagesson, C., Gustafsson, M., Ljungman, A.G., 2011. Wear
362 Particles from Studded Tires and Granite Pavement Induce Pro-inflammatory Alterations in Human
363 Monocyte-Derived Macrophages: A Proteomic Study. *Chem. Res. Toxicol.* 24, 45–53. doi:10.1021/tx100281f

364 Keuken, M., Denier van der Gon, H., van der Valk, K., 2010. Non-exhaust emissions of PM and the efficiency of
365 emission reduction by road sweeping and washing in the Netherlands. *Sci. Total Environ.* 408, 4591–9.
366 doi:10.1016/j.scitotenv.2010.06.052

367 Kim, S.-Y., Peel, J.L., Hannigan, M.P., Dutton, S.J., Sheppard, L., Clark, M.L., Vedal, S., 2012. The temporal lag
368 structure of short-term associations of fine particulate matter chemical constituents and cardiovascular and
369 respiratory hospitalizations. *Environ. Health Perspect.* 120, 1094–9. doi:10.1289/ehp.1104721

370 Kioumourtoglou, M.-A., Coull, B. a, Dominici, F., Koutrakis, P., Schwartz, J., Suh, H., 2014a. The impact of
371 source contribution uncertainty on the effects of source-specific PM_{2.5} on hospital admissions: A case study
372 in Boston, MA. *J. Expo. Sci. Environ. Epidemiol.* 1–7. doi:10.1038/jes.2014.7

373 Kioumourtoglou, M.-A., Spiegelman, D., Szpiro, A. a, Sheppard, L., Kaufman, J.D., Yanosky, J.D., Williams, R.,
374 Laden, F., Hong, B., Suh, H., 2014b. Exposure measurement error in PM_{2.5} health effects studies: a pooled
375 analysis of eight personal exposure validation studies. *Environ. Health* 13, 2. doi:10.1186/1476-069X-13-2

376 Kreyling, W.G., Semmler-Behnke, M., Möller, W., 2006. Ultrafine Particle-Lung Interactions: Does Size Matter? J.
377 Aerosol Med. 19, 74–83.

378 Lall, R., Ito, K., Thurston, G.D., 2011. Distributed lag analyses of daily hospital admissions and source-apportioned
379 fine particle air pollution. *Env. Heal. Perspect* 119, 455–460. doi:10.1289/ehp.1002638

380 Lindbom, J., Gustafsson, M., Blomqvist, G., Dahl, A., Gudmundsson, A., Swietlicki, E., Ljungman, A.G., 2006.
381 Exposure to Wear Particles Generated from Studded Tires and Pavement Induces Inflammatory Cytokine
382 Release from Human Macrophages. *Chem. Res. Toxicol.* 19, 521–530. doi:10.1021/tx0503101

383 Mazzarella, G., Ferraraccio, F., Prati, M.V., Annunziata, S., Bianco, A., Mezzogiorno, A., Liguori, G., Angelillo,
384 I.F., Cazzola, M., 2007. Effects of diesel exhaust particles on human lung epithelial cells: An in vitro study.
385 *Respir. Med.* 101, 1155–1162. doi:10.1016/j.rmed.2006.11.011

386 Norman, M., Johansson, C., 2006. Studies of some measures to reduce road dust emissions from paved roads in
387 Scandinavia. *Atmos. Environ.* 40, 6154–6164. doi:10.1016/j.atmosenv.2006.05.022

388 Pant, P., Harrison, R.M., 2013. Estimation of the contribution of road traffic emissions to particulate matter
389 concentrations from field measurements: A review. *Atmos. Environ.* 77, 78–97.
390 doi:10.1016/j.atmosenv.2013.04.028

391 Peng, R.D., Chang, H.H., Bell, M.L., Mcdermott, A., Zeger, S.L., Samet, J.M., Dominici, F., 2008. Coarse
392 particulate matter air pollution and hospital admissions for cardiovascular and respiratory diseases among
393 medicare patients. *J. Am. Med. Assoc.* 299, 2172–2179. doi:10.1001/jama.299.18.2172

394 Pun, V.C., Tian, L., Yu, I.T.S., Kioumourtzoglou, M.-A., Qiu, H., 2015. Differential Distributed Lag Patterns of
395 Source-Specific Particulate Matter on Respiratory Emergency Hospitalizations. *Environ. Sci. Technol.*
396 150204132337001. doi:10.1021/es505030u

397 Pun, V.C., Yu, I.T.-S.S., Qiu, H., Ho, K.-F.F., Sun, Z., Louie, P.K.K., Wong, T.W., Tian, L., 2014a. Short-term
398 associations of cause-specific emergency hospitalizations and particulate matter chemical components in
399 Hong Kong. *Am. J. Epidemiol.* 179, 1086–95. doi:10.1093/aje/kwu026

400 Pun, V.C., Yu, I.T., Ho, K., Qiu, H., Sun, Z., Tian, L., 2014b. Differential Effects of Source-Specific Particulate
401 Matter on Emergency Hospitalizations for Ischemic Heart Disease in Hong Kong. *Env. Heal. Perspect.* (in

402 press), 391–396.

403 Ren, C., Williams, G.M., Tong, S., 2006. Does particulate matter modify the association between temperature and
404 cardiorespiratory diseases? *Environ. Health Perspect.* 114, 1690–6.

405 Rexeis, M., Hausberger, S., 2009. Trend of vehicle emission levels until 2020 – Prognosis based on current vehicle
406 measurements and future emission legislation. *Atmos. Environ.* 43, 4689–4698.
407 doi:10.1016/j.atmosenv.2008.09.034

408 Risom, L., Moller, P., Loft, S., 2005. Oxidative stress-induced DNA damage by particulate air pollution. *Mutat Res*
409 592, 119–137.

410 Samet, J., Dominici, F., 2000. Fine particulate air pollution and mortality in 20 US cities, 1987–1994. *New Engl.*
411 *J. ...* 343, 1742–1749.

412 Sarnat, J.A., Marmur, A., Klein, M., Kim, E., Russell, A.G., Sarnat, S.E., Mulholland, J.A., Hopke, P.K., Tolbert,
413 P.E., 2008. Fine particle sources and cardiorespiratory morbidity: an application of chemical mass balance and
414 factor analytical source-apportionment methods. *Env. Heal. Perspect* 116, 459–466.

415 Saxena, G., Flora, S.J.S., 2004. Lead-induced oxidative stress and hematological alterations and their response to
416 combined administration of calcium disodium EDTA with a thiol chelator in rats. *J. Biochem. Mol. Toxicol.*
417 18, 221–33. doi:10.1002/jbt.20027

418 Schreuder, A.B., Larson, T. V., Sheppard, L., Claiborn, C.S., 2006. Ambient woodsmoke and associated respiratory
419 emergency department visits in Spokane, Washington. *Int J Occup Env. Heal.* 12, 147–153.

420 Schulz, H., Brand, P., Heyder, J., 2000. *Particle Deposition in the Respiratory Tract.* Marcel Decker, New York, NY.

421 Shi, T., Knaapen, a M., Begerow, J., Birmili, W., Borm, P.J. a, Schins, R.P.F., 2003. Temporal variation of
422 hydroxyl radical generation and 8-hydroxy-2'-deoxyguanosine formation by coarse and fine particulate matter.
423 *Occup. Environ. Med.* 60, 315–21.

424 Stanek, L.W., Sacks, J.D., Dutton, S.J., Dubois, J.-J.B., 2011. Attributing health effects to apportioned components
425 and sources of particulate matter: An evaluation of collective results. *Atmos. Environ.* 45, 5655–5663.
426 doi:10.1016/j.atmosenv.2011.07.023

427 Thorpe, A., Harrison, R.M., 2008. Sources and properties of non-exhaust particulate matter from road traffic: a

428 review. *Sci. Total Environ.* 400, 270–82. doi:10.1016/j.scitotenv.2008.06.007

429 Tian, L., Ho, K., Louie, P.K.K., Qiu, H., Pun, V.C., Kan, H., Yu, I.T.S., Wong, T.W., 2013. Shipping Emissions
430 Associated with Increased Cardiovascular Hospitalizations. *Atmos. Environ.* 74, 320–325.
431 doi:10.1016/j.atmosenv.2013.04.014

432 Tong, H., Cheng, W.-Y., Samet, J.M., Gilmour, M.I., Devlin, R.B., 2010. Differential cardiopulmonary effects of
433 size-fractionated ambient particulate matter in mice. *Cardiovasc. Toxicol.* 10, 259–67. doi:10.1007/s12012-
434 010-9082-y

435 Yuan, Z., Yadav, V., Turner, J.R., Louie, P.K.K., Lau, A.K.H., 2013. Long-term trends of ambient particulate matter
436 emission source contributions and the accountability of control strategies in Hong Kong over 1998–2008.
437 *Atmos. Environ.* 76, 21–31. doi:10.1016/j.atmosenv.2012.09.026

438 Yuan, Z.B., Lau, A.K.H., Zhang, H.Y., Yu, J.Z., Louie, P.K.K., Fung, J.C.H., 2006. Identification and
439 spatiotemporal variations of dominant PM10 sources over Hong Kong. *Atmos. Environ.* 40, 1803–1815.
440 doi:10.1016/j.atmosenv.2005.11.030

441 Yuan, Z.B., Yadav, V., Turner, J.R., Louie, P.K.K., Lau, A.K.H., 2013. Long-term trends of ambient particulate
442 matter emission source contributions and the accountability of control strategies in Hong Kong over 1998–
443 2008. *Atmos. Environ.* 76, 21–31. doi:10.1016/j.atmosenv.2012.09.026

444

445